

1 **Evaluating Variations in Great Salt Lake Inflow to Infer Human Consumptive Water**
2 **Use, A Volume Reconstruction Approach**

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6 **Key Points:**

- 7 • Current estimates of consumptive water use that apply direct calculation methods for the
8 Great Salt Lake Basin are disparate
- 9 • Flow reductions into the lake due to consumptive use can be estimated using volume
10 reconstruction and lake level measurements
- 11 • Our volume reconstruction consumptive use estimate of 2.3 km³/yr independently
12 corroborates direct calculations

13 **Abstract**

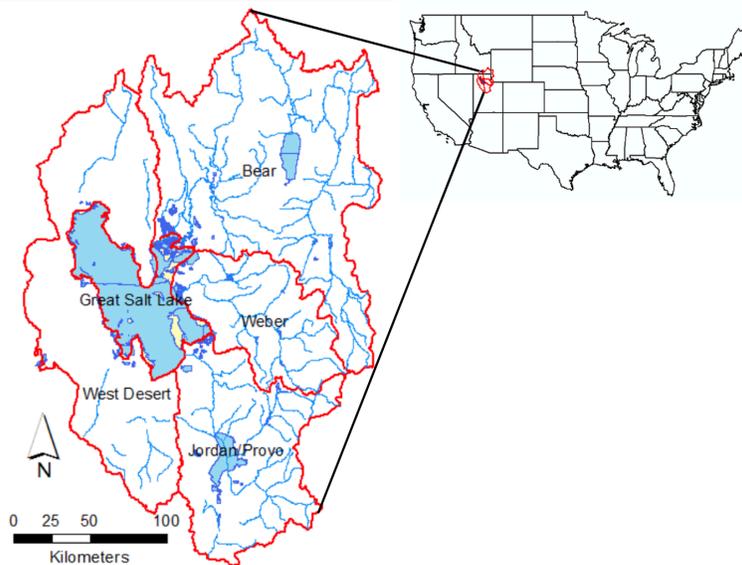
14 The declining water level in Great Salt Lake (GSL) has been attributed to human
15 consumptive water use that depletes natural streamflow into the lake. Understanding depletions
16 due to historical consumptive water use within the GSL Basin is important to managing present
17 and future lake conditions. Direct calculations of consumptive water use in the basin are made by
18 summing detailed uses and return flows. However, this method is limited by insufficient data and
19 resulting estimates thus far have been disparate. In this study, we reconstructed total GSL water
20 inputs and stream inflows using lake levels recorded from 1847-2023 to estimate the magnitude
21 of reductions due to consumptive use and the associated lake level decline. To do so, we
22 developed a method that uses lake volume changes derived from bathymetry and water surface
23 elevation measurements along with estimates of annual evaporation and precipitation over the
24 lake to hindcast inflow volume to the lake. The declining trend in lake inflow, without associated
25 precipitation or natural streamflow trends, was used to quantify basin wide water depletions to be
26 up to 2.3 km³/yr and the current lake level decline associated with this estimate to be as much as
27 4.6 meters. This basin wide depletion estimate depends only on lake level, precipitation, and
28 evaporation estimates and is not limited by the challenges of aggregating individual diversions
29 and return flows.

30 **Plain Language Summary**

31 The Great Salt Lake (GSL) is an ecologically and economically important resource that is
32 at risk due to declining lake levels. This decline has been attributed to human consumptive water
33 use that depletes streamflow into the lake. Understanding depletions due to consumptive water
34 use within the GSL Basin is important to managing present and future lake conditions. We
35 reconstructed total GSL water inputs using lake levels recorded from 1847-2023. Calculations of
36 lake volume changes associated with measured lake level changes, together with information on
37 direct precipitation to and evaporation from the lake, support the estimation of lake inflow back
38 to the date lake levels were first recorded and prior to the dates for which streamflow has been
39 measured. Trends associated with these prior streamflow hindcasts were used to infer overall
40 consumptive use depletion of streamflow and the impact this has had on lake levels. This
41 approach complements and extends direct approaches for calculating consumptive use depletion
42 of streamflow in the GSL Basin. The results are important because they quantify how
43 consumptive use has impacted the lake and serve as a reference for efforts towards conservation
44 and other water management actions aimed at restoring the lake to more healthy levels.

45 **1 Introduction**

46 The Great Salt Lake (GSL) is a terminal lake located in northern Utah within the Great
47 Basin (Figure 1). It is the largest terminal lake in North America. The lake is divided into two
48 arms by a railroad causeway that was built in 1959. The south arm, Gilbert Bay, receives



49

50 **Figure 1.** Location of the Great Salt Lake and its contributing basins.

51 approximately 95% of the surface inflow to the lake entering along the southeastern shores from
 52 three major rivers, Bear, Weber, and Jordan (Loving et al., 2000). In the north arm, Gunnison
 53 Bay, input is limited to direct precipitation and intermittent runoff. Gilbert Bay is roughly twice
 54 the area and volume of Gunnison Bay. Evaporation is the only outflow from either arm, resulting
 55 in the lake's high salinity (e.g., Loving et al., 2000). Level, area, and volume of GSL adjust to
 56 balance differences in these inflows and outflows and, with its large size and relatively shallow
 57 depth (13 meters at its deepest), area and volume vary greatly with fluctuating levels.

58 In recent years, the GSL water surface elevation (WSE) has declined to critically low
 59 levels, with much of the decline being attributed to depletion of streamflow through human
 60 consumptive water use (Wurtsbaugh et al., 2017). This is superimposed on natural climate driven
 61 low frequency cycles and periodic extremes (Lall and Mann, 1995). There is thus a need to
 62 quantify historical reductions in inflow to the GSL due to the depletion of streamflow and
 63 associated declines of GSL WSE. However, while GSL WSE has been measured since 1847,
 64 streamflow measurements from all the main lake inflows are complete only following 1949.
 65 Much development of agricultural water use within the GSL Basin started immediately upon
 66 settlement of the GSL valley by European (Mormon) settlers in 1847. The impacts of the
 67 resultant reductions in streamflow are therefore already part of streamflow measurements
 68 making it difficult to determine reductions in streamflow due to water use depletions from
 69 current trends in streamflow measurements alone.

70 Mohammed and Tarboton (2012) defined a GSL volume sensitivity measure, denoted as
 71 ϕ , as the ratio of the standard deviation of input variables to the standard deviation of the lake
 72 volume change at an annual time scale. They found that lake volume is eight times more
 73 sensitive to streamflow ($\phi = 0.83$) than evaporation ($\phi = 0.10$) and nearly three times more
 74 sensitive than precipitation ($\phi = 0.30$). The strong relationship between lake volume changes and
 75 streamflow suggests that volume changes derived from changes in WSE may be useful to infer
 76 inflow to GSL back to the time when levels were first measured and thus extract information not

77 available in post 1949 streamflow measurements on how streamflow may have been reduced by
78 water use development.

79 Tree-ring reconstructions of streamflow extend several centuries into the past and provide
80 information on streamflow prior to the availability of measurements. Based on climate and soil
81 moisture driven growth of trees, tree ring reconstructions are not impacted by diversions and thus
82 represent estimates of natural flow. Streamflow reconstructions from tree-ring estimates dating
83 back approximately 1200 years for the Bear River (DeRose et al., 2015), 600 years for the Weber
84 River (Bekker et al., 2014), and 800 years for the Jordan River (Tikalsky, 2007) show no
85 significant long-term trends. Therefore, any trend in streamflow into the lake inferred from lake
86 volume changes over the period of human development prior to the availability of complete lake
87 inflow streamflow measurements can be attributed to human influence reflecting streamflow
88 depletion due to consumptive use (CU).

89 The goal of this paper was to investigate and reconstruct streamflow into GSL from
90 volume change measurements and, from trends in these reconstructions, infer how human
91 consumptive water use has impacted streamflow. Our approach is an alternative to, and check or
92 validation of, uncertain and unknown early direct consumptive use depletion estimates. We
93 developed a volume reconstruction approach to reconstruct the historical record of annual inflow
94 to GSL for U.S. water years 1848-2023 by using lake volume changes along with evaporation
95 and precipitation hindcasts as inputs to a water mass balance for the lake. Our hindcast and mass
96 balance process, along with supporting assumptions, are described in the methods section. The
97 objective was to establish the magnitude of streamflow reductions due to CU based on mass
98 balance evidence in the historical record for GSL WSE as an alternative to and check on direct
99 CU streamflow depletions that are only available during recent periods and are uncertain. Trends
100 from the reconstructed inflow were used to estimate CU in the GSL Basin and how, over the
101 period that WSE has been measured, CU as led to changes in WSE.

102 **2 Background**

103 Both arms of GSL have a salt concentration several times higher than the ocean. The
104 hypersaline lake and associated marshes are vital for breeding and migrating shorebirds (IWJV,
105 2016). The Western Hemisphere Shorebird Reserve Network has designated the GSL as a site of
106 Hemispheric Importance, the highest rank of importance (WHSRN, 2016). A major food source
107 for the birds is the lake's brine shrimp, which are also considered a keystone species because
108 they control the lake's phytoplankton (Stephens, 1990; Aldrich and Paul, 2002). Brine shrimp
109 harvesting is also important economically. GSL is the largest source of brine shrimp for the
110 global aquaculture industry, feeding an estimated 10 million metric tons of farmed seafood each
111 year (Penrod, 2023; Brown et al., 2023). The north arm of GSL can no longer support brine
112 shrimp or brine fly populations due to extremely high salinities (Barnes & Wurtsbaugh, 2015).
113 Although the less saline south arm does not support fish, brine shrimp and brine flies flourish in
114 most years (WHSRN, 2016). However, salinity nearing or reaching biotic thresholds can be
115 harmful to the brine shrimp and brine flies. Disruptions in the brine shrimp and brine fly
116 population propagate and result in disturbances further up the food chain, affecting migratory
117 bird populations and food production for human consumption.

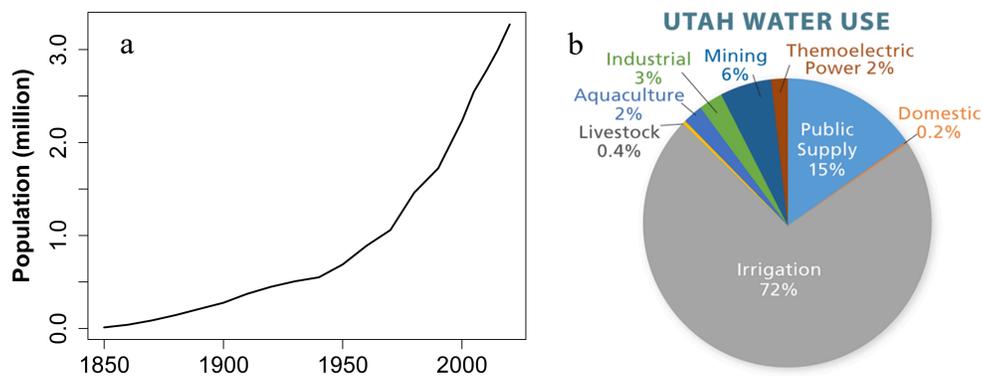
118 The south arm is also home to the estuaries of Bear River Bay and Farmington Bay. Both
119 bays receive freshwater inflows and, therefore, have lower salinity than the rest of the lake.
120 When GSL WSE is at or above what is regarded as normal elevations (average ~4200 feet or

121 1280 meters), these bays provide shorebird habitat and are commonly used for recreation.
 122 However, being shallower areas of the lake, both bays are nearly dry at current low lake levels
 123 and the possibility of complete desiccation is a real threat. In addition, the ecosystems in these
 124 bays are fragile and struggle under low WSE conditions (Wurtsbaugh et al., 2016).

125 Low WSE and high salinities throughout the lake also make it more difficult to pump
 126 lake water to the evaporation ponds for mineral extraction, which are located on the periphery of
 127 the lake. In 2014, Morton Salt dug a five-mile-long canal in the south arm to deliver brine from
 128 the distant lake water to their salt ponds and processing plants (Wurtsbaugh et al., 2016). US
 129 Magnesium (MagCorp), a producer of magnesium located on the south arm, is also extending
 130 their canals by several miles (UDEQ, 2022). In the north arm, accumulating salt precipitate as
 131 thick as 8 feet requires repeated dredging from the Behrens Trench, an underwater canal used in
 132 the mineral extraction process by Compass Minerals (Standard Examiner, 2016; FGSL, 2018).

133 Another problem created by low WSE is increased lakebed exposure, which increases the
 134 potential for dust pollution and related human health impacts due to dust from the exposed
 135 shores (Hahnenberger and Nicoll, 2012; 2014). This is worrisome due to the proximity to Salt
 136 Lake City and its growing metropolitan area. Recent work has also attributed significant dust
 137 from exposed GSL lakebed area with earlier snowmelt due to dust on snow albedo reductions
 138 (Lang et al., 2023). All of the concerns mentioned here that threaten the lake's uses are
 139 exacerbated by low WSE. Therefore, water use in the GSL Basin and its effect on the lake are
 140 important to understand and manage.

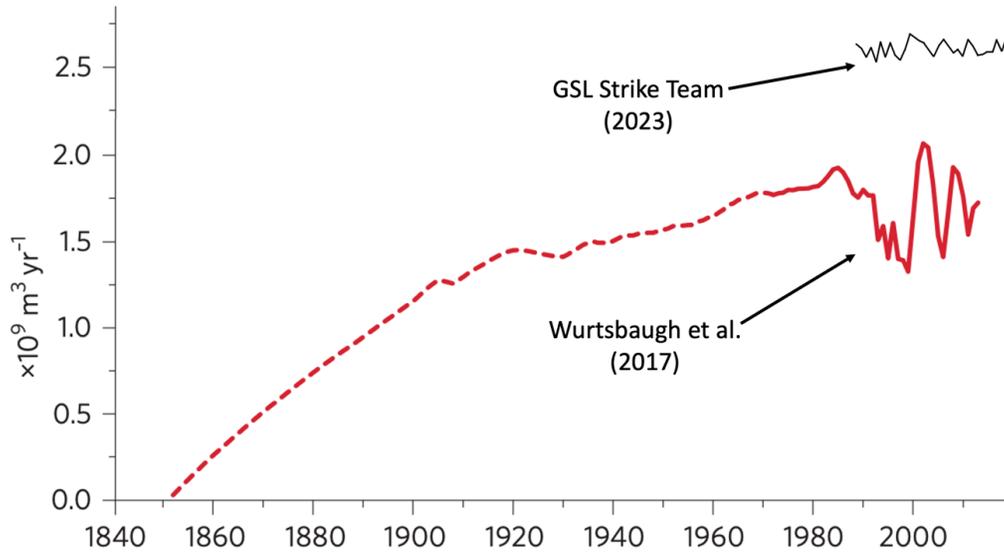
141 Following the settlement of Salt Lake City in 1847 by Mormon pioneers, the population
 142 of Utah has grown steadily (Figure 2a). Much of this growth is centered around Salt Lake City
 143 and is within the basins draining into GSL. With increased population comes increased demands
 144 on the land and its resources, including water. Agricultural irrigation is the largest water
 145 consumer in Utah, accounting for over 70% of the state's water use (UGS, 2021), with other uses
 146 making up the balance (Figure 2b). Increases in water use within the GSL Basin impact WSE,
 147 volume, and salinity and, therefore, the many uses of the lake.



148

149 **Figure 2.** a) Population of Utah, 1850-2020. b) Water use in Utah (UGS, 2021).

150 In recent years, there have been efforts to estimate CU for the GSL Basin. However,
 151 existing estimates of CU in the literature are disparate. For example, Wurtsbaugh et al. (2017)
 152 estimate present day total annual CU to be $\sim 1.7 \text{ km}^3$ while the GSL Strike Team (GSLST, 2023)
 153 estimate is $\sim 2.6 \text{ km}^3$ (Figure 3), more than a 50% difference. The GSL Strike Team reported
 154 individual year CU for the contemporary period, 1989-2020, and found CU within that period to



155

156 **Figure 3.** Contrasting estimates of consumptive use for the GSL Basin, including historical
 157 consumptive water use from Wurtsbaugh et al (2017, red line), and contemporary consumptive
 158 use reported by GSL Strike Team (2023, black line). Consumptive uses prior to 1989 (dashed
 159 line) are estimated and based on unpublished data from Utah Division of Water Resources
 160 (Wurtsbaugh et al., 2017).

161 be fairly consistent with no overall trend (black line, Figure 3). Wurtsbaugh et al. infers CU back
 162 to ~1850 and their estimate (red line, Figure 3) suggests that the majority of CU growth occurred
 163 shortly after the GSL Basin was settled, prior to the turn of the century and prior to complete
 164 stream gaging (red dashed line, Figure 3). Wurtsbaugh et al. (2017) also estimated that the total
 165 CU in the GSL Basin between ~1850 and present day has decreased the WSE of GSL by over 3
 166 meters. These direct approaches to quantifying CU are challenging as they involve the
 167 aggregation of uncertain inputs, including summaries of land use and estimates of
 168 evapotranspiration, ungauged inflows, and irrigation return flows (Wurtsbaugh et al., 2017). This
 169 disparity is a motivating factor for this study to estimate and corroborate water use independently
 170 through a whole basin water balance approach based on the recorded history of lake levels.

171 3 Methods

172 The water mass balance equation for a closed basin lake such as GSL may be stated as:

$$173 \text{Change in Volume} = \text{Inflow} + \text{Precipitation} - \text{Evaporation} \quad (1)$$

Or equivalently, given our interest in estimating streamflow prior to its measurement, as:

$$174 \text{Inflow} = \text{Change in Volume} - \text{Precipitation} + \text{Evaporation} \quad (2)$$

175 This offers opportunities for estimating hindcast Inflow through measurements of the change in
 176 volume and estimates of lake precipitation and evaporation. Change in Volume can be calculated
 177 from WSE data, which are available starting in 1847 through the USGS National Water
 178 Information System website (USGS Site Numbers 10010000 and 10010100, in
 179 <https://maps.waterdata.usgs.gov/mapper>, accessed on 1 November 2023). Precipitation data over
 180 the GSL Basin are available to a varying degree of quality from sources such as PRISM since
 approximately 1900 (Daly et al., 2008, <https://prism.nacse.org>). Evaporation has not been

181 directly measured and is generally calculated from mass balance closure, which limits its
 182 availability to 1949 to present due to the availability of streamflow data. Therefore, to extend
 183 estimates of Inflow back over times prior to its measurement, suitable assumptions or
 184 approximations must be made for Precipitation and Evaporation.

185 Here we evaluated a total of four approaches to hindcasting annual inflow to GSL for
 186 U.S. water years 1848-2023. Note that in the U.S. a water year is defined to run from October 1
 187 of the previous year to September 30 of the designated year. Year in this paper always refers to
 188 water year unless noted otherwise. For all four approaches, we assumed that the depth of
 189 freshwater equivalent lake evaporation each year was constant at its average value. The first
 190 three approaches use different assumptions for approximating precipitation. First, because
 191 precipitation data is not available for the full duration of our study, we assumed precipitation
 192 depth was constant at its long-term average value. Second, recognizing that precipitation and
 193 streamflow are correlated, we assumed a constant runoff ratio and constant proportion between
 194 basin and lake precipitation. Third, we hindcast streamflow using PRISM precipitation data for
 195 the period they are available. For the fourth hindcast approach, we included precipitation volume
 196 with the inflow quantity being hindcast, effectively estimating streamflow plus precipitation
 197 from volume changes as the hindcast quantity. This approach avoids introducing uncertainty due
 198 to not knowing precipitation, and instead transfers the uncertainty to the interpretation of the
 199 combined input quantity hindcast. These four approaches all exploit the fact, reported by
 200 Mohammed and Tarboton (2012), that GSL Change in Volume is significantly more sensitive to
 201 Streamflow variability than Precipitation or Evaporation variability and, therefore, that the
 202 sensitivity of hindcast Inflow to errors introduced through approximating Precipitation and
 203 Evaporation is likely to be relatively small. The uncertainty in these inflow hindcasts was
 204 quantified through comparison of their estimates to actual streamflow and, in the case of the
 205 runoff ratio based precipitation hindcast, to PRISM precipitation over the periods when there is
 206 streamflow and precipitation data. Common hydrologic performance metrics (Nash-Sutcliffe
 207 Efficiency, Kling-Gupta Efficiency, Root Mean Square Error) were used to quantify this
 208 uncertainty. These hindcasts were then used to infer how lake inputs (either streamflow or
 209 streamflow plus precipitation) have reduced since lake levels were first measured, presumably
 210 due to the development of water resources and CU. In each of the subsections below, we give
 211 details of the calculations and equations involved.

212 3.1 GSL Specific Water Mass Balance

213 To fully describe the process of developing a hindcast estimate of the Inflow specific to
 214 streamflow into GSL, a more detailed expression of Equation 1 is shown in Equation 3.

$$\Delta V = V_i(WSE_i) - V_{i-1}(WSE_{i-1}) = Q_S + Q_G + Q_{ME} + Q_{WD} + P_L * A_L - E_L * A_L \quad (3)$$

215 As expressed here, ΔV = Change in Volume; Inflow has been separated into Q_S = surface inflow
 216 (streamflow) volume, Q_G = groundwater flow volume, Q_{ME} = mineral extraction pumping
 217 (negative) and return flow volume (positive), and Q_{WD} = west desert pumping (negative) and
 218 return flow volume (positive); Precipitation has been expressed as P_L = precipitation depth over
 219 the lake times A_L = area of the lake; and Evaporation as E_L = evaporation depth over the lake
 220 times A_L . Lake volume and lake area are both functions of WSE through the bathymetry of the
 221 lake; WSE is the quantity that has been measured over time. Equation 3 can be reformulated to
 222 solve for surface inflow volume, Q_S .

$$Q_S = \Delta V - Q_G - Q_{ME} - Q_{WD} - P_L * A_L + E_L * A_L \quad (4)$$

223 This formulation can be used to estimate the hindcast of historical inflow to GSL once data,
 224 assumptions, and approximations for variables on the right-hand side are established. In this
 225 formulation, ΔV and A_L are calculated directly from WSE measurements back to 1847. Constant
 226 Groundwater Inflow, Q_G , is estimated based on prior work (Waddell and Fields, 1977). Mineral
 227 extraction, Q_{ME} , and west desert pumping, Q_{WD} , are known quantities and are 0 prior to when
 228 they began in 1960 and 1987, respectively. This leaves the variables P_L and E_L to be estimated in
 229 order to use Equation 4 for hindcasting. Flow quantities and lake areas were evaluated using
 230 annual or 12-month average data for the U.S. water year starting October 1. Lake volumes were
 231 taken as end of water year values so that ΔV was a water year difference associated with the
 232 period over which the flows were averaged.

233 3.2 Inflow

234 In Equation 4, inflow to GSL is separated into streamflow volume (Q_S), groundwater
 235 flow volume (Q_G), mineral extraction pumping and return flow volume (Q_{ME}), and west desert
 236 pumping and return flow volume (Q_{WD}). We used streamflow data from the United States
 237 Geological Survey (USGS), <https://maps.waterdata.usgs.gov/mapper/> (accessed 1 June 2023).
 238 Missing data was filled in using regression with nearby stations to create a complete $Q_{S,measured}$
 239 dataset for 1950-2023 using methods documented by Tarboton (2023) that were adapted from
 240 Loving et al. (2000) and Mohammed and Tarboton (2012). Waddell and Fields (1977) estimated
 241 annual GSL groundwater inflow to be $Q_G = 0.0925 \text{ km}^3/\text{year}$. We took this quantity as a constant
 242 for the entire 1848-2023 period due to the lack of information on temporal variability of
 243 groundwater inflows. There is uncertainty in this assumption, but errors due to this assumption
 244 are expected to be small due to the small overall contribution from groundwater. The West
 245 Desert Pumping Project circulated water out of GSL from 1987-1989 and then back to GSL from
 246 1990-1992. Flow estimates for Q_{WD} were obtained from Loving et al. (2000). Mineral extraction
 247 in GSL started in the 1960s. Flow rates for Q_{ME} were estimated based on withdrawals reported
 248 by mineral industries to the Utah Division of Water Rights. More detailed information can be
 249 found on the Q_{WD} and Q_{ME} time series in Merck and Tarboton (2023).

250 3.3 Area and Volume

251 Lake area and lake volume are functions of WSE. They were tabulated in 0.15 m (0.5
 252 foot) increments, for elevations from 1269.5 m to 1286.45 m, from the lakebed bathymetry
 253 digital elevation model (Tarboton and Merck, 2023) using methods detailed in Merck and
 254 Tarboton (2023). The historical time series of WSE was used to interpolate time series of A_L , V ,
 255 and hence ΔV needed for our analysis from the lake area and volume bathymetry tables. This
 256 paper used the total volume in each arm of the lake, including areas now separated into mineral
 257 extraction evaporation ponds because, at the times the lake was high, these pond areas were
 258 effectively part of the lake due to either being breached or not yet being constructed and our
 259 focus is on inferring what inflows were prior to pond construction.

260 3.4 Evaporation

261 Measured evaporation data is not available for GSL. However, all other terms in
 262 Equation 3 are available from 1949 to present. Thus, Equation 3 can be rearranged to calculate
 263 evaporation for the years 1950-2023 using mass balance closure.

$$E_{L,MB} = \frac{Q_S + Q_G + Q_{ME} + Q_{WD} + P_L * A_L - \Delta V}{A_L} \quad (5)$$

264 Recognizing that evaporation is sensitive to salinity, which changes from year to year, the
 265 evaporation term is further expressed as

$$E_{L,MB} = E_{f,MB} * SCF \quad (6)$$

266 where $E_{f,MB}$ = mass balance freshwater evaporation and SCF = salinity correction factor, which
 267 reduces freshwater evaporation based on salinity. The freshwater evaporation depth over the
 268 lake, $E_{f,MB}$, can be calculated for 1950-2023 using Equation 6 once the SCF has been calculated.
 269 Mohammed and Tarboton (2012) developed a modification to the Penman method for
 270 calculating lake evaporation based on salinity using an activity coefficient to reduce saturation
 271 vapor pressure. This was used to calculate GSL evaporation over the period of their study.
 272 Comparing their salinity dependent evaporation and freshwater evaporation, we developed the
 273 following empirical equation for SCF as a function of salinity, S .

$$SCF = -1 \times 10^{-11} * S^4 + 3 \times 10^{-9} * S^3 - 2 \times 10^{-4} * S + 1 \quad (7)$$

274 Salinity is required to use Equation 7 and was calculated based on salt mass divided by
 275 lake volume, $S = M_{salt}/V_{Lake}$, where V_{Lake} is a function of bathymetry and WSE. Merck and
 276 Tarboton (2023) estimated GSL salt mass based on salinity and volume measurements for the
 277 period 1966 to present. Salt mass prior to 1966 was inferred based on the salinity of river inputs
 278 (Hahl, 1968). The mean $E_{f,MB}$ for 1950-2023, $E_{f,mean}$, was used along with estimated SCF to
 279 calculate the times series of evaporation depth over the lake, E_L , for 1848-2023.

$$E_L = E_{f,mean} * SCF \quad (8)$$

280 3.5 Precipitation

281 Gridded data for precipitation depth are available with a 4 km resolution for the years
 282 1895 to present from the PRISM Climate Group based at Oregon State University
 283 (<https://prism.nacse.org>, accessed 1 June 2023). PRISM uses interpolation based on
 284 physiographic similarity to estimate precipitation from discontinuous and intermittent point
 285 observations (Daly et al., 2008). PRISM estimates do suffer from uncertainty, especially during
 286 the early years, when the underlying point observations are sparse. There are also other gridded
 287 precipitation data sources (e.g., NLDAS, DAYMET, gridMET) available in easy-to-use format
 288 from sites such as climateengine.org (Huntington et al., 2017), but these start no earlier than
 289 1950. We performed spot check comparisons of PRISM precipitation versus some of these other
 290 sources, and versus point observations for some long running precipitation stations in the GSL
 291 basin, and found differences to be generally small, though there were some sites with notable
 292 biases. We decided that relying on the specialized functionality built into PRISM methods for
 293 interpolation and addressing shortcomings due to missing data was preferable to us attempting
 294 our own interpolation from point precipitation observations. Thus, our analysis has used PRISM
 295 data.

296 PRISM precipitation data were extracted for the area of the three river basins (Bear,
 297 Weber and Jordan) that contribute to streamflow into GSL, $P_{B,PRISM}$, and the area over the lake,
 298 $P_{L,PRISM}$, for the full duration of the PRISM dataset. A hindcast of precipitation was still needed
 299 for 1848-1895, the years for which PRISM is not available, and, as noted above, we evaluated
 300 two options: (1) constant precipitation at its long-term average value; and (2) lake precipitation
 301 estimated from runoff ratio and the ratio with basin precipitation being assumed constant.

302 Detailing the second approach, streamflow volume in a basin is related to the
 303 precipitation volume over that basin through the runoff ratio, r , expressed as

$$r = \frac{Q_S}{P_B * A_B} \quad (9)$$

304 where Q_S = streamflow volume in the basin, P_B = precipitation depth over the basin, and A_B =
 305 35,637 km², the area of the combined Bear, Weber, and Jordan River basins. While r varies year
 306 to year, an average value for r was estimated based on years 1950-2023, the time span data is
 307 available for all three variables in Equation 9. Another assumption is that the precipitation depth
 308 over the basin and the precipitation depth over the lake are linearly related

$$P_B = a * P_L \quad (10)$$

309 where a = basin to lake precipitation ratio and P_L = precipitation depth over the lake. A value for
 310 a was estimated based on PRISM data, $P_{B,PRISM}$ and $P_{L,PRISM}$, for the years 1896-2023. Average
 311 values for r and a were substituted into Equation 3 and then solved for precipitation depth over
 312 the lake

$$P_{L,ra} = \frac{\Delta V - Q_G - Q_{ME} - Q_{WD} + E_{f,mean} * SCF * A_L}{(r * a * A_B + A_L)} \quad (11)$$

313 All variables on the right side of Equation 11 now have either measured data or have been
 314 estimated for the years 1848-2023 and can be used to estimate a time series of precipitation depth
 315 over the lake, $P_{L,ra}$.

316 3.6 GSL Inflow Hindcasts

317 To supplement the USGS measured streamflow record, $Q_{S,measured}$, Equation 4 was used to
 318 estimate two hindcast time series for inflow to GSL for the full period of record, 1848-2023,
 319 using the two different assumptions about precipitation over the lake, and another inflow
 320 hindcast was estimated using PRISM precipitation data for 1896-2023, the years PRISM data is
 321 available. Note that both Q_{WD} and Q_{ME} are zero for years 1848-1949, the time period over which
 322 the hindcasts of inflow to GSL will be used to supplement measured data, but both variables
 323 have been retained in the equations that follow for the purpose of validating hindcast estimates
 324 for the years 1950-2023.

325 The first hindcast inflow to GSL, $Q_{S,Pmean}$ or just the $Pmean$ hindcast, was calculated for
 326 1848-2023 using Equation 4 based on the assumption of constant depths for precipitation and
 327 freshwater evaporation over the lake. The mean of the PRISM time series data for the years
 328 1896-2023, $P_{L,mean}$, was used for precipitation depth, P_L . The mean of the mass balance
 329 freshwater evaporation depth for the years 1950-2023, $E_{f,mean}$, was used for freshwater
 330 evaporation depth, E_f , along with estimated SCF for the years 1848-2023. Substituting these
 331 variables into Equation 4 we get

$$Q_{S,Pmean} = \Delta V - Q_G - Q_{ME} - Q_{WD} - P_{L,mean} * A_L + E_{f,mean} * SCF * A_L \quad (12)$$

332 The second hindcast inflow to GSL, $Q_{S,ra}$ or just the *ra* hindcast, was also calculated for
 333 1848-2023 using constant freshwater evaporation depth over the lake in Equation 4. However,
 334 this hindcast used the assumption that the runoff coefficient, r , and precipitation ratio, a , are both
 335 constant over the years 1848-2023, and therefore the hindcast precipitation, $P_{L,ra}$, was used for
 336 precipitation, P_L . Substituting into Equation 4 we get

$$Q_{S,ra} = \Delta V - Q_G - Q_{ME} - Q_{WD} - P_{L,ra} * A_L + E_{f,mean} * SCF * A_L \quad (13)$$

337 A third hindcast inflow to GSL, $Q_{S,PRISM}$ or just the *PRISM* hindcast, was calculated using
 338 *PRISM* data, $P_{L,PRISM}$, and constant freshwater evaporation depth over the lake in Equation 4 for
 339 the dates *PRISM* data is available, 1896-2023. Substituting into Equation 4 we get

$$Q_{S,PRISM} = \Delta V - Q_G - Q_{ME} - Q_{WD} - P_{L,PRISM} * A_L + E_{f,mean} * SCF * A_L \quad (14)$$

340 The *PRISM* hindcast was used in both reconstructions of inflow to GSL for the years 1896-1949,
 341 the years *PRISM* data are available but streamflow data is not.

342 3.7 Primary Inputs Hindcast: $Q + P$ Using ΔV and Constant E_L

343 Equation 3 can be rearranged to calculate the primary inputs to GSL, inflow and
 344 precipitation, using mass balance closure as follows

$$Q_S + P_L * A_L = \Delta V + E_L * A_L - Q_G - Q_{ME} - Q_{WD} \quad (15)$$

345 All variables on the right-hand side of Equation 15 are known or have reasonable approximations
 346 for the years 1848-2023. This fourth hindcast provides information on the total input to the lake
 347 without the uncertainties introduced by precipitation assumptions and serves as another line of
 348 evidence quantifying reduced inputs.

349 3.8 GSL Inflow Reconstructions

350 Inflows to GSL were reconstructed using inflow hindcasts and measured streamflow. The
 351 *Pmean* and *ra* hindcasts were used for 1848-1895, the years there is no precipitation,
 352 evaporation, or streamflow data available. The *PRISM* hindcast was used for 1896-1949, the
 353 years there is *PRISM* precipitation data available but no evaporation or streamflow available.
 354 And measured streamflow was used for 1950-2023.

$$Q_{R,Pmean} = \text{concatinate}(Q_{S,Pmean}(1848 - 1895), \quad (16)$$

$$Q_{S,PRISM}(1896 - 1949), Q_{measured}(1950 - 2023))$$

$$Q_{R,ra} = \text{concatinate}(Q_{S,ra}(1848 - 1895), \quad (17)$$

$$Q_{S,PRISM}(1896 - 1949), Q_{measured}(1950 - 2023))$$

355 These inflow reconstructions were used to determine the associated input reductions, natural
 356 flow, and WSE for GSL.

357 3.9 Consumptive Use and Natural Flow

358 The natural flow of a stream, Q_N , is equal to the measured streamflow, Q_S , plus the
 359 reductions due to water depletions for its basin, Q_D .

$$Q_N = Q_S + Q_D \quad (18)$$

360 Historical streamflow in the GSL Basin measured at points above diversions, or tree-ring
 361 reconstructed flows not subject to diversions that quantify the natural flow of the basin, does not
 362 show long-term trends. Therefore, any long-term, ongoing trend in the historical time series of
 363 inflow to GSL can be attributed to human water use depletions within the GSL Basin. Mean
 364 inflows to GSL were determined for the first and last 30 years of the historical time series in
 365 order to establish starting and ending points for determining input reductions. There is some
 366 arbitrariness in this selection of 30 years, but it is a common averaging period used for climate
 367 averages as it spans a period longer than the typical annual and decadal frequencies present in
 368 many hydrologic and climate series. A 30-year moving average was used to determine the
 369 general trend of input reductions in the GSL Basin, $Q_{D,Pmean}$, $Q_{D,ra}$, $Q_{D,Q+P}$, for their respective
 370 inflow reconstructions. Note that the input reductions associated with the primary inputs hindcast
 371 were determined using the same methods as the reductions associated with the reconstructed
 372 inflows, with early and late means for start and end points and trends for decreasing primary
 373 inputs.

374 3.10 Natural Water Surface Elevation

375 The water mass balance in Equation 3 was solved using implicit finite differences to
 376 calculate what the lake volume and corresponding WSE would have been given reconstructed
 377 natural flows, $Q_{N,Pmean}$ and $Q_{N,ra}$. With an initial WSE of $WSE_{i=0} = 1280.267$ meters on 1847-10-
 378 01 (the beginning of water year 1848), the inflow, precipitation, and evaporation for each water
 379 year, i , were used to step ahead to a new lake volume, V_i , and corresponding WSE_i at the end of
 380 the water year. Evaluation of the fluctuations in the annual WSE indicate that end of water year
 381 $WSE_i + 0.15$ meters better represents the mean annual WSE over a water year, and therefore this
 382 value was used to determine $A_{L,i}$ in the calculation of area dependent precipitation and
 383 evaporation volumes. SCF was calculated based on the lake volume of the previous water year.
 384 The natural flow and associated WSE were not calculated for the primary inputs hindcast
 385 because it has built into it the historical area of the lake (A_L is on the left-hand side of equation
 386 15) and adding estimated depletion to this would not account for the effect of larger lake area
 387 associated with the resulting higher lake levels.

388 3.11 Hindcast Performance, Validation and Error

389 The metrics used to evaluate and validate hindcast performance and estimate error
 390 include the coefficient of variation (CV), Root Mean Square Error ($RMSE$), normalized RMSE
 391 ($nRMSE$), Nash-Sutcliffe Efficiency (NSE), and Kling-Gupta Efficiency (KGE). The formulas
 392 for each are as follows:

$$CV = \frac{\sigma}{\mu} \quad (19)$$

$$RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (S_i - O_i)^2} \quad (20)$$

$$nRMSE = \frac{RMSE}{\mu_{obs}} \quad (21)$$

$$NSE = 1 - \frac{\sum_{i=1}^n (S_i - O_i)^2}{\sum_{i=1}^n (O_i - \bar{O})^2} \quad (22)$$

$$KGE = 1 - \sqrt{(r - 1)^2 + \left(\frac{\sigma_{sim}}{\sigma_{obs}} - 1\right)^2 + \left(\frac{\mu_{sim}}{\mu_{obs}} - 1\right)^2} \quad (23)$$

393 where σ = standard deviation; μ = mean; S = the simulated value; O = the observed value; and r
 394 = Pearson correlation coefficient. The CV was used to evaluate the performance of the
 395 evaporation hindcasts while the $RMSE$, $nRMSE$, NSE , and KGE were used to evaluate the
 396 precipitation and validate inflow hindcasts. In order to determine which hindcast to use in the
 397 final inflow reconstruction during certain time periods, all three inflow hindcasts were compared
 398 with measured inflow for a validation period, 1950-2023, while the P_{mean} and ra hindcasts were
 399 compared with the $PRISM$ hindcast for a test period, 1896-1949.

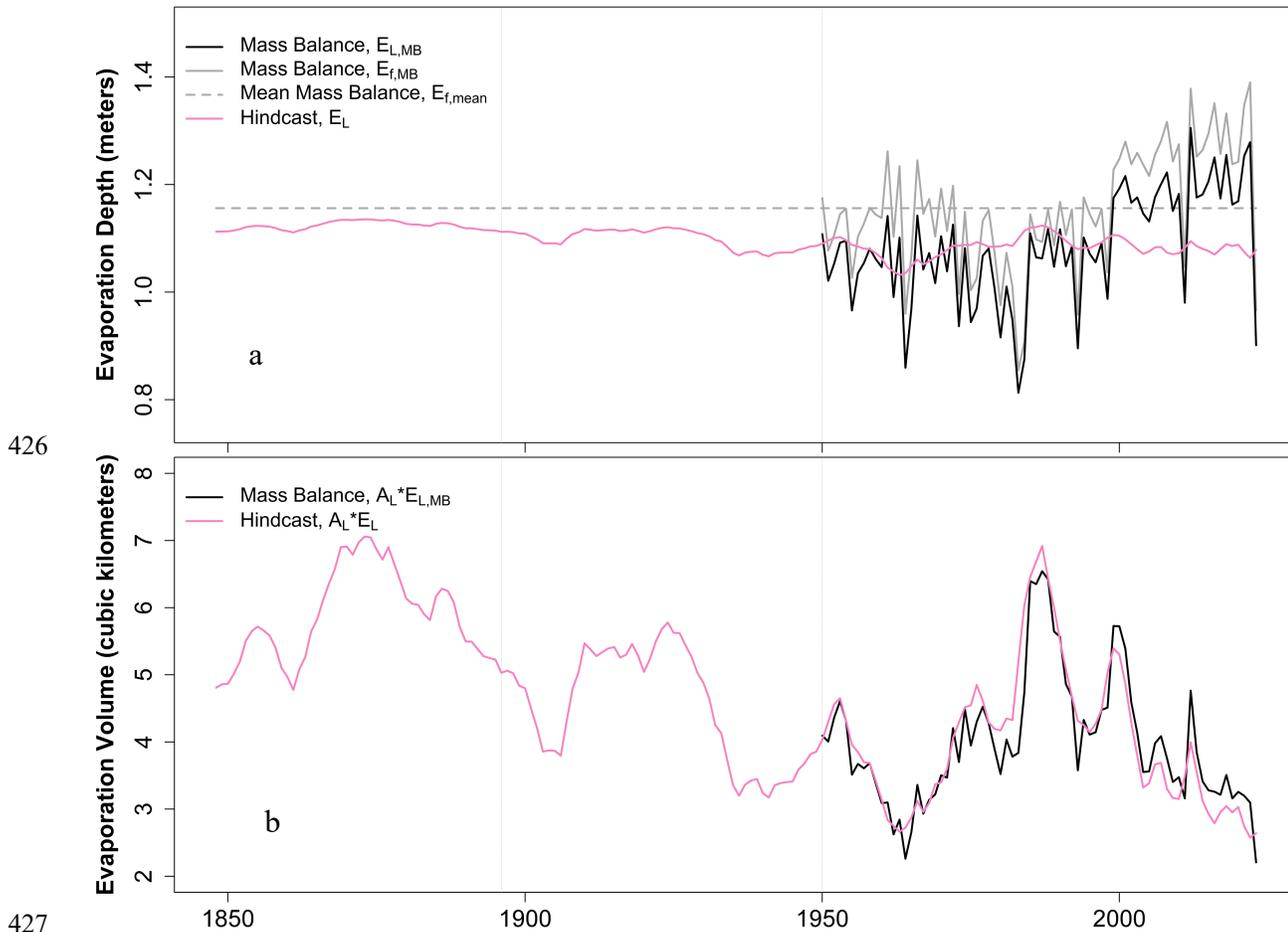
400 4 Results

401 4.1 Evaporation

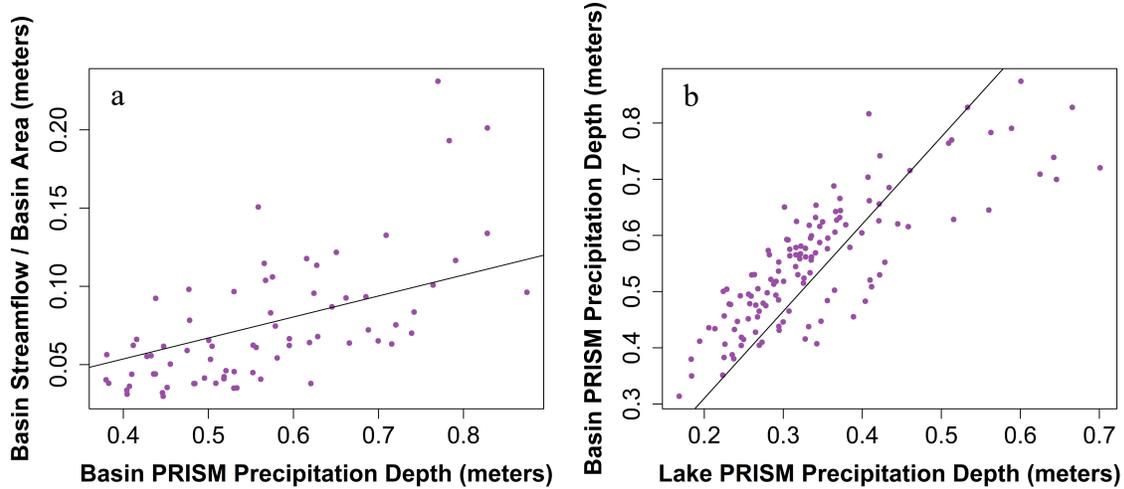
402 Evaporation depths and volumes over GSL are shown in Figure 4. The mass balance time
 403 series of evaporation depth, $E_{L,MB}$, for the years all measured data were available, 1950-2023
 404 (Figure 4a, black solid line), was found to have a mean annual depth of 1.08 m. The mass
 405 balance freshwater evaporation depth, $E_{f,MB}$ (Figure 4a, gray solid line), has mean, $E_{f,mean}$, of 1.16
 406 m (Figure 4a, gray dashed line) and standard deviation 0.1 m, reflecting a CV of less than 0.1.
 407 This small CV supports the approximation of E_f as a constant (i.e., $E_f = E_{f,mean} = 1.16$ m) for use
 408 in Equation 8 when calculating the hindcast evaporation depth over the lake, E_L (Figure 4a, pink
 409 solid line).

410 4.2 Precipitation

411 $PRISM$ precipitation depths were used along with measured USGS streamflow in
 412 Equation 9 for years both datasets were available, 1950-2023, to estimate the average GSL Basin
 413 runoff ratio, $r = 0.13$ (Figure 5a), and in Equation 10 for years 1896-2023 to estimate the basin to
 414 lake precipitation ratio, $a = 1.55$ (Figure 5b). The mean annual $PRISM$ precipitation depth over
 415 the lake, $P_{L,Pmean}$, is 0.341 m (Figure 6a, gray line). Based on the assumption of constant values
 416 for r , a , and $E_{f,mean}$, a hindcast times series was calculated for precipitation depth over the lake,
 417 $P_{L,ra}$ (Figure 6a, purple line), for the years from 1848-2023 using Equation 11, which is driven by
 418 WSE that inform lake volume change, lake area, and SCF. The mean depth of $P_{L,ra}$ for the years
 419 1896-2023 is 0.363 m, which is 6.5% more than the mean $P_{L,PRISM}$ (Figure 6a, black line) for the
 420 same period. Comparing evaluation metrics for $P_{L,ra}$ and $P_{L,PRISM}$ on the overlapping years, the
 421 $RMSE$ is 0.122 m, reflecting a normalized $RMSE$ of 36%, which is rather high. Interestingly, the
 422 NSE is -0.33, indicating that $P_{L,Pmean}$ is a better predictor of precipitation depth over the lake than
 423 $P_{L,ra}$, while the KGE is 0.44, indicating $P_{L,ra}$ is a reasonable predictor of precipitation depth over
 424 the lake. Both $P_{L,ra}$ and $P_{L,Pmean}$ were used to estimate hindcasts of inflow to GSL and these
 425 metrics for the precipitation should be held in mind when interpreting each inflow hindcast.

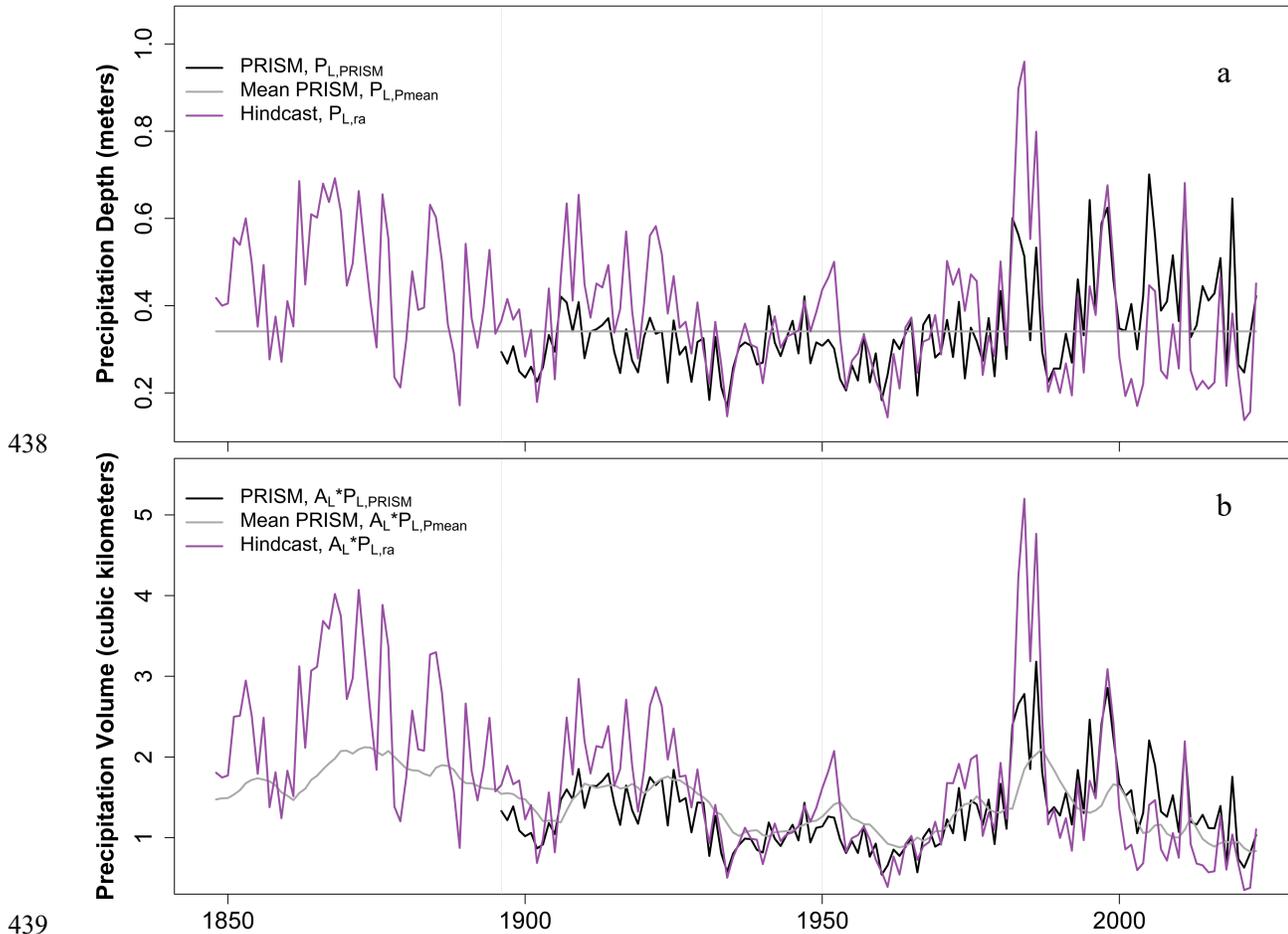


428 **Figure 4.** Evaporation (a) depths and (b) volumes over GSL. Results using measured data for
 429 mass balance closure over the years 1950-2023 include mass balance freshwater evaporation
 430 depth ($E_{f,MB}$, gray solid line), its mean of 1.16 m ($E_{f,mean}$, gray dashed line), and mass balance
 431 evaporation depth ($E_{L,MB}$, black solid line). Hindcast results for the years 1848-2023 include
 432 evaporation depth over the lake (E_L , pink solid line).



433

434 **Figure 5.** (a) Streamflow volume and precipitation volume in the GSL Basin and the linear
435 relationship between the two (black line) representing the runoff ratio, $r = 0.13$ (Equation 9). (b)
436 Precipitation depth over the GSL Basin and lake and the linear relationship between the two
437 (black line) representing the basin to lake precipitation ratio, $a = 1.55$ (Equation 10).



439

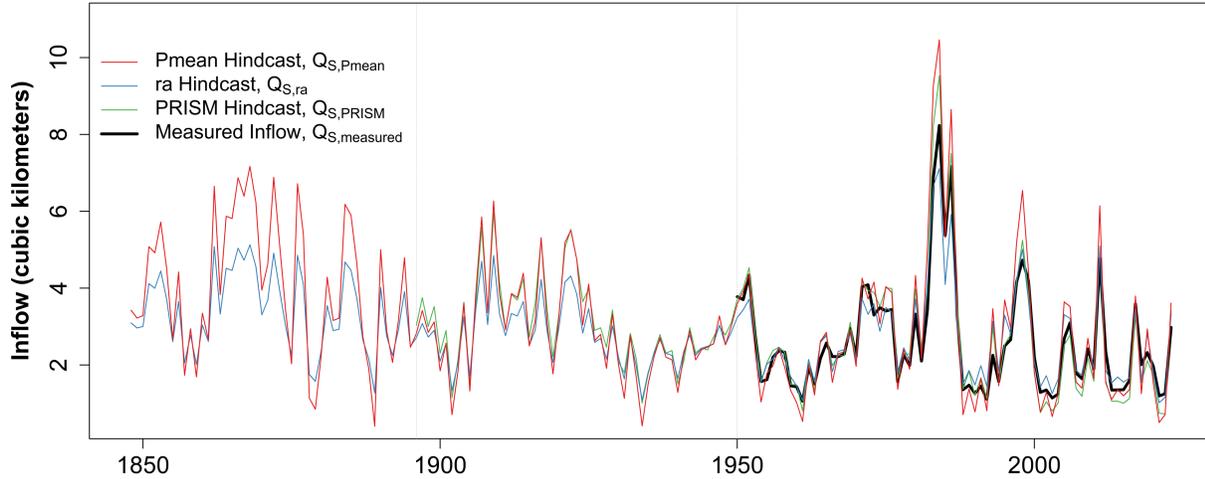
440 **Figure 6.** Precipitation depths (a) and volumes (b), including PRISM data over the lake (black
 441 solid line), mean PRISM data over the lake (gray solid line), and calculated depth over the lake
 442 using constant r and a (purple solid line).

443

4.3 Inflow Hindcasts

444

445 Inflow hindcast results are shown in Figure 7. All three inflow hindcasts were compared
 446 with measured inflow (black line) in a validation period, 1950-2023, and with the *PRISM*
 447 hindcast (green line) in a test period, 1896-1949. All evaluation metrics establish the *ra* hindcast
 448 (blue line) as a more accurate hindcast than the *Pmean* hindcast (red line) for the validation
 449 period, 1950-2023 (Table 1, “Validation” columns), and the *Pmean* hindcast as a more accurate
 450 hindcast than the *ra* hindcast for the test period, 1896-1949 (Table 1, “Test” columns). These
 451 metrics for the inflow hindcasts should be held in mind when interpreting each inflow
 reconstruction.



452

453 **Figure 7.** Inflow hindcast results for $Pmean$ ($Q_{S,Pmean}$, red line), ra ($Q_{S,ra}$, blue line), and $PRISM$
 454 ($Q_{S,PRISM}$, green line). Measured inflow (black line) is included for comparison.

455

456 **Table 1.** Evaluation metrics for the inflow hindcasts in Figure 7 for the periods 1896-1949 and
 457 1950-2023.

458

Evaluation Metric	Test (1896-1949)		Validation (1950-2023)		
	$Pmean$	ra	$Pmean$	ra	$PRISM$
NSE	0.93	0.81	0.76	0.92	0.92
KGE	0.86	0.70	0.61	0.86	0.82
$RMSE$	0.28 m	0.47 m	0.69 m	0.41 m	0.39 m
$nRMSE$	9%	15%	27%	16%	15%

459

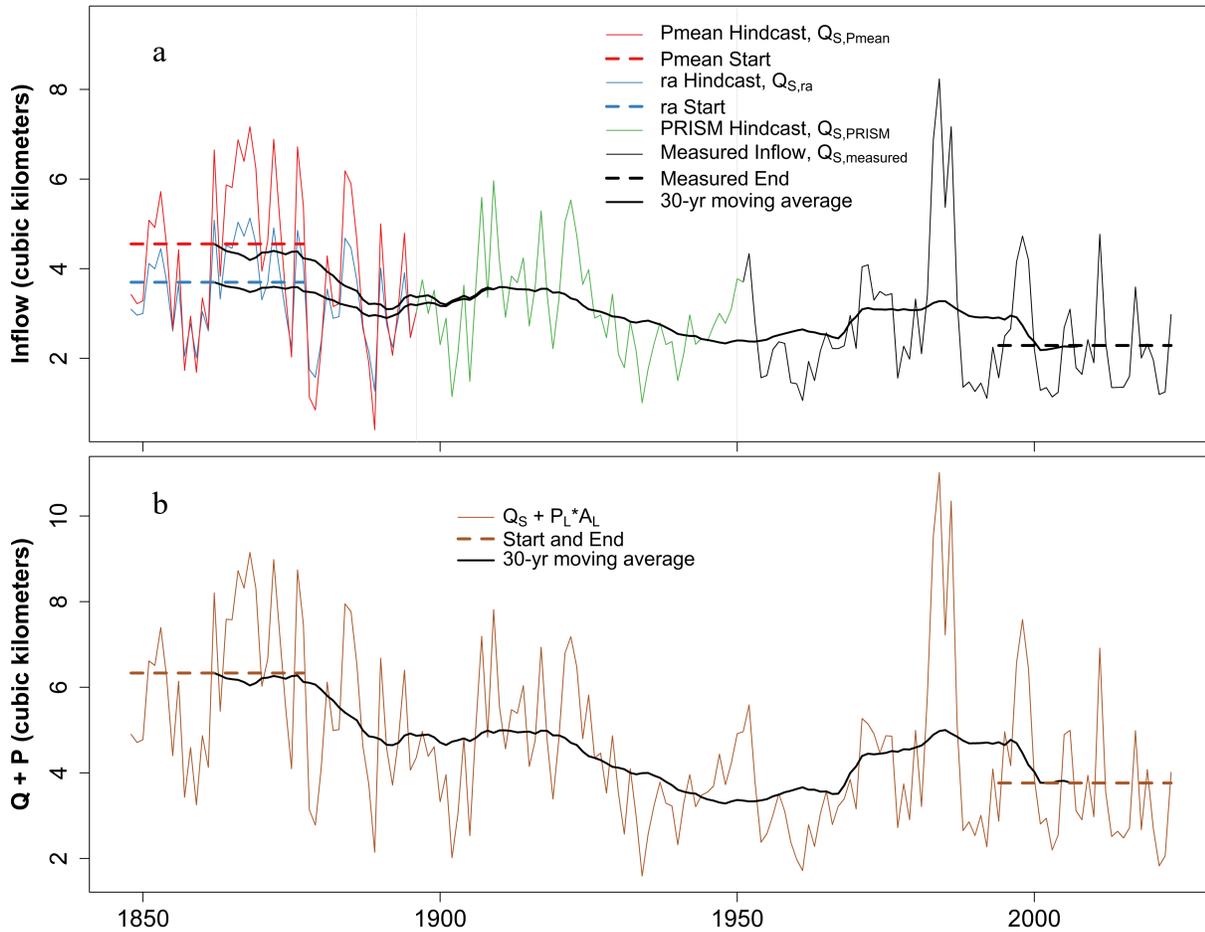
460 4.4 Input Reductions Inferred from Reconstructions and Hindcasts

461 Two reconstructed times series were developed for inflow to GSL, $Q_{R,Pmean}$ and $Q_{R,ra}$
 462 (Equations 16 and 17, respectively) using the respective $Pmean$ and ra inflow hindcasts for
 463 1848-1895, the PRISM inflow hindcast, $Q_{S,PRISM}$, for 1896-1949, and the measured streamflow,
 464 $Q_{S,measured}$, for 1950-2023. The $Pmean$ and ra reconstructed inflows are shown in Figure 8a,
 465 specifying the individual components, $Pmean$ hindcast (red solid line), ra hindcast (blue solid
 466 line), $PRISM$ hindcast (green solid line), and measured streamflow (black solid line). Also
 467 included are the 30-year moving averages (black solid lines). Note the moving averages are
 468 separate for $Pmean$ and ra for 1848-1895 and then overlap for the remainder of the
 469 reconstruction where both use the $PRISM$ inflow hindcast and measured streamflow. Also
 470 indicated are the starting and ending points of the moving average represented by 30-year spans
 471 (horizontal dashed lines).

472 Similarly, a hindcast of primary inputs to GSL, $Q_{S+PL} * A_L$, was calculated for 1848-2023
 473 using Equation 15. As with the inflow hindcasts, change in lake volume and inflows due to

474 groundwater, mineral extraction, and west desert pumping are known volumes for the entire
475 period and therefore predetermined, and the mean freshwater mass balance evaporation, $E_{f,mean}$,
476 was also used in Equation 15. The results for the primary inputs hindcast are shown in Figure 8b
477 along with the 30-year moving average (black solid line) as well as the 30-year spans
478 representing the starting and ending points of the moving average (horizontal dashed lines).

479 The reduced input, representing changes to inflow into GSL between 1848 and 2023, was
480 inferred by taking the difference between the starting and ending points, defined by the 30-year
481 spans (dashed horizontal lines) for each reconstruction method. The decline in the 30-year
482 moving averages seen in P_{mean} and ra in Figure 8a and the primary inputs in Figure 8b
483 approximate the input reductions due to water depletions over the historical record. These input
484 reductions were used to determine time series for the P_{mean} and ra reconstructions, $Q_{D,P_{mean}}$ and
485 $Q_{D,ra}$, and the primary inputs hindcast, $Q_{D,Q+P}$. Starting points of $Q_D = 0$ in 1848 were used based
486 on the respective starting point mean flows for P_{mean} (red dashed line, Figure 8a), ra (blue
487 dashed line, Figure 8a), and primary inputs flows (brown dashed line, Figure 8b). Ending points
488 in 2023 were used based on the recent mean measured inflow (black dashed lines, Figures 8a and
489 b). The Q_D in all three time series were increased from $Q_D = 0$ in 1848 to their respective recent
490 magnitudes by 1960, $Q_D = Q_{start} - Q_{end}$, based on the rate inferred by the general trend and shape
491 of the 30-year moving averages. The resulting current depletion magnitudes, Q_D , are 2.26
492 km^3/yr , 1.41 km^3/yr , and 2.60 km^3/yr for P_{mean} , ra , and the primary inputs, respectively (Figure
493 9). Note that the selection of 1960 as the year to switch from increasing consumptive use to flat
494 consumptive use is based on the 30 years means in Figure 8 essentially flattening out by then.
495 While recognizing the uncertainty in selection of this point is about plus or minus 15 years, this
496 uncertainty does not tangibly affect our results.

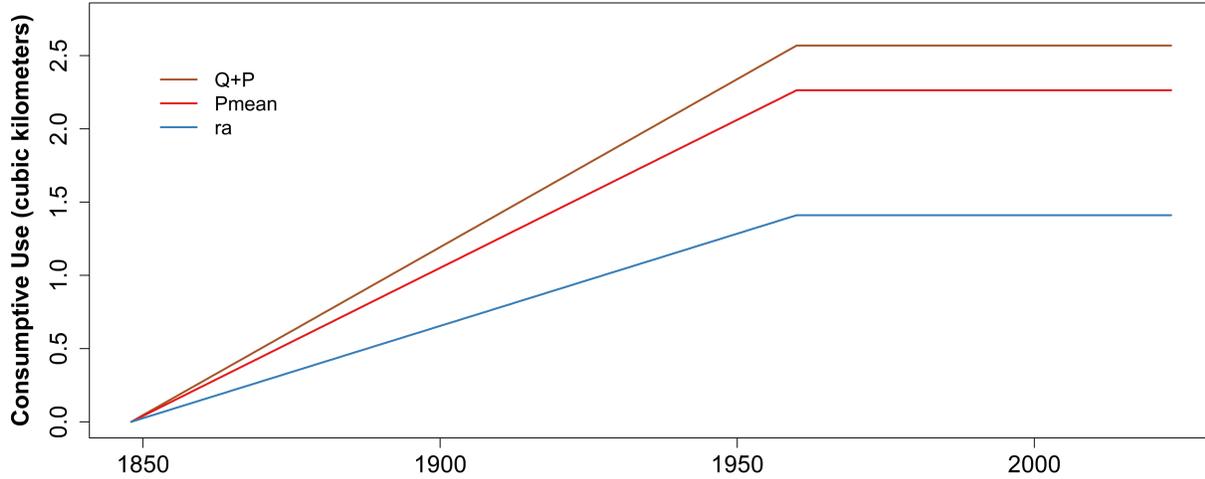


497

498

499 **Figure 8.** (a) Inflow reconstructions using $Q_{S,Pmean}$ (red line) and $Q_{S,ra}$ (blue line) for 1848-1895,
 500 $Q_{S,PRISM}$ (green line) for 1896-1949, and $Q_{S,measured}$ (black line) for 1950-2023. Also included are
 501 the 30-year moving averages (black lines) and the 30-year spans representing the start and end of
 502 the moving average (dashed lines). (b) Primary inputs hindcast (brown line) with 30-year moving
 503 average and 30-year spans representing start and end of the moving average (dashed lines).

504



505

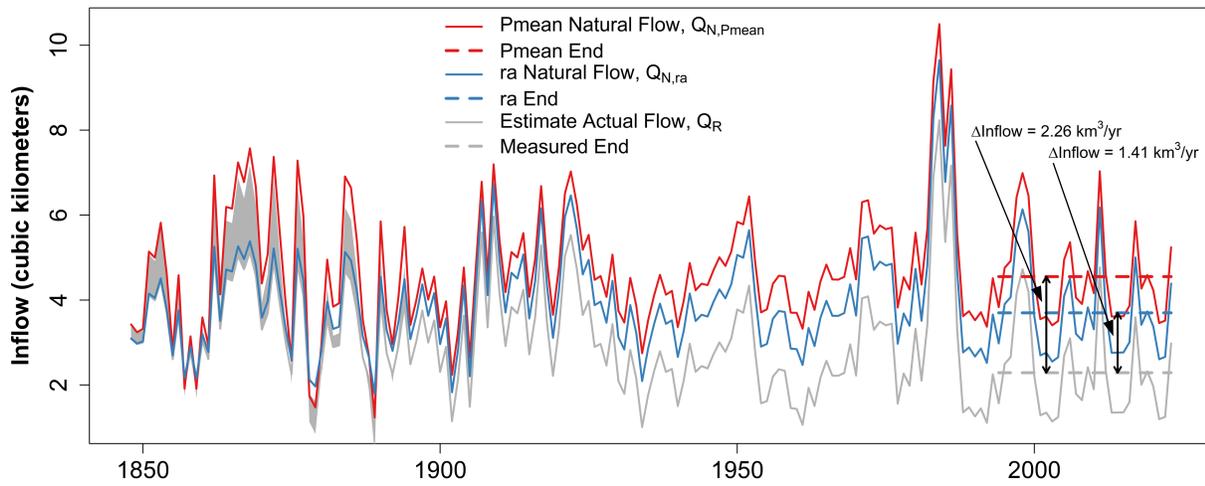
506 **Figure 9.** Input reductions inferred from reconstructions and hindcasts are 2.26 km³/yr, 1.41
 507 km³/yr, and 2.60 km³/yr for *Pmean*, *ra*, and the primary inputs, respectively.

508

4.5 Natural Inflow Reconstructions and Associated Water Surface Elevation

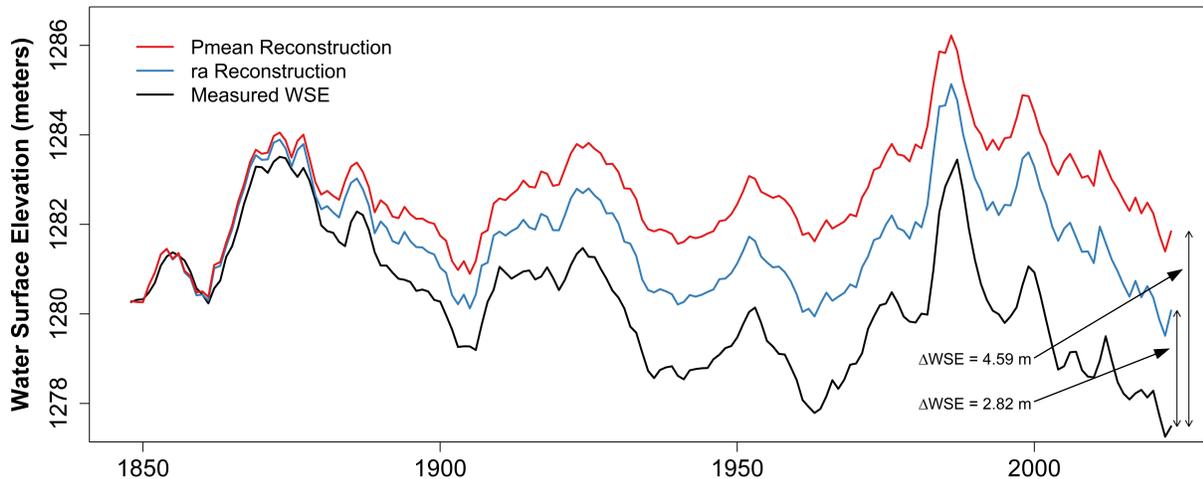
509

510 The Q_D from Figure 9 were added to the respective *Pmean* and *ra* reconstructed flows in
 511 Figure 8a based on Equation 18 to develop two estimated time series of natural flow into GSL,
 512 $Q_{N,Pmean}$ and $Q_{N,ra}$ (Figure 10). By construction, neither the *Pmean* nor *ra* time series for natural
 513 flow have a trend. The time series of estimated WSE corresponding to the *Pmean* and *ra* natural
 514 flows are presented in Figure 11 and show the current decline in WSE due to reduced inflow to
 be 4.59 m and 2.82 m, respectively.



515

516 **Figure 10.** Reconstructed natural inflows to GSL $Q_{N,Pmean}$ (red line) and $Q_{N,ra}$ (blue line), not
 517 including depletions, shown in comparison to the respective reconstructed inflows (gray line
 518 with shading where they diverge prior to 1896).



519

520 **Figure 11.** Great Salt Lake water surface elevation based on respective reconstructed natural
 521 inflows, $Q_{N,Pmean}$ (red line) and $Q_{N,ra}$ (blue line), shown in comparison to USGS measured water
 522 surface elevation (black line).

523 4 Discussion

524 The time series of mass balance evaporation depth over GSL, $E_{L,MB}$ (dotted black line,
 525 Figure 4a), is generally consistent with the mass balance results reported in Figure 10 of
 526 Mohammed and Tarboton (2012), and both time series produce a mean annual evaporation depth
 527 of approximately 1.1 m. The time series of mass balance freshwater evaporation depth, $E_{f,MB}$
 528 (solid gray line, Figure 4), calculated using Equation 5 has mean depth, $E_{f,MBmean}$, of 1.16 m,
 529 which differs from the mean annual freshwater evaporation depth of 1.29 m that Mohammed and
 530 Tarboton calculated using climate data. They point out that GSL WSE modeled using a
 531 freshwater evaporation depth of 1.29 m is consistently lower than observed WSE (Figure 11 of
 532 their publication), suggesting that 1.29 m may be an overestimate. The calculations in
 533 Mohammed and Tarboton are based on data through 2008, including estimated bathymetry and
 534 PRISM modeling methods that have both since been refined. In addition, since the publication of
 535 their work, there have been refinements to streamflow datasets and the state equation for
 536 estimating GSL salinity, which affect measured streamflow and the calculation of SCF and
 537 therefore the mass balance. It is for these reasons that we recalculated the value of $E_{f,MBmean}$ for
 538 use in our mass balance hindcasts rather than using their published values.

539 Overall, the metrics comparing the $Pmean$, ra , and $PRISM$ inflow hindcasts indicate that
 540 all hindcast methods give a good prediction of inflow that enables interpretation of how inflows
 541 have been reduced over the historical record of WSE measurement and the associated
 542 uncertainty (Figure 7, Table 1). In evaluating these, we feel that the $Pmean$ hindcast best
 543 captures long-term average lake inflow because it has better test period evaluation metrics and is
 544 not subject to runoff ratio biases. The differences among the methods serve as an indicator of
 545 uncertainty which is small relative to the overall magnitude of the inflow reductions that are
 546 interpreted as being due to human consumptive use depletions. The $Pmean$ hindcast using
 547 constant precipitation underestimates dry periods and overestimates wet periods ($Q_{S,Pmean}$, red
 548 solid line, Figure 7) with a relatively high $RMSE$ (0.70 m) during the validation period, which is
 549 27% of the mean ($nRMSE$) during that same period. Because of this, we were motivated to
 550 develop a hindcast that better represents precipitation and its natural variation to use in the mass

551 balance. Taking advantage of the correlations between basin precipitation, lake precipitation, and
552 lake inflow, the *ra* hindcast precipitation we developed was estimated using a constant basin
553 runoff ratio, *r*, and constant basin to lake precipitation relationship, *a*. The performance metrics
554 are better for the *ra* inflow hindcast ($Q_{s,ra}$, blue line, Figure 7) than the *Pmean* hindcast for the
555 validation period, 1950-2023. It is apparent in Figure 7 that the variability of the *Pmean* hindcast
556 is amplified relative to the *ra* hindcast for the full period (1848-2023). However, the *ra* hindcast
557 is generally below the *Pmean* hindcast prior to 1900. Although the two hindcasts overlap after
558 1900, crossing nearly every time there is a high-low or low-high transition, the low offset of the
559 *ra* hindcast noticed before 1900 does not appear in later years and also does not appear in
560 comparison to measured streamflow. This early period coincides with the period prior to
561 significant water development and, therefore, significant depletions, when the GSL Basin was
562 just being settled. Since *r* was calculated using present day precipitation and streamflow, it
563 includes some CU effects. The *r* for the first 50 years or more (e.g., 1848-1900) may be expected
564 to be more than our calculated *r* and we surmise that this may be the cause for the *ra* hindcast
565 being less than the *Pmean* hindcast during that time. Another place this effect can be seen is
566 when comparing the mean precipitation depth for *ra* for the years 1900-2023 and 1848-1900,
567 which are 0.36 and 0.45 m. Using a larger runoff ratio prior to 1900 would result in higher flow
568 than the *ra* hindcast flow, likely closer to the *Pmean* hindcast flows. In addition, when
569 comparing these hindcasts (Figure 8a) with the primary inputs hindcast (Figure 8b) for the early
570 period, 1848-1895, the *Pmean* and primary inputs hindcast are very similar and there is no bias
571 or error included in the primary inputs hindcast due to precipitation. Thus, while the *ra* hindcast
572 is better at capturing one to two year time scale variability, we feel that the *Pmean* hindcast,
573 averaged over a longer period, better captures long-term average lake inflow.

574 When summing detailed CU to determine overall CU for a basin, like the method used
575 for existing estimates for the GSL Basin, any diversions into or out of the basin would be
576 included in the final result. The transbasin diversions that enter the GSL Basin fulfill CU within
577 the basin but do not contribute to the decrease in flow into the lake or otherwise affect the lake
578 volume. We estimated the decrease in GSL inflow based on changes in volume of the lake and
579 therefore these types of diversions are not included in our result. However, we believe that they
580 are effectively part of the CU reported by Wurtsbaugh et al. (2017) and the GSL Strike Team
581 (GSLST, 2023). Therefore, for comparison to our results it is necessary to remove these
582 diversions from the existing estimates of CU. The average annual transbasin diversion from the
583 Colorado River Basin to the GSL Basin in Utah between 2001 and 2020 was 145,700 AF or 0.18
584 km³ (Bureau of Reclamation, 2022). This is similar to the 0.16 km³/yr transbasin diversion that
585 Wurtsbaugh et al. (2017) reported and included in their water mass balance calculations.
586 Subtracting these transbasin inputs from the reported CU values we are comparing to, we get an
587 equivalent 2.42 km³/yr for GSL Strike Team and 1.54 km³/yr for Wurtsbaugh et al. (2017). This
588 is similar to the range of our results, 1.41-2.26 km³/yr, based on the range of *ra* and *Pmean*.
589 However, as noted above, our *ra* hindcast appears to be biased low in early years due to *r* having
590 been calculated with streamflow that includes CU and applying it to a time when there
591 presumably were nearly no depletions. Therefore, the estimate of 2.26 km³/yr from the *Pmean*
592 hindcast is more defensible and is very close to the GSL Strike Team results, yet it is
593 approximately 50% higher than the results from Wurtsbaugh et al. (2017). This leads us to
594 believe that Wurtsbaugh et al. (2017) may not have had as complete information or data as was
595 available for this study and that was used by the GSL Strike Team. On the other hand, the GSL
596 Strike Team included CU from areas of the basin that are hydrologically disconnected from the

597 lake, like the west desert, which does not affect the GSL WSE and makes an estimate based on
598 summing CU higher. It is also worth noting that the characteristics of the early trend in both the
599 *Pmean* and *ra* flow hindcasts mimic what Wurstbaugh et al. (2017) asserted about the majority
600 of CU occurring shortly after the GSL Basin was settled, and the characteristics of the recent
601 trend mimic the constant CU shown in the estimate by the GSL Strike Team.

602 **5 Conclusion**

603 Reductions in streamflow into the GSL are responsible for declines in the terminal lake's
604 WSE. Present estimates of consumptive water use within the GSL Basin based on summing
605 detailed uses and return flows are disparate and range from 1.5 to 2.4 km³/yr. This paper offers
606 an additional line of evidence for quantifying consumptive use by estimating inflow depletions
607 through volume reconstruction from records of WSE. We analyzed annual changes in lake
608 volume for GSL to reconstruct inflow corresponding to the full historical record of WSE and
609 estimated the magnitude of CU in the basin and associated lake level decline for the water years
610 1848-2023. Our analysis included developing hindcasts for annual evaporation and precipitation
611 depths over the lake and inflow volume to the lake, from which we reconstructed annual
612 streamflow using a water mass balance. Two methods were used to estimate precipitation, giving
613 us a range for CU of 1.41-2.26 km³/yr and the associated lake level decline of 2.82-4.59 m. Our
614 hindcast of evaporation depth over the lake is consistent with previous mass balance estimates
615 for the years 1950-2010 but was extended to include the entire 1848-2023 period. A bias may be
616 present in the *ra* hindcast of precipitation using a constant runoff ratio due to its dependency on
617 measured streamflow that includes CU. This bias would be more pronounced, and would
618 therefore result in an underestimate of CU, during years immediately after the initiation of
619 irrigated agriculture following Mormon Pioneer settlement in the GSL Basin starting in 1847.
620 Because of this bias in the *ra* based estimate, we believe that the *Pmean* estimate of CU in GSL
621 Basin of 2.26 km³/yr based on constant average precipitation is better and lake level decline due
622 to this estimate is closer to 4.59 meters. These findings are important because they quantify the
623 degree to which CU has impacted the lake and serve as a reference for efforts towards
624 conservation and other water management actions aimed at restoring the lake to levels where it
625 better supports its uses.

626 **Acknowledgments**

627 This research was funded by the Utah Water Research Laboratory and the Sant
628 Foundation Endowed Professorship to David Tarboton. The authors are grateful for this support.

629 **Open Research**

630 All scripts and data used in this paper have been published in HydroShare (Tarboton,
631 2023a,c; Tarboton and Merck, 2023; Merck and Tarboton, 2023) and added to the publicly
632 available HydroShare Collection of Great Salt Lake Data (Tarboton, 2023b).

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