

# Where in the world are vegetation patterns controlled by hillslope water dynamics?

**Shuping Li<sup>1</sup>, Dai Yamazaki<sup>2</sup>, Xudong Zhou<sup>2</sup>, Gang Zhao<sup>2</sup>**

<sup>1</sup>Department of Civil Engineering, The University of Tokyo, Tokyo, Japan

<sup>2</sup>Institute of Industrial Science, The University of Tokyo, Tokyo, Japan

Corresponding author: Shuping Li (shuping@rainbow.iis.u-tokyo.ac.jp)

### Key points:

- A catchment-based strategy is proposed to represent explicit land cover heterogeneity using discretized height bands along a hillslope.
- Landscapes shaped by hillslope water dynamics are detected in many regions of the world in diverse climate zones.
- The proposed strategy can neatly resolve land cover heterogeneity in land surface modeling with relatively high accuracy.

18   **Abstract**

19   Some recent land surface models can explicitly represent land surface process and focus more  
20   on sub-grid terrestrial features. Many studies have involved the analysis of how hillslope  
21   water dynamics determine vegetation patterns and shape ecologically and hydrologically  
22   important landscapes, such as desert riparian and waterlogged areas. However, the global  
23   locations and abundance of hillslope-dominated landscapes remain unclear. To address this  
24   knowledge gap, we propose a globally applicable method that employs high-resolution  
25   elevation, hydrography, and land cover data to neatly resolve explicit land cover  
26   heterogeneity for the mapping of hillslope-dominated landscapes. First, we aggregate pixels  
27   into unit catchments to represent topography-based hydrological units, and then vertically  
28   discretize them into height bands to approximate the hillslope profile. The dominant land  
29   cover type in each height band is determined, and the uphill land cover transition is analyzed  
30   to identify hillslope-dominated landscapes. The results indicate that hillslope-dominated  
31   landscapes are distributed extensively worldwide in diverse climate zones. Notably, some  
32   landscapes, including gallery forests in northeastern Russia and desert riparian in the Horn of  
33   Africa, are newly revealed. Furthermore, the proposed strategy enables more accurate  
34   representation of explicit land cover heterogeneity than does the simple downscaling of a  
35   rectangular grid from larger to smaller units, revealing its capability to neatly resolve land  
36   cover heterogeneity in land surface modeling with relatively high accuracy. Overall, we  
37   present the extensive global distribution of landscapes shaped by hillslope water dynamics,  
38   underscoring the importance of the explicit resolution of heterogeneity in land surface  
39   modeling.

40

## 41    **Plain language summary**

42    Local land cover distributions are influenced profoundly by various factors that are not  
43    represented fully in current land surface models. In alpine regions, changes in vegetation  
44    layers from mountain bases to tops are apparent; this phenomenon is driven largely by  
45    climatic factors, such as temperature. Interestingly, similar vegetation changes occur in  
46    relatively flat regions due to uneven water distribution on hillslopes. Hillslope water  
47    dynamics contribute to the development of unique landscapes, such as gallery forests, and  
48    substantially influence local ecological and hydrological conditions. Despite this importance,  
49    the global locations and abundance of such landscapes remain mysterious. In this study, we  
50    propose a method for the mapping of the global distribution of hillslope-dominated  
51    landscapes using high-resolution land-cover, terrain, and climate data. The results reveal that  
52    the global distribution of these landscapes, including some newly revealed landscapes such as  
53    gallery forests in northeastern Russia, is extensive. In conclusion, our study sheds light on the  
54    significant role of hillslope water dynamics in determining vegetation patterns in many parts  
55    of the world, highlighting the importance of resolving local features in land surface modeling.

56

## 57    **1 Introduction**

58        Land surface models (LSMs) are integrated with climate models as the land components  
59        for the simulation of land–atmosphere water and energy exchange. For global- or continental-  
60        scale modeling, these models generally operate on large (~20–200-km) grid units. However,  
61        LSMs have a limited ability to capture fundamental land surface processes that are  
62        heterogeneous, such as hydrological processes, which are linked closely to spatially complex  
63        factors such as topography, land cover, and soil properties. These processes occur at sub-grid  
64        scales and are not readily resolved using current models (Clark et al., 2015; Fan et al., 2019;  
65        Fisher and Koven, 2020; Wood et al., 2011).

66        At the sub-grid scale, land surface heterogeneity is profoundly differentiated by factors  
67        such as local climatic, topographic, and hydrological conditions (Fisher and Koven, 2020; Li  
68        and Sawada, 2022; Tai et al., 2020). In mountainous regions with significant topographic  
69        relief, climatic gradients tend to determine vertical vegetation zonation from the valley to  
70        hilltop (see Fig. 1a, von Humboldt, 1807; Schimper et al., 1903; Zou et al., 2023). In regions  
71        where the terrain is much flatter, which are governed by hillslope-scale water dynamics, such  
72        vegetation gradients are also observed at the sub-grid scale (see Fig. 1a, Fan et al., 2017), as  
73        gravity drives vertical and lateral surface and subsurface water flow downhill, leading to a  
74        wetter (and sometimes more saline) valley and drier hill.

75        In light of the importance of addressing the fundamental phenomenon underlying these  
76        patterns at the sub-grid scale, model developers have recently aimed to resolve the high-  
77        resolution land surface heterogeneity in LSMs (Ajami et al., 2016; Burton et al., 2019;  
78        Chaney et al., 2016; Hazenberg et al., 2015; Lawrence et al., 2019; Naudts et al., 2015; Subin  
79        et al., 2014; Swenson et al., 2019). In particular, the concept of a representative hillslope,  
80        commonly incorporated into hydrological models, has been applied to LSMs. This concept is  
81        used to aggregate hydrologically similar areas in single catchments into hydrological  
82        response units, allowing the catchment area draining into the main channel to be treated as an  
83        integral hillslope. This representation consists of multiple vertical bands with varying widths  
84        and elevations, and water is routed linearly from higher to lower bands. By treating the band  
85        as the basic modeling unit for the incorporation of representative hillslopes into LSMs, this  
86        method can efficiently resolve the explicit land surface heterogeneity and the key  
87        hydrological processes occurring at the hillslope scale (Newman et al., 2014). This method  
88        has been applied in some studies, resulting in considerable reduction of the computational  
89        cost of the LSM (Hazenberg et al., 2015; Swenson et al., 2019). Furthermore, Chaney et al.  
90        (2018) divided the hillslope into connected bands and then aggregated the hyper-resolution

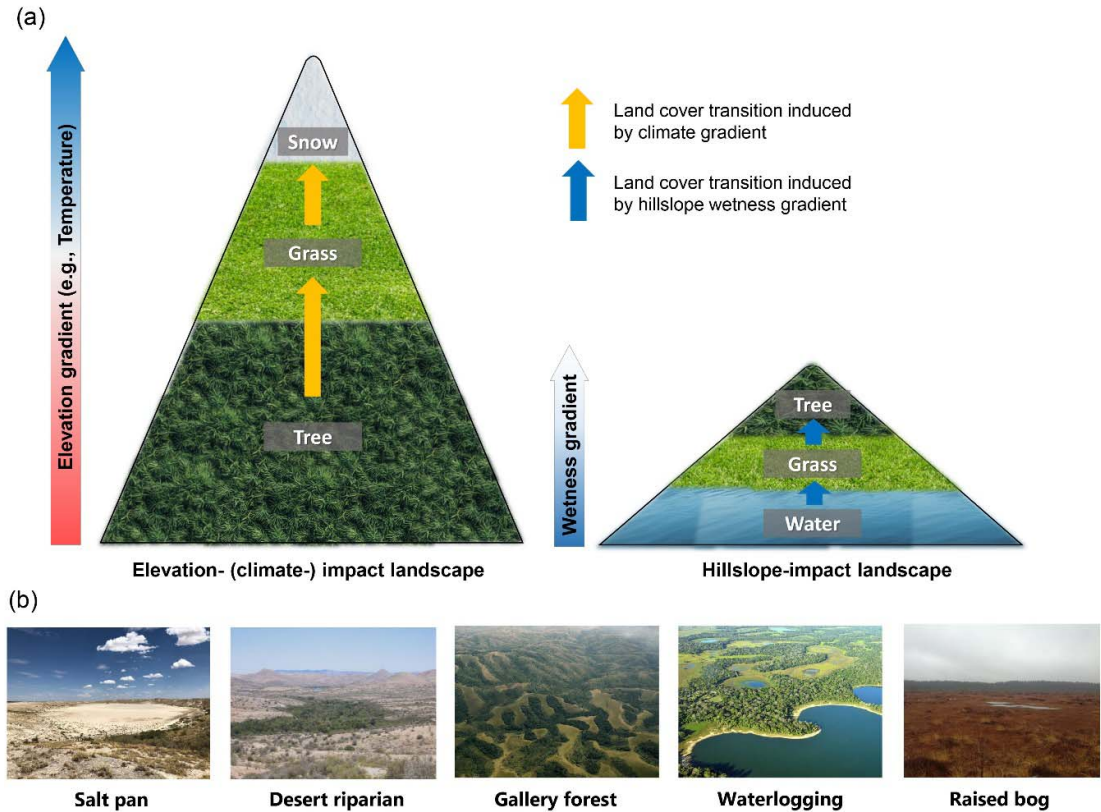


91 pixels with similar hydrological behavior in each band into complex tiles or clusters for the  
92 more efficient representation of land surface processes. This method improves computational  
93 efficiency while minimizing the degradation caused by simulation at high resolution,  
94 providing promising results for the resolution of heterogeneous land-surface processes at  
95 local and regional scales. Nevertheless, previous hillslope modeling studies have focused on  
96 model development, and relevant analysis of the global distribution of hillslope-dominated  
97 landscapes remains lacking. Research aiming to close the gap between model simulation  
98 results and the impacts on sub-grid land cover heterogeneity would provide great benefit.

99 Plant growth is strongly suppressed under extremely dry (plant water stress), humid (plant  
100 water excess), and saline conditions, leading to the development of unique landscapes (Fig.  
101 1b; detailed information is provided in Supplementary Text S1). For example, under arid or  
102 semiarid climate conditions, drier uphill areas constrain plant growth, forming desert riparian  
103 (DR) and gallery forest (GF) landscapes. When excess water accumulates in down-valley or  
104 upper hill areas, plants drown due to limited root respiration, leading to the development of  
105 waterlogging (WL) or raised bog (RB) landscapes, respectively. Salt pan (SP) landscapes  
106 develop due to excessive evaporation relative to groundwater inflow and precipitation, which  
107 causes saline soil conditions in down-valley areas and hinders plant growth. Given their key  
108 roles in influencing local land-atmosphere water and energy budgets and global  
109 biogeochemical cycles (Clark et al., 2015), these landscapes have been studied separately and  
110 regionally, with examination of their spatiotemporal distribution and evolution using state-of-  
111 the-art remote sensing techniques (Kirpotin et al., 2021; Lehner and Döll, 2004; Macfarlane  
112 et al., 2017; Nguyen et al., 2015; Safaee and Wang, 2020; Xu et al., 2018). As noted by Fan  
113 et al. (2019), such landscapes are likely to exist in diverse parts of the world. However, their  
114 distribution patterns and abundance have not been assessed to date. We still lack a global  
115 overview of hillslope-dominated landscapes, which is necessary to fully elucidate how  
116 hillslope water dynamics affect land surface heterogeneity.

117 In this study, the global tessellation of catchments and height bands are generated based  
118 on up-to-date high-resolution topographic data from the MERIT DEM and hydrographic data  
119 from MERIT Hydro (Yamazaki et al., 2017, 2019). Using those data combined with high-  
120 resolution land cover and climate classification maps, we aim to construct a global  
121 distribution map of hillslope-dominated landscapes and discuss the importance of explicitly  
122 resolving land surface heterogeneity in land surface modeling. First, we propose an  
123 aggregation method called the catchment-based strategy, which is used to derive a global  
124 distribution map of landscapes shaped by hillslope water dynamics. Second, to assess and

verify the detection results, we perform visual examination to calculate the detection accuracy. Third, to determine the abundance of hillslope-dominated landscapes, we construct a global distribution map of landscapes shaped by climate impacts for comparison. Finally, through comparison of the catchment-based strategy with the simple downscaling of a rectangular grid from larger to smaller units, we demonstrate the superiority of this new method for the resolution of land cover heterogeneity.



**Figure 1.** (a) Examples of landscapes dominated by elevation (climate) and hillslope dynamics. Temperature is considered to be the factor controlling uphill land cover transition in climate-dominated landscapes, and wetness is regarded as the major controller of hillslope-dominated landscapes. Note that we selected one pathway of land cover transition for illustration; many other paths for elevation- and hillslope-dominated landscapes exist. (b) Typical landscapes shaped by hillslope water dynamics: salt pan (SP; salt lakes of Pinos Wells; photograph by David Ryan, <https://gentleartofwandering.com/wandering-around-the-salt-lakes-of-pinos-wells/>, used with permission), desert riparian (DR; forest corridor in arid Arizona, [https://en.wikipedia.org/wiki/Desert\\_riparian](https://en.wikipedia.org/wiki/Desert_riparian)), gallery forest (GF; forest corridor in the Luama Katanga Reserve of eastern Congo; photograph by Andrew Plumpton/WCS, <https://news.mongabay.com/2014/11/mapping-mistake-leaves-wildlife-at-risk/>, used with

permission), waterlogging (WL; oxygen-stressed environment in Pantanal, <http://wikimapia.org/8582923/Pantanal-Mato-Grossense-National-Park>), and raised bog (RB; uplifted peatland in Teijo National Park, Finland, [https://en.wikipedia.org/wiki/Raised\\_bog](https://en.wikipedia.org/wiki/Raised_bog)). The images represent climatic gradients from hot and dry on the left to cold and wet on the right.

## 2 Data

The datasets used in this study are listed in Table 1.

We employed the MERIT DEM dataset as our topographic data. Major error components of other DEMs have been eliminated from this dataset through the separation of types of bias (absolute and tree height) and noise (stripe and speckle) using multiple satellite datasets and filtering techniques (Yamazaki et al., 2017). In particular, significant improvements have been achieved in flat regions with height errors exceeding their topographic variability, and landscapes such as river networks and hill–valley structures are represented clearly.

We used MERIT Hydro as our hydrographic data, representing the global hydrographic network. These data are derived from the MERIT DEM and water body datasets (G1WBM, Global Surface Water Occurrence, and OpenStreetMap). Due to the increasing availability of high-quality baseline geospatial datasets, this dataset has more spatial coverage (between 90°N and 60°S) and representation of small streams than do other datasets (Yamazaki et al., 2019).

The land use/land cover (LULC) product derived from ESA Sentinel-2 imagery was used as the land cover data (Karra et al., 2021). A global LULC map was created based on a large novel dataset of more than 5 billion human-labeled Sentinel-2 pixels, with a high resolution of 10 m. The LULC data represents 11 types of land cover: clouds, snow/ice, bare ground, built areas, scrub/shrub areas, crops, flooded vegetation, grass, trees, water, and oceans. LULC products from 2017 to 2022 are available; we used the 2020 product in our analysis. To consistently match the spatial resolution of the MERIT DEM and MERIT Hydro data, the LULC data were aggregated from 10 m to 3 arcsec (i.e., 90 m at the equator) using the nearest-neighbor interpolation method.

To account for climate impacts, we used the present-day Koppen–Geiger map as described by Beck et al. (2018). The map was generated from an ensemble of four high-resolution, topographically corrected climatic maps, and has greater classification accuracy and more detailed information than do previous versions, especially in regions with sharp spatial or elevation gradients. To maintain consistency with the MERIT DEM and MERIT

Hydro data, this map was resampled from 1-km to 3-arcsec spatial resolution.

In general, the topographic, hydrographic, land cover and climate classification data are used for landscape detection, as described in section 3. In section 4, the results obtained with the combined application of satellite imagery, topographic and land cover information are evaluated and discussed.

**Table 1.** Datasets used in this study.

Dataset	Name	Spatial resolution	Temporal range	Reference
Topography	MERIT DEM	3 arcsec	-	Yamazaki et al. (2017)
Hydrography	MERIT Hydro	3 arcsec	-	Yamazaki et al. (2019)
Land cover	LULC Sentinel-2	10 m	2017-2022	Karra et al. (2021)
Climate classification	Koppen-Geiger map	1 km	1980-2016	Beck et al. (2018)
Optical satellite image	Google static map	-	-	-

### 3 Methods

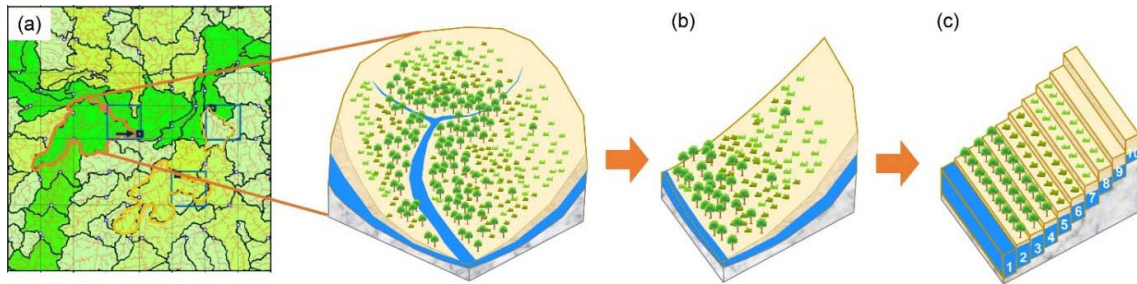
#### 3.1 Catchment-based strategy

Land surface processes are commonly modeled in LSMs based on a large rectangular grid, with the topographic factor parameterized uniformly within each grid cell (Takata et al., 2003; Wood et al., 2011). In this case, the major river channel in each grid unit is not determined explicitly, and relative height above the river channel cannot be defined. Without explicit consideration of the main river channel and hillslope drainage into the channel in each calculation unit, an LSM cannot resolve observed land cover heterogeneity that is shaped by hillslope water dynamics. Thus, in this study, we propose a catchment-based strategy for LSMs that can neatly resolve the sub-grid heterogeneity related to hillslope water dynamics, as follows.

- 1) Based on the MERIT Hydro high-resolution hydrographic dataset, the flow directions of pixels are merged to create a terrestrial boundary map of unit catchments using the flexible location of waterways (FLOW, Yamazaki et al., 2009) upscaling method. By maintaining uniform catchment size and river channel connectivity, FLOW allocates outlets throughout river networks to define the main river channels of unit catchments. Each rectangular grid unit is spatially paired with one unit catchment of similar size; thus, the catchment boundary map roughly aligns with the Cartesian grid

coordinate system commonly used in LSMs (Fig. 2a). Although the catchment boundary map can be created flexibly to discretize unit catchments into multiple sizes, we discretize unit catchments to match the 0.25° rectangular grid units used in this study.

- 2) A catchment generally consists of numerous complex hillslope forms, among which land cover transitions from channel to ridgeline are assumed to be highly similar. For conceptual clarity and computational efficiency, these complex hillslope forms are theoretically collapsed into a neat representative hillslope based on the relative height above the main river channel (Fig. 2b).
- 3) The representative hillslope is discretized into 10 vertical height bands, with uniform surface area in each band (Fig. 2c).
- 4) The proportion of each land cover type in each band is summarized, and the dominant land cover type (that accounting for the largest proportion) is identified. To efficiently represent the explicit land cover heterogeneity using height bands, we assume that the land cover of pixels within each band is represented uniformly by the dominant land cover type (Fig. 2c).



**Figure 2.** Schematic diagram of the catchment-based strategy. (a) The terrestrial area is first segmented into unit catchments of similar sizes; (b) the representative hillslope is applied as a conceptual approximation of the unit catchment; and (c) the representative hillslope is discretized vertically into 10 height bands, each with uniform surface area.

### 3.2 Search for hillslope-dominated landscapes

For the five major hillslope-dominated landscape types (Fig. 1b), information regarding the transition paths of dominant land cover types from lowland to highland was obtained from relevant studies (i.e., Fan et al., 1997, 2017, 2019; Rodríguez-González et al., 2010; MacKay, 2013; Schulz et al., 2015; Roebroek et al., 2020; Safaei and Wang, 2020; van der Velde et al., 2021) and summarized in Table 2. Some land cover types share similar

characteristics of plant adaptation to water excess or stress (e.g., in GF Path I, the grass and shrub herbaceous plant types on the upper hillslope can both withstand water stress), and in certain circumstances, several land cover types are assumed to collectively represent the dominant land cover in one height band. By summing the proportions of these LULC types within a height band, the characteristically similar land cover types are merged into one type prior to landscape detection.

**Table 2.** Summary of vertical land cover transition paths for five hillslope-dominated landscape types.

	Salt pan	Desert riparian (Arid/semiarid)	Gallery forest (Seasonally dry)	Waterlogging	Raised bog
Abbreviation	SP	DR	GF	WL	RB
Path I	Shrub ↑ Bare ground	Grass/Shrub ↑ Tree		Tree * ↑ Water+Flooded veg.	Flooded veg. ↑ Tree
Path II	-	Grass+Shrub ↑ Tree ↑ Water+Flooded veg.		Tree ↑ Grass+Shrub ↑ Water+Flooded veg.	Flooded veg. ↑ Grass+Shrub ↑ Tree
Path III	-	-		-	Flooded veg. ↑ Tree ↑ Water+Flooded veg.
Path IV	-	-		-	Flooded veg. ↑ Grass+Shrub ↑ Tree ↑ Water+Flooded veg.
PTV	40%	40%	30%	30%	30%
Reference	Fan et al. (1997) Schulz et al. (2015) Safaei and Wang (2020)	MacKay (2013) Fan et al. (2019) Roebroek et al. (2020)		Rodríguez-González et al. (2010) Fan et al. (2017, 2019)	van der Velde et al. (2021)

Note. Flooded veg. in the Sentinel-2 LULC data incorporates multiple flooded vegetation types, such as swamps and bogs For WL Path I, labeled with ‘\*’, the proportion of flooded

vegetation should be larger than 0. The optimal proportion threshold value (PTV) for the detection of each landscape type is determined in the validation step and summarized here.

The procedure for hillslope-dominated landscape detection is as follows.

- 1) For each unit catchment (representative hillslope), the proportion of each land cover type in each height band is calculated. The land cover type with the largest proportion is defined as the dominant land cover type.
- 2) Similar to step 1, the proportion of each climate type in each height band is calculated and the dominant type is identified.
- 3) Starting from the lowest band and stopping flexibly at any upper band, if the uphill transition of the dominant land cover type matches any path listed in Table 2, the unit catchment is maintained as a preliminarily detected hillslope-dominated landscape. Note that DR and GF share identical transition paths, but develop under different climatic conditions (Fan et al., 2019); therefore, these classes are first detected and then differentiated based on arid/semiarid or seasonally dry climate conditions, respectively, with reference to the Koppen–Geiger climate map.
- 4) The preliminarily detected landscapes are evaluated by setting a proportion threshold value (PTV) for the dominant land cover type in each band. Detected landscapes are invalidated and removed when the proportion of the dominant land cover type falls below the PTV. Note that this process leads to the non-detection of some landscapes if the PTV is set too high. To explore the optimal value, we tested PTVs of 20%, 30%, 40%, 50%, and 60% for the detection of each landscape type. The optimal PTV for each landscape type is discussed in section 4.2 and summarized in Table 2.
- 5) The PTV is set to examine the proportion of the dominant climate type for further evaluation of the detected landscapes in step 6. Spatially, the climate type distribution is more homogeneous than the land cover distribution, and uphill transition trends are expected to be less frequent for the former than the latter (e.g., the climate type distribution is uniform for the hillslope-dominated landscape illustrated in Fig. 1a). Thus, to minimize the bias caused by heterogeneity in the climate type distribution and the risk of false detection of hillslope-dominated landscapes, we set the PTV as high as 90% to define the dominant climate type in each height band.
- 6) The preliminarily detected landscapes are evaluated, and those satisfying any of the following conditions are excluded:
  - Following the uphill transition of the dominant land cover type shown in Table 2,

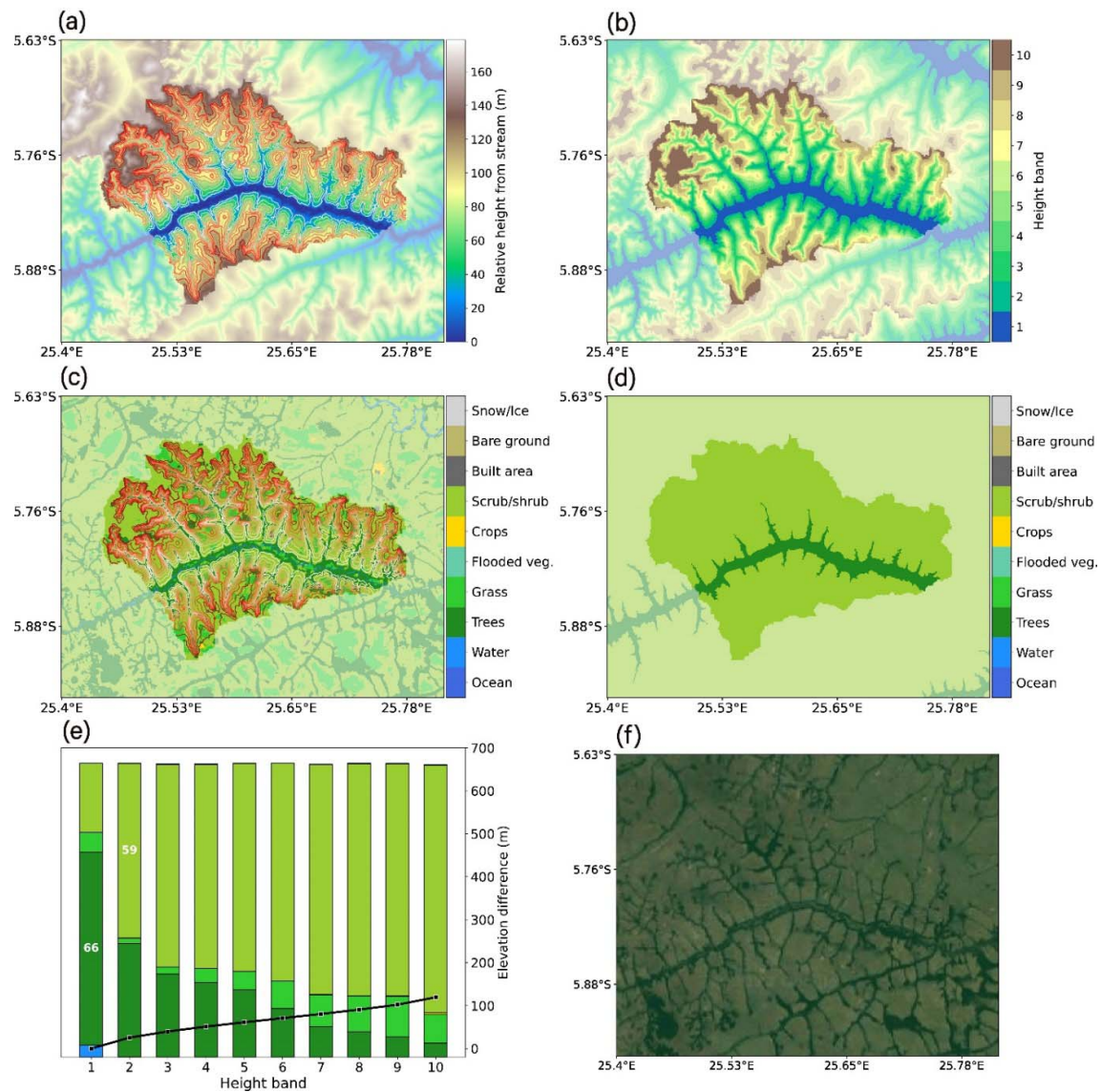
a change in the dominant climate type occurs. This rule aims to exclude climate impacts to focus on the impact of hillslope water dynamics in the catchment. Note that the hillslope dynamics may drive a land cover transition on the lower hillslope while climate drives a land cover transition on the upper hillslope; in such cases, the unit catchments are classified as both climate-dominated and hillslope-dominated landscapes. To focus on the impact of hillslope water dynamics, such mixed catchments are treated uniformly as hillslope-dominated landscapes.

- The integrated proportion of built area and cropland in any height band exceeds 1%. Human impacts are excluded to consider only the impacts of natural factors on the land cover distribution in this study. However, the land cover distribution could be altered strongly by human activities such as groundwater depletion, deforestation, and grazing. A small portion of area affected by human activities may strongly impact the surrounding land cover distribution (Fig. S2). For this reason, the proportions of built areas and cropland should be constrained to small values.

- For SPs, (a) the proportion of flooded vegetation exceeds 0 or the climate type is not arid/semiarid. Because these landscapes often appear in terminal lake basins where the climate is extremely hot and dry, the growth of aquatic plants is largely constrained due to water scarcity and saline conditions. (b) An ocean pixel is detected near or in the unit catchment. This rule is used to avoid confusion with another SP type that is distributed in coastal regions and affected mainly by seawater with or without downhill waterflow, such as SPs in tidal salt marshes (Pethick, 1974).

- 7) For the remaining hillslope-dominated landscapes, the elevation range over which hillslope impacts are detected, i.e., the relative height of the band in which the final transition occurs, is summarized. To avoid the false detection of landscapes where the vertical land cover distribution is prone to climate impacts, the threshold defining the elevation range of hillslope impacts was determined heuristically to be 100 m (Supplementary Text S2). Landscapes in which the elevation range of hillslope impacts exceeds this threshold are excluded.





**Figure 3.** Example of GF detection in Kinda-Mwampu, Congo. (a) Relative height from the mainstream, (b) height bands aggregated through the catchment-based strategy, (c) land cover map, and (d) land cover map aggregated through the application of the catchment-based strategy near the location of the detected landscape. In (a)–(d), the target catchment is highlighted in a brighter color than the surrounding area. In (a) and (c), the boundaries of height bands in the landscape are represented as contours, with redder contours enclosing height bands at higher elevations. (e) Bar plot showing the proportions of land cover types in each height band. Bars with value tags represent the height bands involved in step 3 of the detection procedure, and the values indicate the proportions of dominant land cover types in the bands. The line plot shows the median difference in elevation for each height band relative to the lowest band. (f) Static Google Earth map of the same area.

To illustrate the procedure outlined above, we provide an example of GF detection in Fig. 3. The local topography of the unit catchment is shown in Fig. 3a. Based on the topography, the catchment-based strategy is used to discretize the catchment into 10 height bands (Fig. 3b). The discretized height bands are applied to the high-resolution land cover map (Fig. 3c) and the dominant land cover type (Fig. 3d) is determined according to the summarized proportion of each land cover type in each band (Fig. 3e). The dominant land cover type changes from trees in the first band to scrub/shrubs in the second band, matching the assumed GF transition path (Path I in Table 2) with relatively high proportions of 66% and 59%, respectively. Fig. 3e illustrates the elevation range where hillslope impacts were detected, which is near 50 m and below the set 100-m elevation threshold. Using these data in combination with the satellite image (Fig. 3f), no built area or cropland is detected nearby, suggesting the development of this landscape with little anthropogenic interference. Overall, these results exemplify the successful detection of a GF that developed mainly under the impact of hillslope water dynamics. Furthermore, the land cover in each band is uniformly represented by the dominant type to generate an aggregated land cover map (Fig. 3d). Subsequently, as described in section 5.1, the accuracy of dominant land cover type representation of catchment-based strategy is determined and compared with that of a simple grid-downscaling method.

### 3.3 Validation of detected landscapes

To validate the landscapes detected as described in section 3.2, we examine the detection results generated with different PTVs (20%, 30%, 40%, 50%, and 60%) for each transition path listed in Table 2.

In general, fewer landscapes will be detected with higher threshold values; underestimation may occur if the PTV is set too high. To avoid this issue, we compared PTV categories to determine the appropriate PTV for each landscape type based on the point at which the number of detected landscapes reaches peaks and shows little further difference. To improve robustness, the highest PTV among the appropriate values is selected as the optimal threshold and employed to derive the global distribution of the corresponding landscape type.

On the other hand, due to inherent deficiencies in the detection method or baseline data, falsely detected landscapes may be included in the results, leading to overestimation for certain landscape types. To evaluate the risk of overestimation, we visually examine the

detected landscapes for each PTV category, identifying false detections and then calculating the detection accuracy ( $a_{PTV}$ ):

$$a_{PTV} = \frac{m}{n}, \quad (1)$$

where the PTV is 20%, 30%, 40%, 50%, or 60%;  $m$  denotes the number of landscapes confirmed to be correct detections by visual examination; and  $n$  denotes the number of landscapes selected for visual examination. As the visual examination of all landscapes is difficult when the detection number is large, we randomly select 10% of the detected landscapes for each landscape category as  $n$ . Specifically, we examine each selected landscape with reference to the corresponding land cover distribution and satellite image. When the spatial information provided by the reference maps is unclear, we additionally check the location at a smaller scale using Google Earth to confirm whether the landscape has been shaped by hillslope water dynamics. Generally, false detections are identified in the following cases:

- 1) Despite the effort to exclude human impacts during detection, landscapes affected by human factors may be falsely detected. For example, regularly trimmed woodland is falsely detected as GF in Fig. S3; the land cover distribution pattern has resulted mainly from human activity, rather than hillslope water dynamics.
- 2) The misclassification of land cover type in the Sentinel-2 LULC product may lead to false detection. Fig. S4 shows an example of a falsely detected landscape, with an abrupt change of land cover from trees to scrub/shrubs visible in its northern part. The scrub/shrubs present in the highest band have been mistakenly detected as the dominant land cover type due to a classification error.
- 3) Local factors such as soil type, wind, wildfire, aspect, and microclimate driven by microtopography may mediate the land cover distribution in the landscape (Aas et al., 2019); in some cases, the detection method may lead to the false detection of these landscapes as shaped by hillslope water dynamics (e.g., Fig. S5).

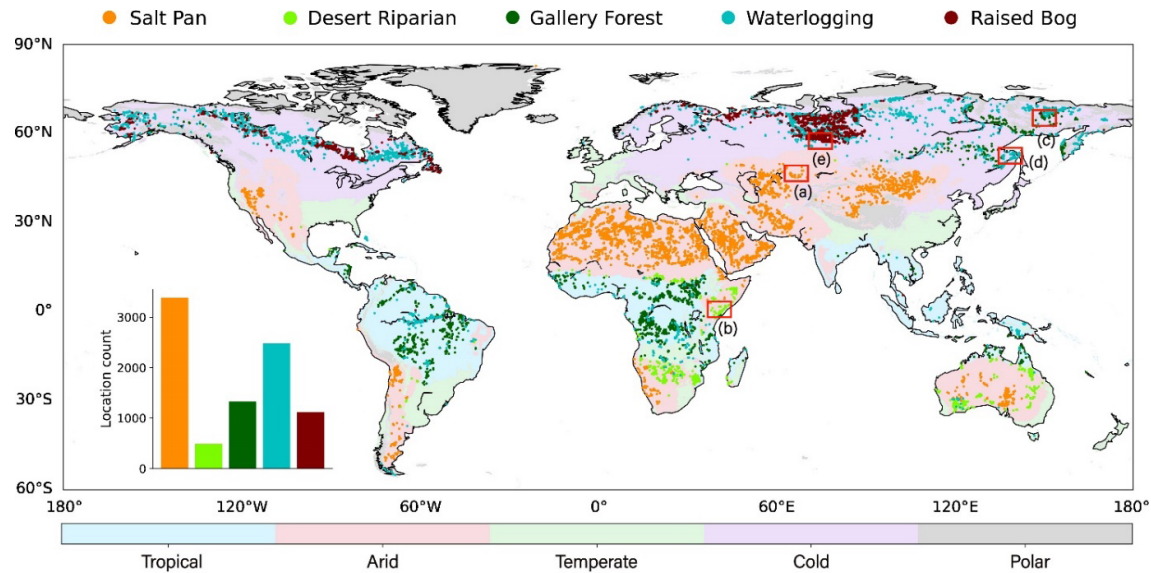
Generally, the occurrence of these issues is independent of the PTV setting. To robustly evaluate the extent of overestimation, we calculate the mean detection accuracy ( $\bar{a}$ ) among the five PTV categories:

$$\bar{a} = \frac{\sum_{i=1}^5 a_{PTV}}{5}. \quad (2)$$

376

## 377 4 Results

### 378 4.1 Global distribution of hillslope-dominated landscapes



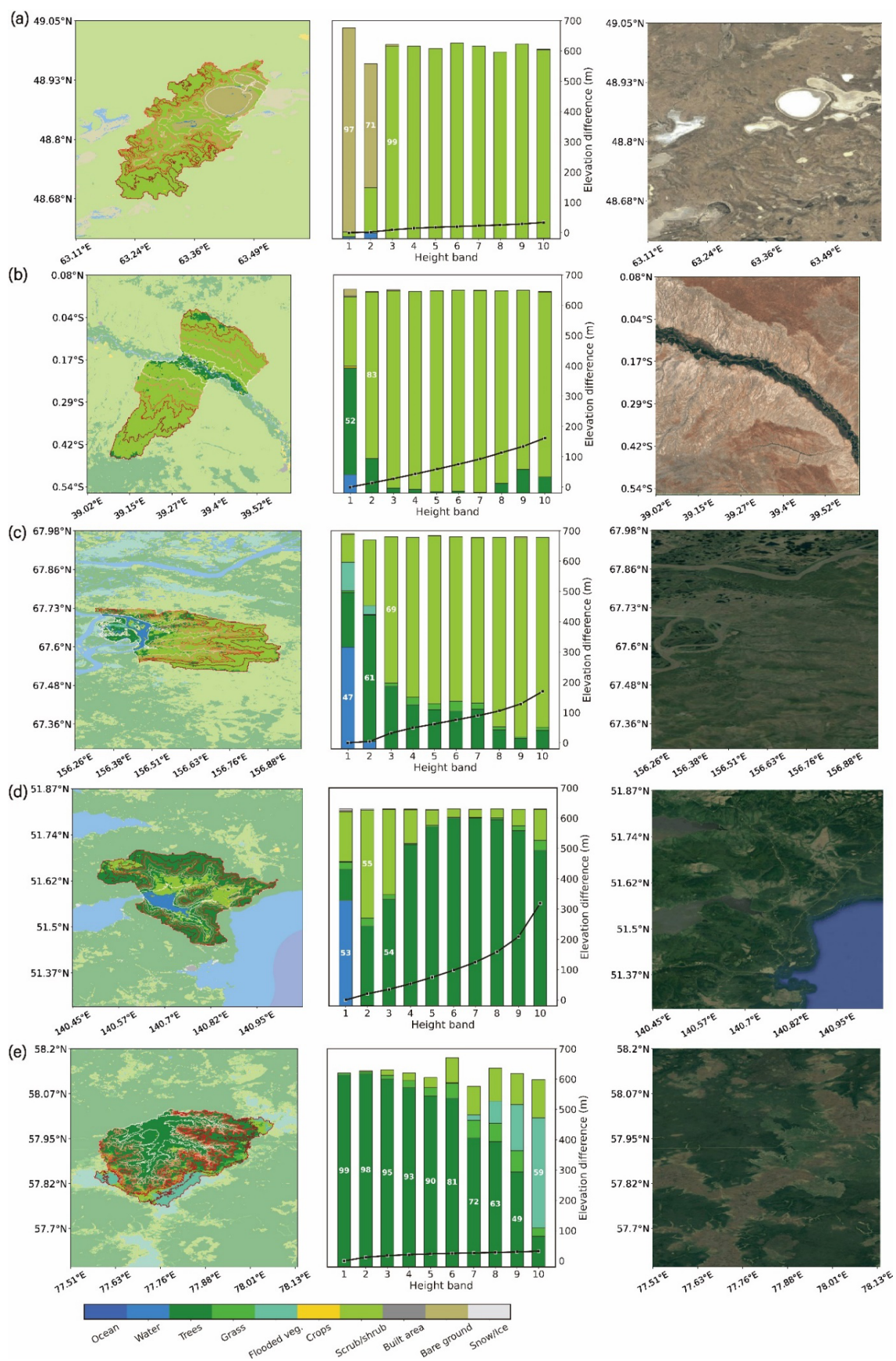
379

380 **Figure 4.** Global distribution of five hillslope-dominated landscape types (derived using the  
 381 optimal PTV, determined as described in section 4.2) and global climate classification map  
 382 (Beck et al., 2018). Locations with overlapping RB and WL are represented by RB. The inset  
 383 bar plot shows the abundance of each landscape type. Red boxes indicate the locations of the  
 384 landscape examples presented in Fig. 5.

385 Fig. 4 illustrates the global distribution of hillslope-dominated landscapes, derived  
 386 through the synthesis of the maps for each detected hillslope landscape type (Fig. 6a–e). SP  
 387 stands out as the most abundant landscape type (3,383 detections worldwide). Overall, the  
 388 distribution map shows some geographical patterns: SPs occur mainly near 30°N and 30°S,  
 389 especially in the Northern Hemisphere, and cover a wide range of dry regions including the  
 390 Sahara Desert and Arabian Peninsula. GFs are located mainly in equatorial regions with  
 391 semiarid climate conditions, such as Amazonia in South America and the Congo Basin in  
 392 Central Africa, generally near the border between tropical and dry regions. Some GFs are  
 393 also detected in subarctic regions, such as Eastern Siberia. The WL and RB distributions  
 394 overlap with some documented wetland types, such as peatlands and swamps. Unlike RBs,  
 395 which occur primarily in boreal regions (south of Hudson Bay in Canada and Tomsk Oblast  
 396 in Russia), WLs cover large areas in boreal (Alaska and Canada in North America, Nordic  
 397 countries and Russia in Eurasia) and equatorial (Amazonia in South America and the Congo

398 Basin in Central Africa) regions.





**Figure 5.** Examples of detected (a) SP (Southern Kostanay region, Kazakhstan), (b) DR (Horn of Africa), (c) GF (northeastern Russia), (d) WL (Lake Ozero Maloye Kizi, Russia) and (e) RB (eastern Tomsk Oblast, Russia) landscapes. On the land cover maps (left), the detected landscapes are highlighted with a brighter color than the surrounding area. The boundaries of height bands in the landscapes are represented as contours, with redder contours enclosing height bands at higher elevations. The bar plots in the middle show the proportions of land cover types in each height band. Bars with value tags represent the height bands used in the identification procedure, and values indicate the proportions of the dominant land cover types in the corresponding height bands. The plotted lines show the median difference in elevation between each height band and the lowest band. Static Google Earth maps (right) show satellite images of the same areas as the land cover maps.

In Fig. 5, we present zoomed-in views and the properties of example landscapes selected from Fig. 4. The SP detected on the land cover map is clearly visible in the satellite image as an expanse of salt evaporites in the northeastern part of the domain (Fig. 5a). In an example of DR, a narrow corridor of forest is observed along a winding stream (Fig. 5b). Compared to the GF example in Fig. 3 that shows an identical transition path to DR landscape in this case (i.e., from trees to grass), virtually identical vegetation patterns are revealed on the land cover maps. However, on the upper hillslope, where the land cover is dominated by scrub/shrubs, the satellite image clearly shows sparser vegetation in the DR than in the GF landscape. This difference suggests that the differentiation of DR and GF landscapes based solely on land cover maps is impossible without the consideration of additional conditions (e.g., wetness or temperature). In addition to the GF example shown in Fig. 3, a GF example corresponding to transition Path II, in which the lowest band is represented by water (a stream), is shown in Fig. 5c. An example of WL landscape corresponding to transition Path II is shown in Fig. 5d, with waterlogging at its center and a change in the dominant land cover type from water to scrub/shrubs and then to trees from low to high bands. For the RB example shown in Fig. 5e, the brown area of the satellite image represents waterlogged peatland on lifted mounds. Note that in some cases, WL landscapes are detected in lower height bands while RBs are detected in both lower and higher height bands of the unit catchment, constituting simultaneous detection (Fig. S7). This phenomenon explains the few overlapping areas in the WL and RB distributions in Fig. 4.

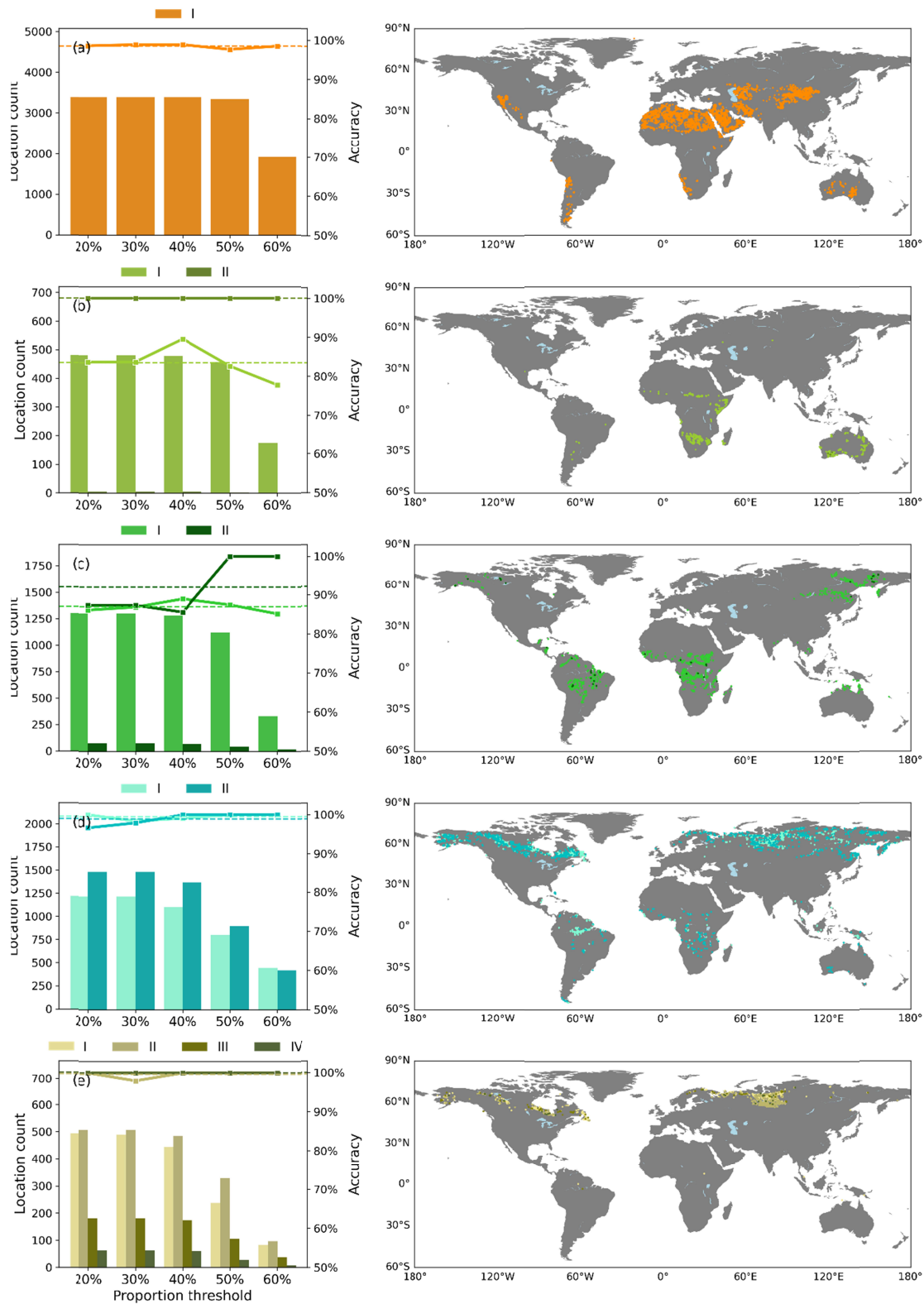
According to the lines plots in the middle panels of Fig. 5, despite differences in topography among landscapes, land cover transitions with relatively flat topography are

observed across lower height bands in areas where vegetation patterns are affected by hillslope water dynamics, e.g., the lowest three bands in the WL landscape (Fig. 5d). Notably, a hillslope impact is observed for the topographically flat RB, in which the elevation difference across all height bands is less than 20 m. This finding suggests that the force of gravity causes hillslope water dynamics to perturb or control the spatial pattern of vegetation in catchments with overall complex topography but relatively flat terrain in the lower hillslope area, in addition to flat catchments. It illustrates the widespread occurrence of hillslope impacts across numerous terrestrial regions.

In some unit catchments, hillslope and climate impacts are observed simultaneously. For some unit catchments detected as GF, the dominant land cover type first changes from trees to shrubs, and then shifts back from shrubs to trees in the highest band (Fig. S6a). The same phenomenon is observed in other landscape types, such as WL landscapes (Fig. S6b), where the dominant land cover type first changes from water and shrubs to trees, and then shifts back from trees to shrubs in the highest band. This pattern is probably due to the impacts of climatic factors, as the elevation difference between the highest and lowest bands of the unit catchment far exceeds 100 m. Thus, the land cover distribution in the same catchment is likely affected simultaneously by hillslope and climate impacts when the elevation difference is very large. To focus on the analysis of hillslope impacts, we labeled such unit catchments as hillslope-dominated landscapes and included them in the distribution results for hillslope-dominated landscapes depicted in Fig. 4.



453 4.2 Validation of the detected landscapes



454

455 **Figure 6.** Validation of the global distributions of (a) SP, (b) DR, (c) GF, (d) WL, and (e) RB

landscapes associated with five PTV categories. In the left panels, bar plots show the number of landscapes detected for each category. Solid and dashed lines denote the detection accuracy of each PTV category and the mean detection accuracy among PTV categories, respectively. The right panels show the distribution of detected landscapes generated using the optimal PTV. Detailed information about accuracy evaluation is provided in Tables S1–5.

Fig. 6 shows the validation results for the detected landscapes. The number of detections peaks for SP and DR landscapes when the PTV is set to 20%, 30% or 40% (Fig. 6a, b), and for GF, WL, and RB landscapes when the PTV is set to 20% or 30% (Fig. 6c–e). Thus, the PTV can be set higher (40%) for the detection of SP and DR landscapes than for the detection of GF, WL, and RB landscapes, indicating a more distinct pattern of transition in dominant land cover type along the hillslope for the former. Despite minor differences in detection numbers, we considered the highest PTV among all appropriate PTVs to be optimal to ensure the robustness of the results; i.e., the PTV is set to 40%, 40%/40%, 30%/30%, 30%/30%, and 30%/30%/30%/30% to derive the global distributions of SP (Path I), DR (Path I/II), GF (Path I/II), WL (Path I/II) and RB (Path I/II/III/IV) landscapes, respectively (Figs. 4, 6, 7, and S11).

The false detection of landscapes due to human factors, classification errors in the baseline data, and other factors can lead to the overestimation of the distribution of a landscape type. According to Fig. 6, mean detection accuracies for all landscape types approximate or exceed 90%, indicating a low likelihood of overestimation. An exception occurs on Path I for DR and GF landscapes, for which the mean detection accuracy is near 85%, indicating a slightly greater possibility of false detection (mainly due to human impacts; Tables S1–5) than for other landscape types. In addition, on transition Path II for GFs, greater detection accuracy is observed when the PTV is set to 50% or 60%, but this value is likely to be invalid due to uncertainty in the selected samples. Based on comparison with the landscape detection results obtained with PTVs of 20%, 30% and 40%, PTVs of 50% and 60% were not used for the construction of the distribution map due to evident underestimation.

Overall, the validation results suggest that limited overestimation occurs in landscape detection when the appropriate threshold values are applied to the proportion of dominant land cover type.

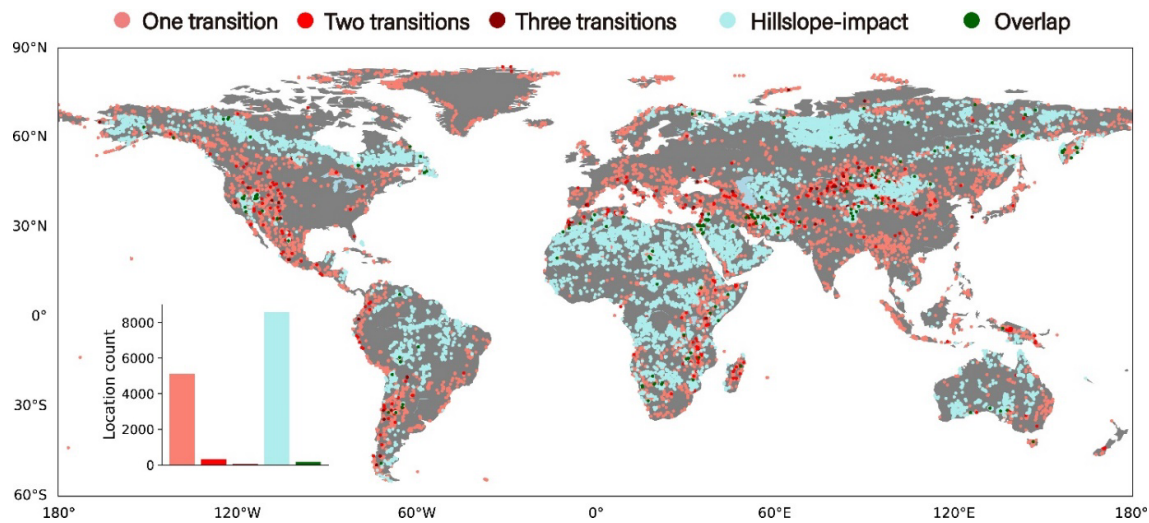
## 5 Discussion

### 5.1 Comparison of the spatial distribution of hillslope-dominated landscapes with previous research findings

The landscape map derived in this study largely agrees with distribution information provided in the relevant literature. The derived distribution of SPs shows strong spatial consistency with well-known regions of SP presence around the world (Safaei and Wang, 2020; Schulz et al., 2015), with a small fraction of mismatches in regions such as the west side of the Caspian Sea. The GF results correspond with previously reported GF distribution information, such as that for the Pantanal and Amazonia regions in central South America (Felfili, 1995; Silva et al., 2008) and West Cameroon in Africa (Momo et al., 2018). The distribution patterns derived for WL and RB landscapes are consistent with the Global Lakes and Wetlands Database (Lehner and Döll, 2004), and they overlap with the global peatland distribution map to differing extents (Kirpotin et al., 2021; Xu et al., 2018).

In addition to showing extensive overlap with documented landscape locations, the newly derived landscape map shows some landscapes that have not, to our knowledge, been previously reported. For example, we identified previously undocumented GFs in eastern Siberia (Fig. 5). Although plant growth is limited by energy (e.g., radiance and temperature) across the high-latitude regions of Eurasia (Li et al., 2021), a massive amount of dry air accumulates in the east and far east of Siberia, creating seasonal water-limited conditions (Beck et al., 2018). This regional water limitation may enhance the impact of hillslope water dynamics on vegetation patterns, leading to GF development in this region. Interestingly, in the Horn of Africa, where the climate is semiarid, a cluster of previously unreported DR landscapes was detected (Fig. 5). According to the global pattern of groundwater table depths (Fan et al., 2013), the water table is relatively shallow in this region relative to that in the surrounding area. This finding reflects the convergence of groundwater in low valleys due to hillslope water dynamics, which may contribute to DR landscape development.

## 5.2 Comparison with the distribution of climate-dominated landscapes



**Figure 7.** Global distributions of hillslope-dominated landscapes (blue) and climate-dominated landscapes with transition patterns of one (light red), two (red), and three (dark red) changes in the dominant land cover (climate) type. Areas of overlap between climate- and hillslope-dominated landscape types are represented with green dots. The inset bar plot indicates the number of each landscape type.

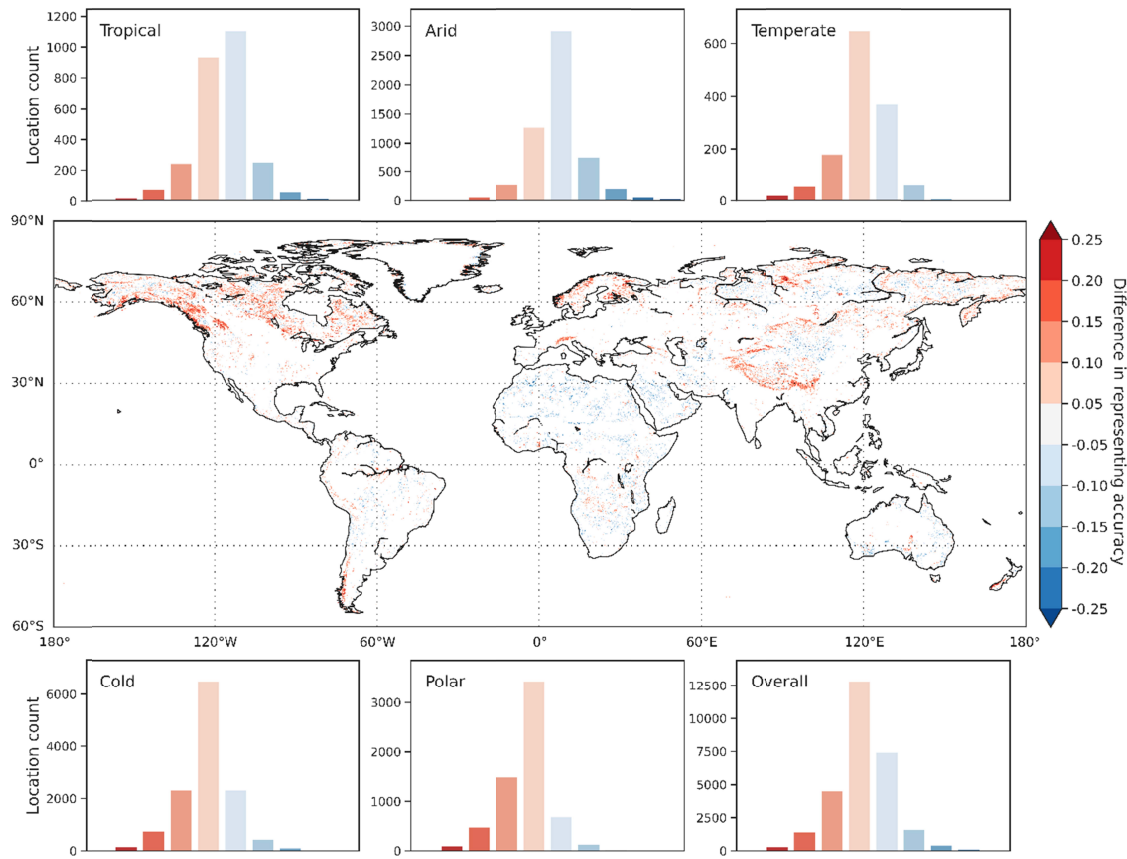
The strategy proposed here for the detection of hillslope-dominated landscapes can also be applied to the search for climate-dominated landscapes (Fig. 1a). Unlike hillslope-dominated landscapes, where vertical land cover transitions occur within the same climate zone, vertical land cover transitions in climate-dominated landscapes occur with climate zone transitions. In addition to the map of hillslope-dominated landscape distribution presented in Fig. 4, the global distribution of climate-dominated landscapes was derived (Fig. 7; the procedure is described in Supplementary Text S3). High consistency is apparent between the distributions of climate-dominated landscapes and mountainous areas globally, such as the Sierra Nevada Mountains in the western US and the Andes Mountains of South America (von Humboldt, 1807), and especially for landscapes with multiple transitions in the dominant land cover (climate) type (red and dark red dots). Fig. 7 also shows the locations where climate- and hillslope-dominated landscapes overlap (examples shown in Fig. S8), most of which are distributed near the boundary between the two landscape types.

About 5,500 climate-dominated landscapes were detected worldwide, indicating that the global coverage of these landscapes is smaller than that of hillslope-dominated landscapes (~8,500). This result suggests that hillslope water dynamics have had more extensive impacts

than climatic factors on sub-grid land cover heterogeneity, and provides further evidence of the importance of investigating these impacts. When deriving the global distributions of the five hillslope-dominated landscape types shown in Fig. 4, the application of the strict proportion threshold (1%) for anthropogenic land cover types, i.e., built area and cropland, leads to the masking out of a large number (2,240) of catchments. Nevertheless, in catchments where anthropogenic and natural land cover types coexist, hillslope impacts may perturb the vegetation pattern to some extent. For this reason, the impacts of hillslope water dynamics on global land cover heterogeneity may be broader than observed for the mapped hillslope-dominated landscapes (Fig. 4).

Overall, the broad coverage of hillslope- and climate-dominated landscapes highlights the importance of resolving sub-grid heterogeneity in LSMs for the more accurate simulation of land surface processes at small scales.

### 5.3 Representation of land cover heterogeneity in the LSM



**Figure 8.** Difference in the accuracy of land cover heterogeneity ( $\phi_{dif}$ ) representation between the catchment-based and grid-downscaling methods. Positive ( $\phi_{dif}$ ) values denote

more accurate representation with the catchment-based strategy than with the grid-downscaling method, and negative ( $\varphi_{dif}$ ) values indicate greater accuracy with the grid-downscaling method than with the catchment-based strategy. The bar plots show the numbers of locations corresponding to various levels of difference in representation accuracy for tropical, arid, temperate, cold, and polar climate types and globally. The categories represented with bars of different colors match those on the global map.

To investigate the merit of accurately resolving explicit land cover heterogeneity, the catchment-based strategy is compared with the downscaling of a rectangular grid from larger to smaller units. Specifically, we compare the ability of 10 height bands and  $3 \times 3$  rectangular grid units to approximate the explicit land cover distribution, as described in Supplementary Text S4. The explicit land cover heterogeneity is neatly resolved by assigning the dominant land cover type to the entire calculation unit (height band or grid cell). This use of the dominant land cover type to represent land cover heterogeneity inevitably leads to inaccurate representation. Thus, we determined the accuracy of representations obtained using the two strategies and the difference ( $\varphi_{dif}$ ) between them. Despite minor differences (between -0.05 and 0.05) for most locations, the  $\varphi_{dif}$  values reveal geographic patterns (Fig. 8).

The catchment-based strategy provides significantly more accurate representations of land cover heterogeneity in flat regions with humid conditions (e.g., northern Siberia and Canada, as well as river mainstems and major tributaries in Amazonia, where WL landscapes are widespread) and regions with high topographic relief (e.g., the Tibetan Plateau and Alps). In topographically flat regions where the climate is homogeneously wet, the catchment-based strategy effectively captures vertical land cover gradients shaped by hillslope water dynamics. In regions with high topographic relief where the climate is distinctly heterogeneous, the catchment-based strategy also generates more accurate representations, indicating that vegetation patterns induced by climate impacts align strongly with the topographic gradient.

On the other hand, the proposed strategy generates less accurate representations of land cover heterogeneity in flat regions with arid climates (e.g., the Sahara and Arabian Peninsula). Separately, the representation by the catchment-based strategy is favorably accurate (Fig. S9), justifying the extensive detection of SPs in the corresponding regions (Fig. 4). Despite the better infiltration conditions in flat terrain than in high-relief terrain due to longer residence of surface water (Han et al., 2020; Huang et al., 2018), the unpronounced

hillslope water dynamics impedes water convergence in lowland valleys and may have resulted in reduced accuracy. In addition, an extremely dry climate leads to substantial evaporation and thus insufficient moisture for plant uptake in the soil root zone. These two factors may collectively attenuate the impact of hillslope water dynamics on local vegetation patterns, explaining the lesser accuracy of representations obtained using the catchment-based strategy.

Overall, with the masking out of locations with trivial differences in representation accuracy, the ratios of locations labeled in red (more accurate representation with the catchment-based strategy) and blue (less accurate representation with the catchment-based strategy) to total locations are 67% and 33%, respectively. These results indicate that the proposed strategy has the advantage of resolving explicit land cover heterogeneity shaped by both climate and hillslope impacts over the simple downscaling of a rectangular grid from larger to smaller units.

#### 5.4 Limitations of the catchment-based strategy

Some landscapes reported in previous studies are absent from Fig. 4. For example, as a typical desert vegetation type in the southwestern US, DR landscapes have been frequently studied at the regional scale (Hultine et al., 2015; Nguyen et al., 2015). The impact of waterlogged conditions on plant growth in temperate regions across North and South America, Europe, and South and East Asia has been discussed intensely (Schulz et al., 2015; Zúñiga-Feest et al., 2017). Nevertheless, only a fraction of WL landscapes was detected in Scotland, UK; Tasmania, Australia; and New Zealand (Fig. 4). Missed detection of the five landscape types may be attributed to the following reasons:

1) The size of a single landscape is ambiguous, as the spatial coverage may range from a few hundred meters to several kilometers. Any pre-defined unit catchment size might be too coarse or too fine to detect landscapes that are visible on satellite imagery. When the catchment is coarsely discretized into height bands, the explicit land cover distribution may be represented inaccurately by the dominant land cover type (Fig. S10). This issue arises because the catchment-based strategy treats the lower height band as the “mainstem” and secondary tributaries as part of the hillslope. Although the satellite imagery shows that trees also line along secondary tributaries (Fig. 3f), these trees are not resolved accurately using the height bands (Fig. 3d). Hence, land cover heterogeneity remains partially resolved. With the development of a finely discretized boundary map of unit catchments and treatment of tributaries as “mainstem” areas, further improvement of the representation accuracy detection

of additional landscapes can be expected. Aside from hillslope water dynamics, other factors such as wind, wildfires, and the hillslope aspect affect local vegetation patterns in various manners (Fan et al., 2019; Gerlach, 1993; Smith and Finch, 2018). To represent this heterogeneity and thereby improve the proposed strategy, multiple tiles in each height band could be used to represent different hydrological response units (Chaney et al., 2018).

2) Landscape detection with the current catchment-based strategy begins from the lowest band, focusing on the identification of landscapes on the lower part of the hillslope. The detection procedure terminates when a change in climate type occurs. However, the climate in alpine regions could exhibit significant vertical heterogeneity within single unit catchments (Beck et al., 2018). Water dynamics may have a greater impact on the middle or upper part of the hillslope when they control vegetation patterns in those areas (von Humboldt, 1807; Zou et al., 2023).

3) The exclusion of anthropogenic factors can lead to incomplete detection results. In the southwestern US and many other places, built areas and croplands are often located near riparian areas due to their proximity to stream water. Anthropogenic land coverage is large in temperate regions due to the favorable climate conditions. Human influence explains the missed detection of a large number of landscapes in Fig. 4. This finding may reflect the substantial underestimation of hillslope impacts, as hillslope water dynamics also have great impacts on the natural distribution of land cover types in unit catchments where human impacts are less significant.

4) The limitations of baseline land cover data also hinder accurate detection. The distribution of hillslope-dominated landscapes was derived from a composite intra-annual land cover product (Table 1). Landscapes influenced by seasonal changes in land cover might be neglected in that dataset. For example, in temperate regions where precipitation has a strong seasonal pattern, seasonally flooded WL landscapes are observed widely during the rainy season. These conditions place a significant constraint on local vegetation, but are not represented in the derived map (Fan et al., 2017; Schulz et al., 2015). Aside from discrepancies between land cover types, different sub-categories of the same land cover type may differ in their adaptation to extreme water conditions. However, different sub-categories of land cover in the LULC dataset have been merged into general types (e.g., broadleaf, needle-leaf, and alpine trees are all categorized as “tree”). The lack of representation of such sub-categories affected the detection results as well.



In light of these factors, we emphasize that we did not intend to create a map that perfectly incorporates all landscape locations in this study. Rather, this study provides an overview of landscapes that are influenced by hillslope water dynamics and an unprecedented global inventory of locations with such landscapes (Fig. 4). The results underline the crucial roles of hillslope impacts in shaping various landscape types that hold hydrological and ecological significance.

## 6 Conclusion

In this study, a globally applicable catchment-based strategy is proposed to neatly resolve explicit land cover heterogeneity using discretized height bands along hillslopes. Our results show that:

- 1) Using the catchment-based strategy, we present an unprecedented global inventory of landscapes in which the vegetation pattern is shaped by hillslope water dynamics. The validated detection results for hillslope-dominated landscapes show high overall accuracy.
- 2) The detected hillslope-dominated landscapes have wide global coverage. Compared with climate factors, hillslope water dynamics affect vegetation patterns more extensively around the world.
- 3) Some landscapes, e.g., GFs in northeastern Russia and DR in the Horn of Africa, are newly revealed in this study. These findings demonstrate the strong impact of hillslope water dynamics on vegetation patterns in dry boreal and semiarid regions.
- 4) The proposed strategy more accurately resolves land cover heterogeneity than does the simple downscaling of a rectangular grid from larger to smaller units. In 67% of terrestrial areas with a distinct difference in representation accuracy, the proposed strategy provides more accurate representation of explicit land surface heterogeneity.

Some hillslope-dominated landscapes, such as DR and GFs, occur near the boundary between climate classification zones, and thus are susceptible to climate change (Fig. S11). Climate change in the coming decades could profoundly affect the status of those landscapes (Beck et al., 2018; Hagedorn et al., 2019). To investigate their spatiotemporal variation patterns from the past to the future, comprehensive elucidation of the underlying mechanism and proper inclusion in LSMs are essential. Classic LSMs provide lower boundary conditions to the atmosphere, and thus address vertical fluxes at a coarse scale and are incapable of tracing water at and near the land surface. To assess the water budget in hillslope-landscape

landscapes, an effective approach to the resolution of lateral flow must be incorporated into LSMs. The proposed catchment-based strategy should greatly aid such analysis by enabling the simulation of land surface processes in existing LSMs at sub-grid scales.

## **Acknowledgements**

We express appreciation for the valuable comments and suggestions on developing mechanism of hillslope-impact landscape by Professor Ying Fan. This research was partially supported by JSPS KAKENHI (21H05002) and by MEXT program for the advanced studies of climate change projection (SENTAN: JPMXD0722680395). S Li acknowledges funding from a PhD scholarship from Ministry of Education, Culture, Science, Sports and Technology of Japan (MEXT) and the support from the China Scholarship Council.

## **Conflict of Interest**

The authors declare no conflict of interest.

## **Data availability statement**

The MERIT DEM topography data is accessible at [http://hydro.iis.u-tokyo.ac.jp/~yamadai/MERIT\\_DEM/](http://hydro.iis.u-tokyo.ac.jp/~yamadai/MERIT_DEM/). The MERIT Hydro hydrography data is accessible at [http://hydro.iis.u-tokyo.ac.jp/~yamadai/MERIT\\_Hydro/](http://hydro.iis.u-tokyo.ac.jp/~yamadai/MERIT_Hydro/). The LULC Sentinel-2 land cover dataset is obtained from <https://livingatlas.arcgis.com/landcover/>. The Koppen-Geiger climate map is available at [www.gloh2o.org/koppen/](http://www.gloh2o.org/koppen/). All links are valid as of Sep 8th, 2023.

## **Reference**

- Aas, K. S., Martin, L., Nitzbon, J., Langer, M., Boike, J., Lee, H., et al. (2019). Thaw processes in ice-rich permafrost landscapes represented with laterally coupled tiles in a land surface model. *The Cryosphere*, 13(2), 591–609. <https://doi.org/10.5194/tc-13-591-2019>
- Ajami, H., Khan, U., Tuteja, N. K., & Sharma, A. (2016). Development of a computationally efficient semi-distributed hydrologic modeling application for soil moisture, lateral flow and runoff simulation. *Environmental Modelling & Software*, 85, 319–331.

710 Beck, H. E., Zimmermann, N. E., McVicar, T. R., Vergopolan, N., Berg, A., & Wood, E. F.  
 711 (2018). Present and future köppen-geiger climate classification maps at 1-km resolution.  
 712 *Scientific Data*, 5, 1–12. <https://doi.org/10.1038/sdata.2018.214>

713 Burton, C., Betts, R., Cardoso, M., Feldpausch, T. R., Harper, A., Jones, C. D., et al. (2019).  
 714 Representation of fire, land-use change and vegetation dynamics in the Joint UK Land  
 715 Environment Simulator v4. 9 (JULES). *Geoscientific Model Development*, 12(1), 179–  
 716 193.

717 Chaney, N. W., Metcalfe, P., & Wood, E. F. (2016). HydroBlocks: a field-scale resolving  
 718 land surface model for application over continental extents. *Hydrological Processes*,  
 719 30(20), 3543–3559. <https://doi.org/10.1002/hyp.10891>

720 Chaney, N. W., Huijgevoort, M. H. J. Van, Shevliakova, E., Malyshev, S., Milly, P. C. D.,  
 721 Gauthier, P. P. G., & Sulman, B. N. (2018). Harnessing big data to rethink land  
 722 heterogeneity in Earth system models. *Hydrology and Earth System Sciences*, 3311–  
 723 3330.

724 Clark, M. P., Fan, Y., Lawrence, D. M., Adam, J. C., Bolster, D., Gochis, D. J., et al. (2015).  
 725 Improving the representation of hydrologic processes in Earth System Models, 5929–  
 726 5956. <https://doi.org/10.1002/2015WR017096>. Received

727 Fan, Y., Li, H., & Miguez-Macho, G. (2013). Global patterns of groundwater table depth.  
 728 *Science*, 339(6122), 940–943. <https://doi.org/10.1126/science.1229881>

729 Fan, Y., Clark, M., Lawrence, D. M., Swenson, S., Band, L. E., Brantley, S. L., et al. (2019).  
 730 Hillslope Hydrology in Global Change Research and Earth System Modeling. *Water*  
 731 *Resources Research*, 1737–1772. <https://doi.org/10.1029/2018WR023903>

732 Fan, Ying, Duffy, C. J., & Oliver Jr, D. S. (1997). Density-driven groundwater flow in closed  
 733 desert basins: field investigations and numerical experiments. *Journal of Hydrology*,  
 734 196(1–4), 139–184.

735 Fan, Ying, Miguez-Macho, G., Jobbágy, E. G., Jackson, R. B., & Otero-Casal, C. (2017).  
 736 Hydrologic regulation of plant rooting depth. *Proceedings of the National Academy of*  
 737 *Sciences of the United States of America*, 114(40), 10572–10577.  
 738 <https://doi.org/10.1073/pnas.1712381114>

739 Felfili, J. M. (1995). Diversity, structure and dynamics of a gallery forest in central Brazil.

740       *Vegetatio*, 117, 1–15.

741   Fisher, R. A., & Koven, C. D. (2020). Perspectives on the Future of Land Surface Models and  
742       the Challenges of Representing Complex Terrestrial Systems. *Journal of Advances in*  
743       *Modeling Earth Systems*, 12(4). <https://doi.org/10.1029/2018MS001453>

744   Gerlach, J. (1993). Invasive melastomataceae in seychelles. *Oryx*, 27(1), 22–26.

745   Hagedorn, F., Gavazov, K., & Alexander, J. M. (2019). Above- and belowground linkages  
746       shape responses of mountain vegetation to climate change. *Science*, 1123(September),  
747       1119–1123.

748   Han, Z., Chen, X., Huang, Y., Luo, B., Xing, H., & Huang, Y. (2020). Effect of slope  
749       gradient on the subsurface water flow velocity of sand layer profile. *Journal of*  
750       *Mountain Science*, 17(3), 641–652. <https://doi.org/10.1007/s11629-019-5644-z>

751   Hazenbergh, P., Fang, Y., Broxton, P., Gochis, D., Niu, G., Pelletier, J. D., et al. (2015). A  
752       hybrid-3D hillslope hydrological model for use in Earth system models. *Water*  
753       *Resources Research*, 51(10), 8218–8239.

754   Huang, Y., Chen, X., Li, F., Zhang, J., Lei, T., Li, J., et al. (2018). Velocity of water flow  
755       along saturated loess slopes under erosion effects. *Journal of Hydrology*, 561, 304–311.  
756       <https://doi.org/https://doi.org/10.1016/j.jhydrol.2018.03.070>

757   Hultine, K. R., Bean, D. W., Dudley, T. L., & Gehring, C. A. (2015). Species Introductions  
758       and Their Cascading Impacts on Biotic Interactions in desert riparian ecosystems.  
759       *Integrative and Comparative Biology*, 55(4), 587–601.  
760       <https://doi.org/10.1093/icb/icv019>

761   von Humboldt, A. (1807). *Essay on the geography of plants (English Translation 2009)*. The  
762       University of Chicago Press.

763   Karra, K., Kontgis, C., Statman-Weil, Z., Mazzariello, J. C., Mathis, M., & Brumby, S. P.  
764       (2021). Global land use/land cover with Sentinel 2 and deep learning. In *2021 IEEE*  
765       *international geoscience and remote sensing symposium IGARSS* (pp. 4704–4707).  
766       IEEE.

767   Kirpotin, S. N., Antoshkina, O. A., Berezin, A. E., Elshehawi, S., Feurdean, A., Lapshina, E.  
768       D., et al. (2021). Great Vasyugan Mire: How the world’s largest peatland helps  
769       addressing the world’s largest problems. *Ambio*, 50(11), 2038–2049.

770 <https://doi.org/10.1007/s13280-021-01520-2>

771 Lawrence, D. M., Fisher, R. A., Koven, C. D., Oleson, K. W., Swenson, S. C., Bonan, G., et  
 772 al. (2019). The Community Land Model Version 5: Description of New Features,  
 773 Benchmarking, and Impact of Forcing Uncertainty. *Journal of Advances in Modeling*  
 774 *Earth Systems*, 11(12), 4245–4287. <https://doi.org/10.1029/2018MS001583>

775 Lehner, B., & Döll, P. (2004). Development and validation of a global database of lakes,  
 776 reservoirs and wetlands. *Journal of Hydrology*, 296(1–4), 1–22.  
 777 <https://doi.org/10.1016/j.jhydrol.2004.03.028>

778 Li, S., & Sawada, Y. (2022). Soil moisture-vegetation interaction from near-global in-situ soil  
 779 moisture measurements. *Environmental Research Letters*, 17(11), 114028.  
 780 <https://doi.org/10.1088/1748-9326/ac9c1f>

781 Li, W., Migliavacca, M., Forkel, M., Walther, S., Reichstein, M., & Orth, R. (2021).  
 782 Revisiting Global Vegetation Controls Using Multi-Layer Soil Moisture. *Geophysical*  
 783 *Research Letters*, 48(11). <https://doi.org/10.1029/2021gl092856>

784 Macfarlane, W. W., McGinty, C. M., Laub, B. G., & Gifford, S. J. (2017). High-resolution  
 785 riparian vegetation mapping to prioritize conservation and restoration in an impaired  
 786 desert river. *Restoration Ecology*, 25(3), 333–341.

787 MacKay, P. (2013). *Mojave desert wildflowers: a field guide to wildflowers, trees, and*  
 788 *shrubs of the Mojave Desert, including the Mojave National Preserve, Death Valley*  
 789 *National Park, and Joshua Tree National Park*. Rowman & Littlefield.

790 Momo, M. C., Njouonkou, A. L., Temgoua, L. F., Zangmene, R. D., Taffo, J. B. W., &  
 791 Ntoupka, M. (2018). Land-use/land cover change and anthropogenic causes around  
 792 Koupa Matapit Gallery Forest, West-Cameroon. *Journal of Geography and Geology*,  
 793 10(2), 1–56.

794 Naudts, K., Ryder, J., McGrath, M. J., Otto, J., Chen, Y., Valade, A., et al. (2015). A  
 795 vertically discretised canopy description for ORCHIDEE (SVN r2290) and the  
 796 modifications to the energy, water and carbon fluxes. *Geoscientific Model Development*,  
 797 8(7), 2035–2065.

798 Newman, A. J., Clark, M. P., Winstral, A., Marks, D., & Seyfried, M. (2014). The use of  
 799 similarity concepts to represent subgrid variability in land surface models: Case study in

800 a snowmelt-dominated watershed. *Journal of Hydrometeorology*, 15(5), 1717–1738.

801 Nguyen, U., Glenn, E. P., Nagler, P. L., & Scott, R. L. (2015). Long-term decrease in satellite  
802 vegetation indices in response to environmental variables in an iconic desert riparian  
803 ecosystem: the Upper San Pedro, Arizona, United States. *Ecohydrology*, 8(4), 610–625.

804 Pethick, J. S. (1974). The distribution of salt pans on tidal salt marshes. *Journal of*  
805 *Biogeography*, 57–62.

806 Rodríguez-González, P. M., Stella, J. C., Campelo, F., Ferreira, M. T., & Albuquerque, A.  
807 (2010). Subsidy or stress? Tree structure and growth in wetland forests along a  
808 hydrological gradient in Southern Europe. *Forest Ecology and Management*, 259(10),  
809 2015–2025. <https://doi.org/https://doi.org/10.1016/j.foreco.2010.02.012>

810 Roebroek, C. T. J., Melsen, L. A., Van Dijke, A. J. H., Fan, Y., & Teuling, A. J. (2020).  
811 Global distribution of hydrologic controls on forest growth. *Hydrology and Earth*  
812 *System Sciences*, 24(9), 4625–4639. <https://doi.org/10.5194/hess-24-4625-2020>

813 Safaee, S., & Wang, J. (2020). Towards global mapping of salt pans and salt playas using  
814 Landsat imagery : a case study of western United States Towards global mapping of salt  
815 pans and salt playas using Landsat imagery : a case study of western United States  
816 ABSTRACT. *International Journal of Remote Sensing*, 41(22), 8693–8716.  
817 <https://doi.org/10.1080/01431161.2020.1781285>

818 Schimper, A. F. W. (Andreas F. W., Fisher, W. R., Groom, P., & Balfour, I. B. (1903). *Plant-*  
819 *geography upon a physiological basis* (Rev. and e). Oxford: Clarendon Press.  
820 Retrieved from <https://www.biodiversitylibrary.org/item/33808>

821 Schulz, S., Horovitz, M., Rausch, R., Michelsen, N., Mallast, U., Köhne, M., et al. (2015).  
822 Groundwater evaporation from salt pans: Examples from the eastern Arabian Peninsula.  
823 *Journal of Hydrology*, 531, 792–801. <https://doi.org/10.1016/j.jhydrol.2015.10.048>

824 Silva, L. C. R., Sternberg, L., Haridasan, M., Hoffmann, W. A., MIRALLES-WILHELM, F.,  
825 & Franco, A. C. (2008). Expansion of gallery forests into central Brazilian savannas.  
826 *Global Change Biology*, 14(9), 2108–2118.

827 Smith, D. M., & Finch, D. M. (2018). Impacts of interacting fire, climate, and hydrologic  
828 changes on riparian forest ecosystems in the Southwest, 32–46.

829 Subin, Z. M., Milly, P. C. D., Sulman, B. N., Malyshev, S., & Shevliakova, E. (2014).

830 Resolving terrestrial ecosystem processes along a subgrid topographic gradient for an  
831 earth-system model. *Hydrology and Earth System Sciences Discussions*, 11(7), 8443–  
832 8492.

833 Swenson, S. C., Clark, M., Fan, Y., Lawrence, D. M., & Perket, J. (2019). Representing  
834 Intrahillslope Lateral Subsurface Flow in the Community Land Model. *Journal of*  
835 *Advances in Modeling Earth Systems*, 11(12), 4044–4065.  
836 <https://doi.org/10.1029/2019MS001833>

837 Tai, X., Anderegg, W. R. L., Blanken, P. D., Burns, S. P., Christensen, L., & Brooks, P. D.  
838 (2020). Hillslope hydrology influences the spatial and temporal patterns of remotely  
839 sensed ecosystem productivity. *Water Resources Research*, 0–2.  
840 <https://doi.org/10.1029/2020wr027630>

841 Takata, K., Emori, S., & Watanabe, T. (2003). MATSIRO. *Global and Planetary Change*,  
842 38(1–2), 209–222. [https://doi.org/10.1016/S0921-8181\(03\)00030-4](https://doi.org/10.1016/S0921-8181(03)00030-4)

843 van der Velde, Y., Temme, A. J. A. M., Nijp, J. J., Braakhekke, M. C., van Voorn, G. A. K.,  
844 Dekker, S. C., et al. (2021). Emerging forest–peatland bistability and resilience of  
845 European peatland carbon stores. *Proceedings of the National Academy of Sciences*,  
846 118(38), e2101742118.

847 Wood, E. F., Roundy, J. K., Troy, T. J., Van Beek, L. P. H., Bierkens, M. F. P., Blyth, E., et  
848 al. (2011). Hyperresolution global land surface modeling: Meeting a grand challenge for  
849 monitoring Earth’s terrestrial water. *Water Resources Research*, 47(5).

850 Xu, J., Morris, P. J., Liu, J., & Holden, J. (2018). PEATMAP: Refining estimates of global  
851 peatland distribution based on a meta-analysis. *Catena*, 160(September 2017), 134–140.  
852 <https://doi.org/10.1016/j.catena.2017.09.010>

853 Yamazaki, D., Oki, T., & Kanae, S. (2009). Deriving a global river network map and its sub-  
854 grid topographic characteristics from a fine-resolution flow direction map. *Hydrology*  
855 *and Earth System Sciences*, 13(11), 2241–2251. [https://doi.org/10.5194/hess-13-2241-](https://doi.org/10.5194/hess-13-2241-2009)  
856 2009

857 Yamazaki, Dai, Ikeshima, D., Tawatari, R., Yamaguchi, T., O’Loughlin, F., Neal, J. C., et al.  
858 (2017). A high-accuracy map of global terrain elevations. *Geophysical Research Letters*,  
859 44(11), 5844–5853. <https://doi.org/10.1002/2017GL072874>

Yamazaki, Dai, Ikeshima, D., Sosa, J., Bates, P. D., Allen, G. H., & Pavelsky, T. M. (2019). MERIT Hydro: A High-Resolution Global Hydrography Map Based on Latest Topography Dataset. *Water Resources Research*, 55(6), 5053–5073. <https://doi.org/10.1029/2019WR024873>

Zou, L., Tian, F., Liang, T., Eklundh, L., Tong, X., Tagesson, T., et al. (2023). Assessing the upper elevational limits of vegetation growth in global high-mountains. *Remote Sensing of Environment*, 286, 113423.

Zúñiga-Feest, A., Bustos-Salazar, A., Alves, F., Martinez, V., & Smith-Ramírez, C. (2017). Physiological and morphological responses to permanent and intermittent waterlogging in seedlings of four evergreen trees of temperate swamp forests. *Tree Physiology*, 37(6), 779–789.

#### References From the Supporting Information

Al-Khaier, F. (2003). Soil salinity detection using satellite remote sensing. ITC.

Dwivedi, R. S., Sreenivas, K., & Ramana, K. V. (1999). Inventory of salt-affected soils and waterlogged areas: a remote sensing approach. *International Journal of Remote Sensing*, 20(8), 1589–1599.

Fan, Y. (2015). Groundwater in the Earth's critical zone: Relevance to large-scale patterns and processes. *Water Resources Research*, 51(5), 3052–3069.

Fan, Y., Duffy, C. J., & Oliver Jr, D. S. (1997). Density-driven groundwater flow in closed desert basins: field investigations and numerical experiments. *Journal of Hydrology*, 196(1–4), 139–184.

Fluet-Chouinard, E., Stocker, B. D., Zhang, Z., Malhotra, A., Melton, J. R., Poulter, B., et al. (2023). Extensive global wetland loss over the past three centuries. *Nature*, 614(7947), 281–286.

Hultine, K. R., Bean, D. W., Dudley, T. L., & Gehring, C. A. (2015). Species Introductions and Their Cascading Impacts on Biotic Interactions in desert riparian ecosystems. *Integrative and Comparative Biology*, 55(4), 587–601. <https://doi.org/10.1093/icb/icv019>

Kirpotin, S. N., Antoshkina, O. A., Berezin, A. E., Elshehawi, S., Feurdean, A., Lapshina, E. D., et al. (2021). Great Vasyugan Mire: How the world's largest peatland helps



890 addressing the world's largest problems. *Ambio*, 50(11), 2038–2049.  
891 <https://doi.org/10.1007/s13280-021-01520-2>

892 Leifeld, J., & Menichetti, L. (2018). The underappreciated potential of peatlands in global  
893 climate change mitigation strategies. *Nature Communications*, 9(1), 1071.

894 MacKay, P. (2013). *Mojave desert wildflowers: a field guide to wildflowers, trees, and*  
895 *shrubs of the Mojave Desert, including the Mojave National Preserve, Death Valley*  
896 *National Park, and Joshua Tree National Park*. Rowman & Littlefield.

897 Nguyen, U., Glenn, E. P., Nagler, P. L., & Scott, R. L. (2015). Long-term decrease in satellite  
898 vegetation indices in response to environmental variables in an iconic desert riparian  
899 ecosystem: the Upper San Pedro, Arizona, United States. *Ecohydrology*, 8(4), 610–625.

900 Page, S. E., Rieley, J. O., & Banks, C. J. (2011). Global and regional importance of the  
901 tropical peatland carbon pool. *Global Change Biology*, 17(2), 798–818.

902 Roebroek, C. T. J., Melsen, L. A., Van Dijke, A. J. H., Fan, Y., & Teuling, A. J. (2020).  
903 Global distribution of hydrologic controls on forest growth. *Hydrology and Earth*  
904 *System Sciences*, 24(9), 4625–4639. <https://doi.org/10.5194/hess-24-4625-2020>

905 Scharlemann, J. P. W., Tanner, E. V. J., Hiederer, R., & Kapos, V. (2014). Global soil carbon:  
906 understanding and managing the largest terrestrial carbon pool. *Carbon Management*,  
907 5(1), 81–91.

908 Shrestha, D. P., & Farshad, A. (2009). Mapping salinity hazard : an integration application of  
909 remote sensing and modeling based techniques . (A. J. Zinck & G. Metternich, Eds.).  
910 Boca Raton : CRC Press (Taylor & Francis) .

911 van der Velde, Y., Temme, A. J. A. M., Nijp, J. J., Braakhekke, M. C., van Voorn, G. A. K.,  
912 Dekker, S. C., et al. (2021). Emerging forest–peatland bistability and resilience of  
913 European peatland carbon stores. *Proceedings of the National Academy of Sciences*,  
914 118(38), e2101742118.

915 Yu, Z., Beilman, D. W., Froking, S., MacDonald, G. M., Roulet, N. T., Camill, P., &  
916 Charman, D. J. (2011). Peatlands and their role in the global carbon cycle. *Eos*,  
917 *Transactions American Geophysical Union*, 92(12), 97–98.

918 Yu, Z. C. (2012). Northern peatland carbon stocks and dynamics: a review. *Biogeosciences*,  
919 9(10), 4071–4085.

920 Zou, J., Ziegler, A. D., Chen, D., McNicol, G., Ciais, P., Jiang, X., et al. (2022). Rewetting  
921 global wetlands effectively reduces major greenhouse gas emissions.  
922 <https://doi.org/10.1038/s41561-022-00989-0>

923

924

925 The English in this document has been checked by at least two professional editors, both  
926 native speakers of English. For a certificate, please see:

927 <http://www.textcheck.com/certificate/YB3vBT>

928