

# **Empirical Dynamic Modeling Reveals Complexity of Methane Fluxes in a Temperate Salt Marsh**

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## Abstract

Methane dynamics within salt marshes are complex because vegetation types, temperature, oscillating water levels, and changes in salinity and redox conditions influence CH<sub>4</sub> production and emission. These non-linear and complex interactions among variables affect the traditionally expected functional relationships and present challenges for interpretation and developing process-based models. We employ empirical dynamic modeling (EDM) and convergent cross mapping (CCM) as a novel approach for characterizing seasonal/multiday and diurnal CH<sub>4</sub> dynamics by identifying causal variables, lags, and interconnections among multiple biophysical variables within a temperate salt marsh using five years of eddy covariance data. EDM/CCM is a nonparametric approach capable of quantifying the coupling between variables while determining time scales where variable interactions are most relevant. We found that gross primary productivity, tidal creek dissolved oxygen, and temperature were important for seasonal/multiday dynamics ( $\rho=0.73-0.80$ ), while water level was most important for diurnal dynamics during both the growing and dormancy phenoperiods ( $\rho=0.72$  and  $0.56$ , respectively). Lags for top causal variables (gross primary productivity, tidal creek dissolved oxygen, temperature, water level) occurred between 1-5 weeks at the seasonal scale and 1-24 hours at the diurnal scale. The EDM had high prediction capabilities for intra-/inter-seasonal patterns and annual CH<sub>4</sub> sums but with limitations to represent large infrequent fluxes. Results highlight the importance of non-linearity, causal drivers, lag times, and interconnections among multiple biophysical variables that regulate CH<sub>4</sub> fluxes in tidal wetlands. This study presents a new dimension for analyzing CH<sub>4</sub> fluxes, which will prove helpful to test current paradigms in wetlands and other ecosystems.

**Keywords:** Methane flux, Saltmarsh, Nonlinear dynamics, Methane prediction, Empirical dynamic modeling, Tidal wetland

## **Plain Language Summary**

The movement of methane gas in salt marshes is complex and influenced by various factors such as plant types, temperature, water level, and changes in water salinity and oxygen levels. These factors interact in intricate ways, making it difficult to predict the outcome of methane production and movement. We applied a new method of studying methane dynamics using Empirical Dynamic Modeling (EDM) and Convergent Cross Mapping (CCM). After analyzing five years of ecosystem-scale measurements of methane fluxes, we show that the amount of oxygen in the water, the temperature, and the amount of light received by plants are crucial for understanding regulating methane fluxes within days and across seasons. The interconnections among these variables are complex, and methane fluxes may have delayed responses, which highlight the importance of these interactions. This research improves our comprehension of how environmental factors interact to affect methane fluxes in wetlands.

54    **Key points**

- 55    - Methane dynamics in salt marshes are complex and influenced by multiple variables.
- 56    - Empirical dynamic modeling (EDM) and convergent cross mapping (CCM) are novel
- 57    approaches to characterizing CH<sub>4</sub> dynamics.
- 58    - Results highlight non-linearity, causal drivers, lag times, and interconnections among multiple
- 59    biophysical variables for CH<sub>4</sub> dynamics.

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## 1. Introduction

Methane (CH<sub>4</sub>) is the second most potent greenhouse gas and a crucial atmospheric trace gas with a warming potential 25-38 times greater than carbon dioxide (CO<sub>2</sub>) (Wuebbles and Hayhoe 2002, Bridgham et al. 2013, Neubauer and Megonigal 2019). Following a largely unreconciled plateau during a stabilized phase from 2000-2006, there has been a steady global rise, with wetlands likely responsible for most natural contributions (Jackson et al. 2020). While there is evidence that this increase is the result of both anthropogenic and biogenic emissions (Stavert et al. 2022), there is no easy mitigation option for biogenic sources, and climate change could increase these emissions resulting in positive feedback for the global carbon cycle (Dean et al. 2018, Zhang et al. 2017). This is especially important for coastal wetlands where warming and sea level rise are expected to impact CH<sub>4</sub> fluxes directly through changes in microbial metabolism (Yvon-Durocher et al. 2014), and indirectly from effects on co-dependent factors such as plant productivity (GPP) and sediment redox conditions (Strom et al. 2015, Liu et al. 2019, Seyfferth et al. 2020).

Mechanisms driving CH<sub>4</sub> fluxes are complex within coastal wetlands with many potential regulators (Huertas et al. 2019, Vázquez-Lule and Vargas 2021). Water level and waterlogged conditions have been identified as a substantial control for methanogenesis (Li 2007), but these ecosystems have been considered low CH<sub>4</sub> emitters despite prevailing anoxic conditions (Conrad 2020, Bartlett et al. 1987, Wang, Zeng and Patrick 1996, Borges and Abril 2011). This paradigm relies on the idea that acetoclastic or hydrogenotrophic methanogenesis are the dominant biogeochemical pathways and prevailing sulfate-reducing bacteria outcompete methanogens (Poffenbarger, Needelman, and Megonigal 2011). However, new evidence indicates coastal wetlands could have high CH<sub>4</sub> fluxes where other biogeochemical pathways, such as methylotrophic methanogenesis, play essential roles (Al-Haj and Fulweiler 2020, Seyfferth et al. 2020, Conrad 2020). In addition, it has been recognized that multiple environmental factors (e.g., water level, temperature, GPP, salinity) regulate CH<sub>4</sub> fluxes at different temporal scales with potential nonlinear interactions (Huertas et al. 2019, Reid et al. 2013, Li et al. 2018). Consequently, it is

imperative that we accurately understand how biophysical controls regulate CH<sub>4</sub> fluxes in coastal wetlands to better understand ecosystem response to environmental change.

Our past knowledge of CH<sub>4</sub> dynamics in tidal wetlands has vastly relied on chamber-based manual measurements. This information has influenced the development of most functional relationships (e.g., salinity thresholds or temperature dependency) and, subsequently their incorporation into models used to predict CH<sub>4</sub> fluxes. However, chambers often face logistical challenges and are sporadic in time and space, restricting our understanding by delivering limited or incomplete information (Hill and Vargas 2022, Kim 2007, Yang et al. 2021). Technological advances have allowed us to collect higher frequency CH<sub>4</sub> fluxes at the ecosystem scale using the eddy covariance technique (EC) (Morin 2019, Baldocchi 2003). This data can be leveraged to explore relationships in greater detail, providing a wealth of information regarding the dynamics of driving mechanisms. Previous studies have analyzed how biophysical variables regulate ecosystem-scale CH<sub>4</sub> fluxes in salt marshes (Vázquez-Lule and Vargas 2021, Hill and Vargas 2022, Huertas et al. 2019, Li et al. 2018, Reid et al. 2013). Other studies outside salt marshes have used machine learning techniques (Rey-Sanchez et al. 2018, Zaki and Abdul-Aziz 2022) or time series analysis and information theory to identify the dominant controls of ecosystem-scale CH<sub>4</sub> fluxes (Knox et al. 2021, Sturtevant et al. 2016). Together, these studies provide insights into the complex mechanisms and hint that nonlinear dynamics may be more relevant for explaining ecosystem-scale CH<sub>4</sub> fluxes. Therefore, there is a need to identify potential lags, interactions, and interconnections to provide insights about causality and biogeochemical mechanisms controlling CH<sub>4</sub> fluxes across wetland types.

The complexity of CH<sub>4</sub> dynamics in coastal wetlands results from interacting variables that cannot be readily isolated to identify independent functional relationships (Morin 2019; Vazquez-Lule and Vargas 2021). This complexity also brings an opportunity to test alternative methods for data analysis. As a novel approach, we employed a form of nonlinear state space reconstruction referred to as empirical dynamic modeling (EDM; Sugihara and May 1990, Sugihara et al. 2012). We propose EDM as an alternative to techniques such as Granger Causality, wavelets, and information theory to unravel

complex non-linear CH<sub>4</sub> dynamics using the wealth of information from continuous time series generated from EC data (Schafer, Tripathee et al. 2014, Sturtevant, Ruddell et al. 2016, Li, Dai et al. 2018, Knox, Bansal et al. 2021). This method falls within nonlinear dynamical systems but differs from other widely used prediction-based methods, such as Granger Causality, which holds the criteria of separability (Granger 1969) and may be less effective in identifying weak-moderate couplings (Guo et al. 2022). The premise of EDM relies upon a simplex projection algorithm, the nearest neighbor forecasting method, that tracks the evolution of nearby points within a lagged coordinate state space reconstruction or embedding (Hsieh et al. 2005). In other words, simplex projection attempts to predict future values based on when similar patterns were observed in the past, with more similar past patterns assigned a higher weight for calculating the mean location of the predicted point (Petchey, 2016). In this study, we implement a form of EDM known as convergent cross mapping (CCM) as a novel approach and alternative technique to disentangling CH<sub>4</sub> dynamics with EC data within a temperate coastal salt marsh (Munch, Rogers, and Sugihara 2022). We chose the CCM approach because it can describe the complex interrelations in a dynamic ecological system where dependencies cannot be evaluated independently.

Here we examine five years of ecosystem-scale CH<sub>4</sub> fluxes using the EC technique and CCM to study seasonal and diurnal CH<sub>4</sub> fluxes (i.e., dependent variable) by quantifying the coupling between independent variables (i.e., predictors) while considering time lags and interconnections. Briefly, CCM uses state space reconstruction methods that allow us to analyze how variables interact and change with the flow of time by examining past variable states to determine if similar dynamics are embedded in the presently observed target variable (i.e., CH<sub>4</sub>). This method is appropriate for informing causality within complex nonlinear systems where lags or inconsistent relationships commonly manifest (McGowan et al. 2017), and works within the time domain, which is fundamentally different from wavelet analysis which focuses on the frequency domain (Vargas et al. 2010). CCM can also identify complex interactions where the indirect effect of one variable is relevant via a second variable that directly influences the dependent variable. Consequently, CCM can provide interpretable empirical models that can be used to develop predictions for gap filling of data or forecasting applications. Because CH<sub>4</sub> dynamics exhibit complex

patterns generated by multiple interactions, we propose that CCM is best suited to identify the nonlinearities and potential causal relationships that regulate CH<sub>4</sub> dynamics in wetlands and other terrestrial ecosystems.

## **2. Materials and Methods**

### **2.1 Study Site**

This study was performed at the St. Jones Reserve as part of the Delaware National Estuarine Research Reserve System (DNERR). The site is representative of a mid-Atlantic salt marsh with a mean elevation of  $0.60 \pm 0.26$  m relative to the NAVD88 datum (McKenna et al., 2018), yet is still influenced by semi-diurnal tidal activity and site hydrology (i.e., riverine and groundwater flows). Vegetation is dominated by a monoculture of short-form *S. alterniflora* (~66%). The remaining cover (~33%) is associated with tall *S. alterniflora*, *S. cynosuroides*, and *P. australis*, which flank tidal creeks, and *S. patens* and *P. australis* along the upland terminus (Vázquez-Lule and Vargas 2021). Previous studies conducted in this wetland suggest the presence of methylotrophic methanogenesis responsible for very high CH<sub>4</sub> concentrations ( $>200 \mu\text{M}$ ) within the sediments (Seyferth et al. 2020) but with low CH<sub>4</sub> sediment-atmosphere emissions (Capooci and Vargas 2022). These findings challenge the current paradigm that low methanogenesis is expected in tidal wetlands and consequently new approaches are needed to identify the complexity of the underlying biophysical drivers.

### **2.2 Data Acquisition**

#### **2.2.1 Eddy Covariance Measurements**

The eddy covariance (EC) technique was used to measure the ecosystem-scale net exchange of CH<sub>4</sub> with an open path near-infrared gas analyzer (Li-7700, Licor, Lincoln, NE, USA) and ecosystem-scale net exchange of CO<sub>2</sub> (NEE) with an enclosed path infrared gas analyzer (Li-7200, Licor, Lincoln, NE, USA), and wind components with a 3D sonic anemometer (Gill Windmaster Pro, Gill Instruments,



Lymington, UK) recording measurements at 10 Hz. Preprocessing was completed in Eddy Pro (version 7.0.6), which consisted of time lag compensations, double coordinate rotation of wind components, and Reynolds block averaging to calculate 30-minute fluxes of CH<sub>4</sub> and CO<sub>2</sub>. Expanded preprocessing and tower set-up descriptions can be found elsewhere (Vázquez-Lule and Vargas 2021, Hill and Vargas 2022). During post-processing, several standardized Ameriflux QA/QC procedures were applied, including removing values flagged for low quality during pre-processing, calculating nighttime storage fluxes determined by Eddy Pro, and range filtering of CO<sub>2</sub> to remove outliers falling beyond +/- 50 umol m<sup>-2</sup> s<sup>-1</sup>. A range filter was not applied to CH<sub>4</sub> data as spikes since ebullition could occur, and the goal was to incorporate all CH<sub>4</sub> dynamics within the 30-minute block averaging period. In addition, we used an optimal friction velocity ( $u^*$ ) threshold of 0.069 m s<sup>-1</sup> to remove low turbulence conditions and applied a fetch/footprint filter to exclude fluxes originating from forested regions beyond the marsh terminus (Vázquez-Lule and Vargas 2021).

Data gaps occurring in NEE were filled using marginal distribution sampling (MDS) with customized site variables consisting of energy fluxes (sensible and latent heat; H and LE), radiation, air temperature (T<sub>air</sub>), soil temperature (T<sub>soil</sub>), relative humidity (RH) and vapor pressure deficit (VPD) (Vázquez-Lule and Vargas 2021). Partitioning of NEE into component fluxes of gross primary productivity (GPP) was completed with the REddyProc R package (Reichstein et al., 2005; Wutzler et al., 2018) based on the standard nighttime method (Reichstein et al., 2005). To fill gaps in CH<sub>4</sub> fluxes, we applied a random forest technique with the Caret R package and used a full suite of relevant site variables (Fig. S2) (Kim et al., 2020; Kuhn et al. 2016). This method was selected because it is less biased when identifying predictors of CH<sub>4</sub> flux and because CCM requires continuous gap-filled data (Kim et al., 2020; Chang, Ushio, and Hsieh 2017).

### **2.2.2 Meteorology and Water Quality Measurements**

Meteorological variables included air temperature (T<sub>air</sub>) and relative humidity (RH, HC2-S3, Campbell Scientific, Logan, UT), precipitation (Precip) (TE 525, Tipping Bucket Rain Gauge, Campbell

Scientific, Logan, UT), photosynthetically active radiation (PAR) (SQ-110, quantum sensor, Apogee, Logan, UT), air pressure (Patm) (CS-106, Vaisala, Vantaa, Finland) and wind speed (WSpd) (05103-L Wind Monitor, Campbell Scientific, Logan, UT). VPD was calculated using Tair and RH based on the Tetens formula (Murray 1967). Water quality data were collected from the tidal creek adjacent to the EC tower using a YSI EXO2 sonde outfitted with EXO sensors. Measured variables included water temperature (Twater), water level (Level), salinity (Sal), and dissolved oxygen (DO). Meteorological and water quality data were averaged for 30 minutes to align with flux data. Minor gaps in meteorological data were filled using available on-site sensors. Gaps in water quality data, which mainly occurred during the dormancy phenoperiod, were filled using empirical relationships from another YSI EXO2 sonde located approximately 2.0 km upstream within the St. Jones River, which shares connectivity to the tidal channel. Meteorological and water quality data were collected under the National Estuarine Research Reserve (NERR) Centralized Data Management Protocol (Kennish, 2019) and can be accessed from the NERR System Centralized Data Management Office (NERR CDMO; station: delsjmet-p).

### **2.3 Data Selection and Phenology**

For examining seasonal CH<sub>4</sub> dynamics, we selected five years spanning 2017-2021 to calculate daily means of non-gap-filled observations. This period was chosen as all years contained minimal continuous gaps (<30 days). To examine diurnal dynamics, we used data only from 2020-21, the period with the most complete record (i.e., containing 15% of data gaps) and when no substantial water surges or storm events occurred. To delineate these periods, we determined phenology dates with the Phenopix R package using the greenness chromatic coordinate (GCC) from a site phenocam (Filippa et al. 2016, Hill, Vázquez-Lule and Vargas 2021). Season start and end dates were calculated based on the upturn and recession dates of the annual GCC curve from daily midday images (Gu et al., 2009). To avoid an artifact of calendar years that would result in discontinuous time series for dormancy data, we included the tail end of 2020 and excluded the tail end of 2021.

## 2.4 Convergent Cross Mapping

While previous studies have described in detail the methodology behind CCM (Ushio and Kawatsu 2020, Sugihara and May 1990, Sugihara et al. 2012, Chang, Ushio and Hsieh 2017, Tsonis et al. 2018, Wang et al. 2018), we provide a brief overview of the implementation of CCM and how we interpreted results as they pertain to the goals of the current study (Fig. S1). All CCM analysis was completed within the rEDM R package (Ye et al. 2016).

To evaluate the causality between variables, we applied CCM between CH<sub>4</sub> and all independent variables via the function *CCM* (Ye et al. 2016). The CCM function is essentially a wrapper for the simplex algorithm but accesses the level improvement in nearest neighbor predictions within the state space as the data sample size increases. Specifically, this tests if lags of a predictor variable can be used to predict a target variable (e.g., does GPP have a causal relationship with CH<sub>4</sub>) (Tsonis et al. 2018, Schiecke et al. 2015, Sugihara et al. 2012). It is a standard procedure also to test the opposite cross mapping (i.e., does CH<sub>4</sub> have a causal relationship with GPP). In a dynamic system, information about the causal variable becomes embedded within the target variable (i.e., information about past causal variables is observed in the present target variable). CCM is carried out using successively larger data samples or libraries. Predictive skill is evaluated with the output parameter  $\rho$  (the overall skill of the cross-mapping) and is expected to increase with increasing library (data) size (i.e., convergence) subsampled randomly from the main dataset as the first criterion for causation (Chang et al. 2017). The predictive skill at maximum library size (the complete data set) is further compared to results generated from a surrogate dataset containing a preserved seasonal phase amplitude with the original data randomized in time. Cross mappings need a higher final predictive skill than the final surrogate predictive skill as a second criterion for causation via this significance test (Tsonis et al. 2018). Aside from verifying convergence with increasing data libraries, all figures in this study present results from the full data library size for both surrogate and real data, representing 5 complete years of data.

Two steps are applied before CCM, which provide information about the time series of interest (CH<sub>4</sub>) (Li et al. 2021). The initial step is determining the embedding dimension ( $e$ ) via the function *EmbedDimension*, which iteratively takes successively longer lags of a set time period defined by the parameter Tau ( $T$ ). For this study, we set  $T$  to the default value of  $T=1$  to represent one day for seasonal analyses and one hour for diurnal analyses. The lagged time series are then projected into an  $e$ -dimensional state space, forming a shadow manifold or state space projection (Ye et al. 2016) (Fig. S3). The resulting manifolds are then projected back onto a coordinate axis with the best embedding dimension resulting in a nearly 1:1 mapping of the original time series (Fig. S4). The number of dimensions ( $e$ ) required to recreate the original time series is both a metric of complexity and an a priori requirement for CCM. The second step is confirming nonlinear dynamics by constructing sequential locally weighted global linear maps (S-map). S-maps are like nearest neighbor predictions from the simplex algorithm, but instead of considering the average localization of the nearest neighbors, all neighboring points are regarded with more weight given to closer neighbors through an exponential localization function (Sugihara et al. 1994). This is done via the function *PredictNonlinear*, which determines if predictions made by S-maps depend on the local state of the variable being predicted as defined by the parameter theta ( $\Theta$ ) in a locally weighted linear regression function. If predictive skill increases at  $\Theta > 0$ , we assume that predictions are highly dependent on ecosystem state space (i.e., variable interactions change with time or under specific conditions) and the hallmark of a nonlinear dynamical system (Chang et al. 2017).

We applied year-round daily data for seasonal dynamics with a minimum library size of  $n=15$  days and a maximum of  $n=1,800$  days. For diurnal dynamics, hourly data was used from the selected growing and dormancy phenoperiods using a minimum library size of  $n=15$  hours and a maximum of  $n=2,900$  and  $n=2,700$  hours, respectively. CCM was carried out using 100 sample runs with the default time period for lags set to 1 unit ( $T_p=1$  day or hour). The execution radius, which ignores nearest neighbors within the state space projection, was set at one day or 24 hours to help eliminate influence from temporal autocorrelation. To generate surrogate datasets, the original data was randomized in time

with the seasonal phase amplitude preserved for daily data by setting the method to ‘seasonal’. For hourly data, the method was set to ‘random’ (Deyle et al. 2016).

## **2.5 Model Interpretation with Extended CCM and Causal Network Maps**

Conventional CCM provides a robust means of identifying causality between variable pairs, yet it fails to capture the optimal time period of the causal influence. Extended CCM is a computationally intensive technique that involves applying CCM iteratively across multiple time period lags to determine time periods where cross-mapping skills are maximized. The procedure is also used to help resolve issues of generalized synchrony where cross-mapping is significant in both directions (i.e., Tair causality on CH<sub>4</sub> and CH<sub>4</sub> causality on Tair) (Example: Fig. S7 panel C) as the most probable causal relationship will generally have a maximum predictive skill which peaks within the negative time domain (i.e., negative lags) (Fig. S8) (Ye et al. 2015b). Causal influence can occur synchronously, over a day, several hours, or an extended period, several consecutive or separate days (Sun et al. 2021). Data were binned into 7-day periods to examine seasonal dynamics, and iterations were made spanning +/- 91 days at weekly intervals. For diurnal dynamics, hourly data was used directly, and iterations were made spanning +/- 24 hours.

By considering both the optimum identified time lags and the overall strength of the causal relationship, network maps can be constructed to assess how the full suite of variables interact to influence CH<sub>4</sub> fluxes. Conceptually, this is achieved by summarizing the strength of coupling and optimal lag times between all tested variable pairs, including all ancillary variables with only positively identified (significant) connections mapped. In general, direct connectivity is exhibited when the predictive skill is high and lag times are short and indirect connectivity is exhibited when the prediction skill is lower and lag times are long (Fan et al. 2020). We selected ranking categories of high, moderate, and low based on rho values of 0.60-1.0, 0.30-0.59, and 0-0.29, respectively. This technique allows variables with indirect links that may appear to influence CH<sub>4</sub> fluxes to be sorted from variables with highly confident direct causal links, yet it remains difficult to distinguish consistent directionality (i.e., sign positive or negative) of the connection as relationships can vary in time (i.e., state dependence) (Deyle et al., 2016).

## 2.6 Multivariate Predictive Modeling

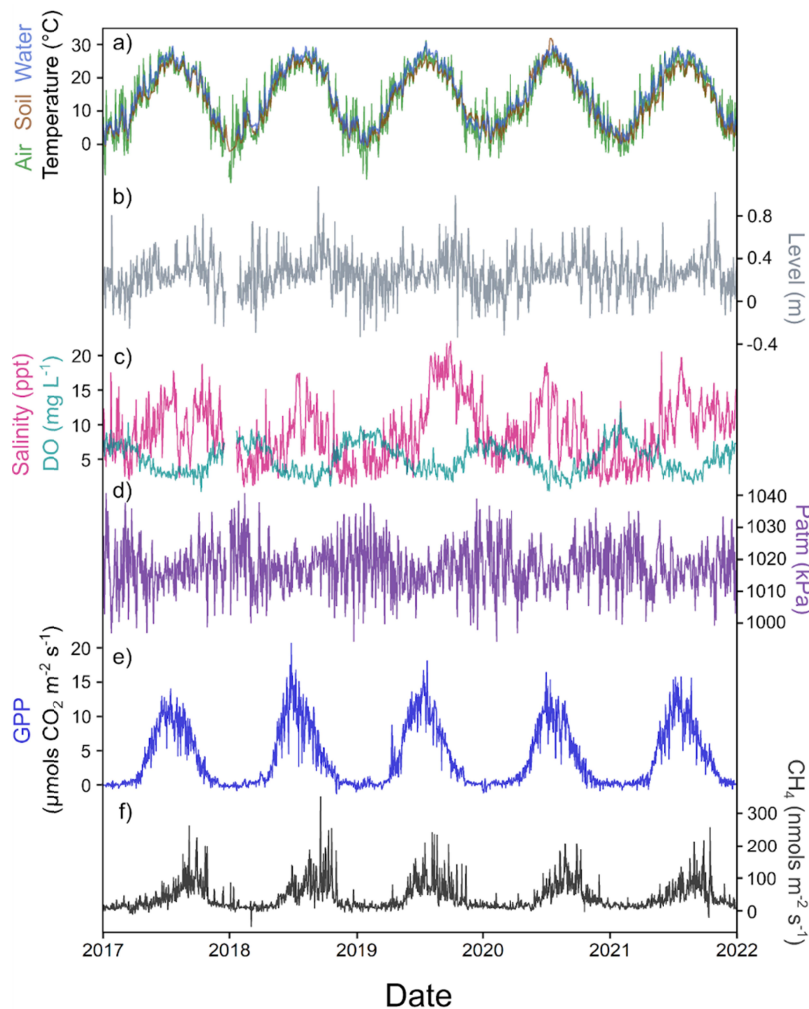
As an additional feature and promising application of EDM, we also made model predictions using a newly constructed multivariate EDM model. We used an out-of-sample validation method with 2017-2019 as the training interval and 2020-2021 as the prediction interval. Only data from the training period was used to make predictions. Predictor data was scaled to reduce dimensional distortion in the state space. The provided prediction variance based on nearest neighbors in the state space was used to calculate a standard deviation as a metric of uncertainty. We included all variables identified as coupled to CH<sub>4</sub> fluxes by daily CCM analysis: temperature components of soil, water, and air, DO, GPP, H<sub>2</sub>O flux, Level, LE, WSpd, PAR, Patm, RH, and Precip. While issues of cross-correlation would potentially violate assumptions for linear models, each of these variables causally impacts CH<sub>4</sub> fluxes in different ways as determined by information embedded in the CH<sub>4</sub> time series. Thus, the inclusion of variables that exhibit cross-correlation is possible with EDM.

## 3. Results

### 3.1 Site Characteristics, Phenology, and CH<sub>4</sub> Fluxes

During the study period (2017-2021), the annual average air temperature (T<sub>air</sub>), salinity (Sal), and water level (Level) were  $13.9 \pm 9.1$  °C,  $8.9 \pm 4.1$  ppt and  $0.25 \pm 0.2$  m above sea level, respectively with a tidal range of 1.2 m and maximum daily salinity as high as 19.5 ppt. The average annual precipitation (Precip) ranged from 567-779 mm, which is received evenly across the year (Fig. 1) (Tables S1-S3). The growing phenoperiod typically begins with green-up events occurring in April or early May, and dormancy commences with complete vegetation senescence in November. For the reference year utilizing hourly data (2020-21), the dormancy phenoperiod began on October 25<sup>th</sup>. It ended at the start of the 2021 growing phenoperiod on April 26<sup>th</sup>, which continued until the subsequent dormancy phenoperiod on October 23<sup>rd</sup>. Despite high levels of GPP, which ranged from 1,412-1,609 g C m<sup>-2</sup> yr<sup>-1</sup>, the site was a net

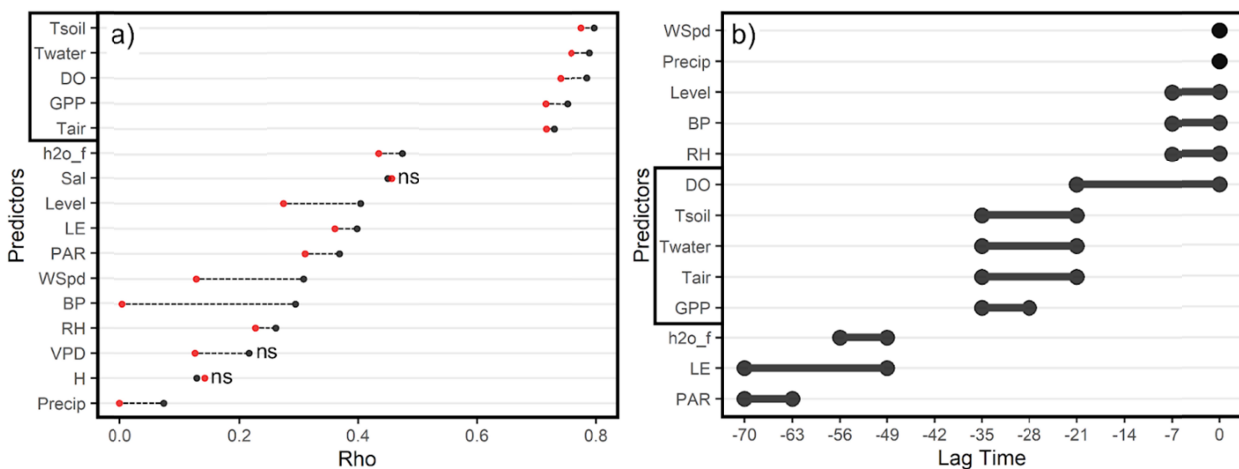
carbon source for four out of the five years (NEE range:  $-126$  to  $221 \text{ g C m}^{-2} \text{ yr}^{-1}$ ) and was a constant annual source of  $\text{CH}_4$  with emissions ranging from  $15.4$ - $16.5 \text{ g C m}^{-2} \text{ yr}^{-1}$  (Fig. 1).



**Fig. 1:** Time series of mean daily data for top identified predictive drivers and  $\text{CH}_4$  flux from 2017-21. a) Air, soil, and water temperature (green, brown, and blue, respectively), b) level (gray), c and d) salinity (pink) and dissolved oxygen (DO) (teal), d) atmospheric pressure (Patm) (purple), e) gross primary productivity (GPP) (blue), and f)  $\text{CH}_4$  flux (black).

### 3.2 CH<sub>4</sub> Time Series Characteristics

We determined that CH<sub>4</sub> is a highly dimensional independent variable that required at least ten dimensions to recreate the original dynamics (Fig. S5). Despite a slightly lower peak at four dimensions, we selected the dimensionality with the greatest predictive skill to capture a full range of potential variable dynamics; because CCM is expected to identify coupling and quantify the overall strength of all possible relationships. Consequently, we confirmed that the response of CH<sub>4</sub> flux was highly dependent on state space or transient ecosystem conditions (i.e., dynamical and nonlinear) (Fig. S6).



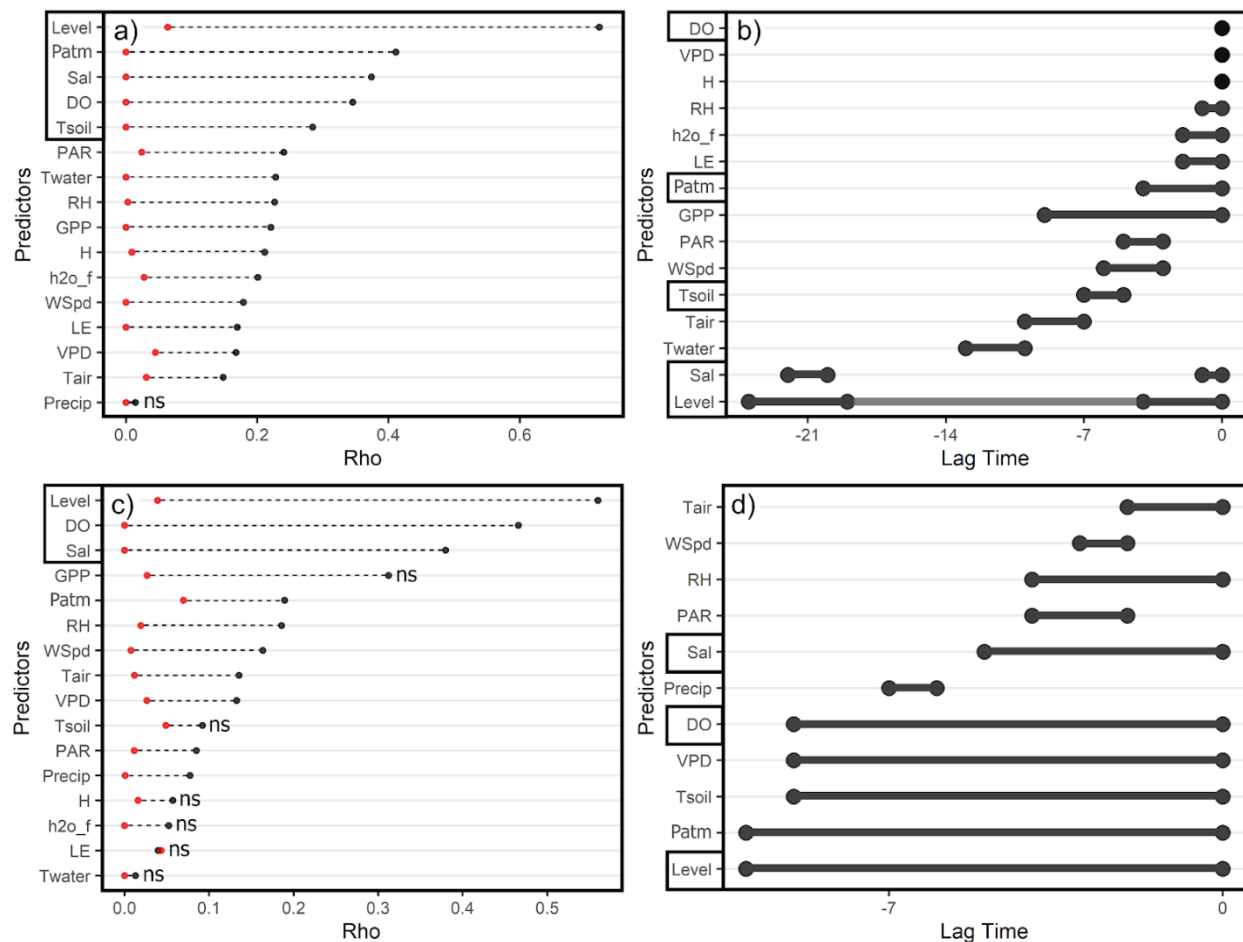
**Fig. 2:** Results from Convergent cross mapping (CCM) and extended CCM for seasonal dynamics from daily data spanning 2017-21. a) Identification of strongest predictor variables for CH<sub>4</sub> flux using CCM. Red points represent the maximum predictive skill achieved with simulated surrogate data. Black points indicate prediction skill of real data. ns denotes non-significant cross-mapping due to a lack of convergence or overlap with surrogate data results. Outlined variables exhibit a prediction skill > 0.6. b) Optimal time lags (in number of days) determined from extended CCM.

### 3.3 Causality Analysis with CCM

Examination of daily data with CCM revealed that the top causal variables at the seasonal/multiday scale ( $\rho > 0.70$ ) consisted of temperature components (Tsoil, Twater, Tair), DO, and



334 GPP (Fig. 2a). These select variables were a distinct grouping, with all other variables falling below a  
335 threshold of  $\rho=0.50$ . The remaining variables ( $H_2O$  flux, Level, LE, PAR, WSpd, Patm, RH, Precip;  
336 sorted from higher to lower coupling strength) were identified as causal ( $\rho=0.10-0.50$ ) except for Sal,  
337 VPD, and H, which did not meet convergence criteria or overlapped with surrogate data (Fig. S7). The  
338 growing phenoperiod was complex regarding the number of predictors and lags, yet physical factors also  
339 played an important role. Examination of hourly data for the diel scale analysis from the growing  
340 phenoperiod revealed Level as the top causal variable ( $\rho\sim 0.70$ ), followed by a grouping that included  
341 Patm, Sal, DO, and Tsoil ( $\rho=0.30-0.50$ ) (Fig. 3a). The remaining variables (PAR, Twater, RH, GPP, H,  
342  $H_2O$  flux, WSpd, LE, VPD, Tair; sorted from higher to lower coupling strength) also had a degree of  
343 coupling ( $\rho=0.15-0.25$ ), except Precip (Fig. S8). Examination of hourly data from the dormancy  
344 phenoperiod revealed Level, DO, and Sal as top causal variables ( $\rho=0.55-0.30$ ) (Fig. 3c). Most  
345 remaining variables (Patm, RH, WSpd, Tair, VPD, PAR, Precip; sorted from higher to lower coupling  
346 strength) also exhibited some coupling ( $\rho=0.10-0.20$ ), yet this period contained the largest number of  
347 non-significant variables, which included GPP, Tsoil, Twater, H, LE, and water fluxes (Fig. S9).



**Fig. 3:** Results from Convergent cross mapping (CCM) and extended CCM for diurnal/seasonal dynamics from select hourly data spanning 2020-21. a) Growing phenoperiod CCM at maximum library size between all variables and  $\text{CH}_4$  flux, b) Growing phenoperiod optimal time lags for extended CCM represented by the number of days (x-axis), c) Dormancy phenoperiod CCM at maximum library size between all variables and  $\text{CH}_4$  flux. d) Dormancy phenoperiod optimal time lags for extended CCM represented by the number of days (x-axis). Red points represent the maximum predictive skill achieved with simulated surrogate data. Black points indicate prediction skill of real data. ns denotes non-significant cross-mapping due to a lack of convergence or overlap with surrogate results. Outlined variables exhibit a prediction skill  $> 0.3$ .

### 3.4 Optimal Lags with Extended CCM

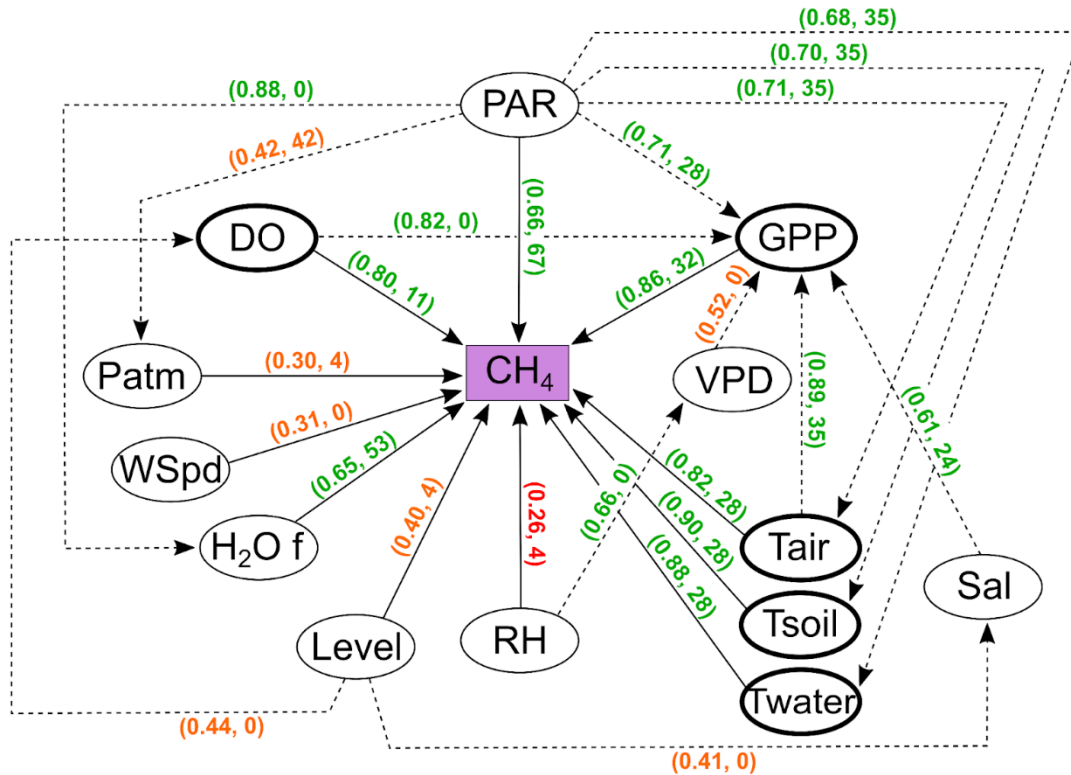
Implementation of extended CCM revealed that the influence of some variables is more immediate while others are associated with a lag. While any identified lags indicate the optimal time period of causal influence, these lags are not necessarily related to the overall strength of the causal relationship. However, longer lags can mean more indirect relationships within the context of causal network mapping. When applying daily data to unravel seasonal/multiday dynamics, we found that PAR, LE, and H<sub>2</sub>O fluxes are associated with the most extended lag times occurring between 49-70 days (seasonal/multiday scale) (moderate causal strength). These were followed by a grouping of top predictor variables identified by the previous CCM analysis (T<sub>soil</sub>, T<sub>water</sub>, T<sub>air</sub>, GPP) (high causal strength), which were most influential between 21-35 days (monthly scale). The remaining variables (DO, Level, WSpd, Patm, RH, Precip) exerted a close temporal relationship (0 days) at a range of causal strengths, but DO was important over a more extended period (0-3 weeks), followed by Level, Patm, and RH (0-1 week) and finally WSpd and Precip (0 days). However, Precip had the weakest coupling (Fig. 2b) (Fig. S10).

When applying hourly data from the growing phenoperiod to unravel diurnal dynamics, we found that lags were more complex. Both Level and Sal (strong and moderate causal strength, respectively) had peaks occurring between 0-5 hours and again between 19-23 hours, but the influence of Level was also persistent throughout the day (denoted as a gray bar in Fig. 3b). Temperature components had successively shorter lags moving from T<sub>water</sub> (9-13 hours) to T<sub>soil</sub> (5-7 hours) and finally T<sub>air</sub> (3-6 hours) (all moderate causal strength). The remaining variables were influential immediately (0 hours), with GPP extending the longest (0-9 hours), followed by Patm (0-5 hours), LE and water fluxes (0-3 hours), RH (0-1 hour) and finally DO, VPD, and H (0 hours) (moderate causal strength) (Fig. 3b) (Fig. S11). Lag times were much shorter for the dormancy phenoperiod. Almost all causal variables (Level, DO, Sal, Patm, RH, T<sub>air</sub>, VPD, T<sub>soil</sub>) exerted a close temporal relationship (0 hours) with DO, VPD, T<sub>soil</sub>, Patm, and Level active over the most extended period (0-10 hours). Precip was influential between

7-8 hours, yet again had the lowest causal strength while PAR, RH, WSpd, and Tair all exerted influence between 0-5 hours) (moderate causal strength) (Fig. 3d) (Fig. S12).

### 3.5 Causal Network Mapping

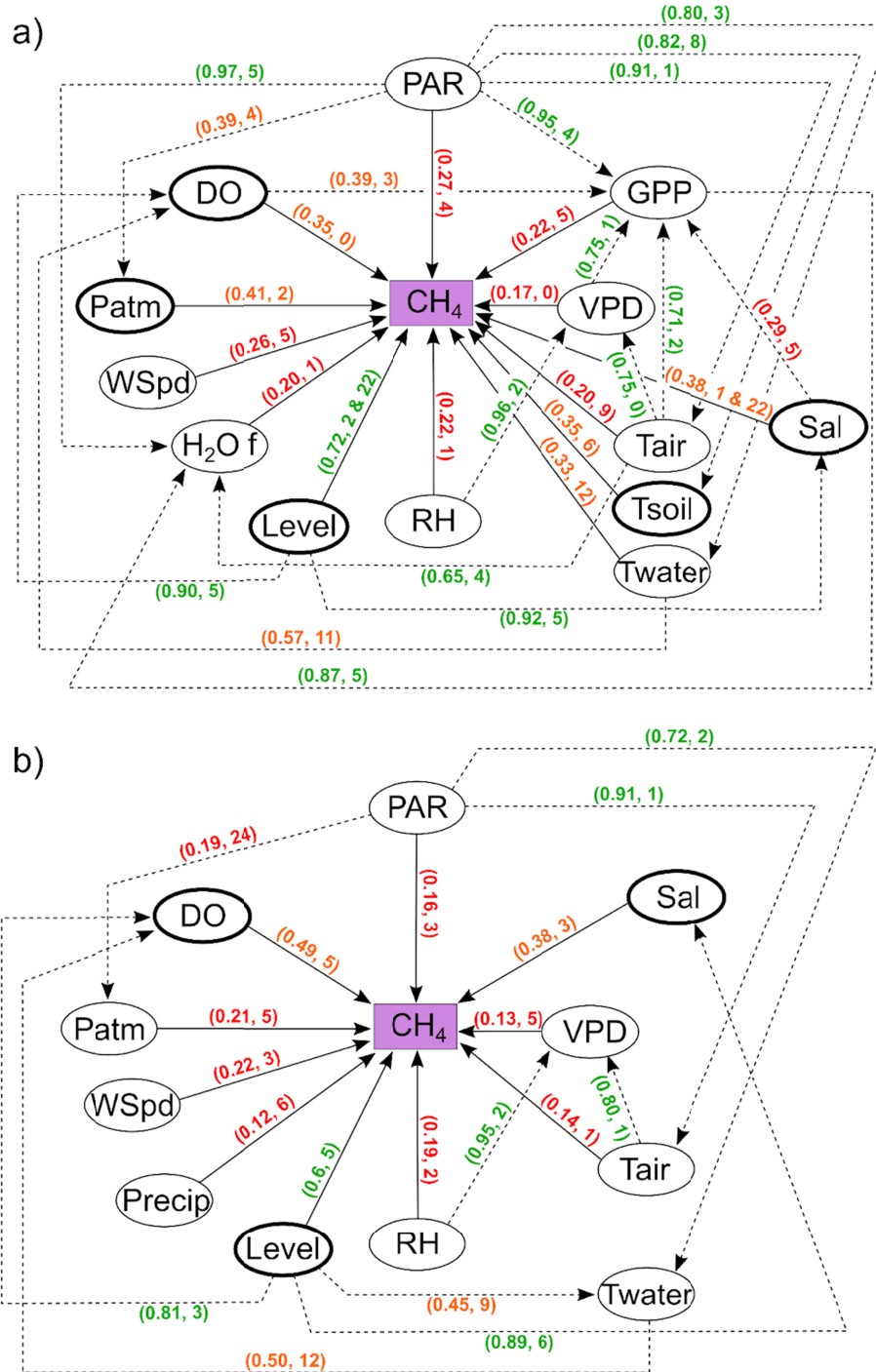
Causal network maps reveal high interconnection among variables at multiple scales (seasonal, multiday, and diurnal) and indicate that fluxes result from a suite of factors. When using daily data, we discovered several interesting patterns (Fig. 4). First, PAR is important for CH<sub>4</sub> ( $\rho=0.66$ ), but its action is likely indirect via greater influence on other relevant variables (e.g., H<sub>2</sub>O fluxes and temperature components) ( $\rho=0.65$  and  $0.90$ ). The strong relationship between GPP and CH<sub>4</sub> fluxes ( $\rho=0.86$ ) was influenced by VPD, Tair, Sal, and DO ( $\rho=0.52-0.82$ ), which are in turn influenced by RH, PAR, and Level, respectively ( $\rho=0.41-0.71$ ). Other variables, such as RH and WSpd, were not causally connected to any variables examined.



**Fig. 4:** Causal network map showing seasonal dynamics between all relevant variables for daily data spanning 2017-21. Numbers represent predictive skill followed by the median value of the optimum time lag. Colors represent the strength of coupling, green=strong (0.6-1.0), orange=moderate (0.30-0.59), and red (weak) (0-0.29). Solid lines represent direct relationships, and dashed lines represent indirect relationships. Variables outlined in bold represent the top causal variables identified by CCM analysis.

When applying hourly data for the growing phenoperiod, many highly connected variables identified from the seasonal analysis (Temperature components, GPP, DO) exert far less influence (Fig. 5a). The water level was the most directly important variable ( $\rho=0.72$ ) and was highly connected to both Sal and DO ( $\rho=0.90-0.92$ ), which exerted moderate influence on CH<sub>4</sub> fluxes ( $\rho=0.35-0.38$ ). Over seven additional variables provided moderate-low influence, but all showed a similar predictive skill ( $\rho=0.17-0.41$ ), highlighting the complexity associated with the sub-daily growing phenoperiod. Hourly data from the dormancy phenoperiod exhibited the most straightforward network connectivity with the

408 least interdependencies (Fig. 5b). As with growing phenoperiod data, Level was the most directly  
409 important variable ( $\rho=0.60$ ) and exerted a strong influence on Sal and DO ( $\rho=0.81-89$ ), which exerted  
410 moderate influence on CH<sub>4</sub> fluxes ( $\rho=0.38-0.49$ ). Additional variables were associated with weaker  
411 casual relationships ( $\rho=0.12-0.22$ ), and only temperature components, VPD, Patm, Sal, and DO were  
412 dependent on the activity of other variables during this phenoperiod.

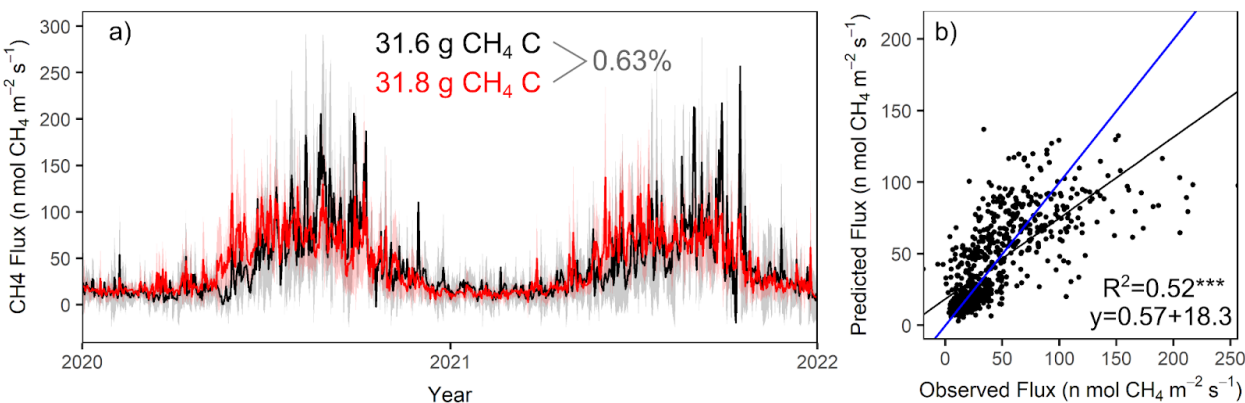


**Fig. 5:** Causal network maps showing diurnal/seasonal dynamics between all relevant variables for select hourly data spanning 2020-21 for a) Growing phenoperiod and b) Dormancy phenoperiod. Numbers represent predictive skill followed by the median value of the optimum time lag. Colors represent the strength of coupling, green=strong (0.6-1.0), orange=moderate (0.30-0.59), and red (weak) (0-0.29). Solid

lines represent direct relationships, and dashed lines represent indirect relationships. Variables outlined in bold represent the top causal variables identified by CCM analysis.

### 3.6 Modeling Output

The predictive multivariate EDM model fell within 1% of the expected 2-year annual sum derived from the gap-filled EC benchmark for CH<sub>4</sub> flux (31.6 g C m<sup>-2</sup>) (Fig. 7). This model incorporated all of the positively coupled variables identified by CCM analysis including temperature components of soil, water, air, GPP, DO, H<sub>2</sub>O flux, Level, LE, PAR, WSpd, Patm, RH, and Precip. This generated a 2-year sum of 31.8 g C m<sup>-2</sup> (0.63% difference; slope 0.57) with an annual uncertainty of +/- 29.3 g C m<sup>-2</sup>, calculated from the state space. This approach had challenges representing high fluxes during senescence and despite good overall predictions, carried a high overall uncertainty and small over-prediction bias for lower magnitude fluxes (< 50 nmol m<sup>-2</sup> s<sup>-1</sup>) (Fig. 6b). It should be noted that the uncertainty estimated for the EDM model is based on the nearest neighbor distance weights within the projected state space and although we scaled predictor variables, including more variables can distort the state space to a greater extent.



**Fig. 6:** a) Multivariate EDM modeling results at the daily time step for prediction years 2020-21 and b) linear regression comparisons. Black lines represent daily CH<sub>4</sub> flux measured by eddy covariance (EC) as a benchmark +/- daily sd (grey shading), and red lines represent daily modeled CH<sub>4</sub> flux +/- prediction variance sd (pink shading). The reported black sum is the 2-year total carbon budget from CH<sub>4</sub> calculated



with EC data (benchmark), and the red sum is the 2-year total carbon budget from CH<sub>4</sub> estimated with modeled data and percent difference (gray). The blue line on regression plots is a 1:1 line.

## 4. Discussion

### 4.1 Comparisons with Previous Wetland CH<sub>4</sub> Studies

There are limited CH<sub>4</sub> studies in salt marshes but it has been reported by linear methods that temperature, PAR, plant productivity, or some proxy (e.g., GPP, biomass, chlorophyll concentration), water level, and salinity are important predictors for CH<sub>4</sub> fluxes (Abdul-Aziz et al. 2018, Huertas et al. 2019, Martin and Moseman-Valtierra 2017, Poffenbarger et al. 2011.) Despite differing temporal averaging periods (daytime, nighttime, year-round, multi-year) and different model structures (linear regression, multivariate canonical correlation, multivariate generalized least squares), multiple approaches have also been able to identify the importance of temperature, DO, and Patm for CH<sub>4</sub> dynamics (Vázquez-Lule and Vargas 2021, Hill and Vargas 2022). Other multi-site studies have also identified similar predictor variables with both linear and other nonlinear methods, especially temperature components, in non-tidal freshwater wetlands (Knox et al, 2020). The mechanisms supported by additional variables such as water level and GPP are more complex, with influence often manifesting across several temporal scales (Sturtevant et al. 2016). Our results indicate that the effects of GPP are highly dependent on the time domain at multiple scales (seasonal and diurnal) and multivariate interactions among variables. While linear models may be capable of extracting similar variables in many cases, these methods fail to identify dependencies, lags, and complex interrelationships. Thus, linear approaches may still limit our interpretations because the influence of one variable can be the result of indirect influence from a third variable, a phenomenon that can be indirectly inferred when applying CCM.

Several studies based on non-tidal freshwater wetlands have successfully applied alternative nonlinear analysis methods (Knox et al. 2021, Sturtevant et al. 2016) or machine learning approaches such as ANN's or random forest techniques (Rey-Sanchez et al. 2018, Zhu et al. 2013, Abbasi et al. 2019, Zaki and Abdul-Aziz 2022, Morin et al. 2017). In these non-tidal freshwater wetlands, interactions regulating fluxes do not include the same dynamics found within tidal wetlands including frequent oscillation of water level and subsequent modulation of sediment salinity, dissolved oxygen, and redox conditions (Vazquez-Lule and Vargas 2021). Further, dominant vegetation species differ, which can influence the biogeochemical conditions of the sediments and CH<sub>4</sub> production (Seyfferth et al. 2020, Yuan et al. 2016, Gao et al. 2018). Despite differences, these studies also show that the strongest relevant drivers are identified by both linear and nonlinear approaches, yet incomplete results from solely applying linear methods limit the full mechanistic understanding of the system (Sturtevant et al. 2016). Specifically, issues of cross-correlation can restrict which variables are included to satisfy model requirements, and important lags and interrelationships fail to be identified. In fact, tidal saltwater wetlands have been purposely excluded from several FLUXNET- CH<sub>4</sub> analyses because they do not follow the general pattern of linear relationships defined by freshwater wetlands, and an overall lack of data from these highly dynamic ecosystems creates difficulties when attempting to generalize across multiple sites (Chang et al. 2021, Delwiche et al. 2021, Knox et al. 2021).

In limited cases where nonlinear methods are applied, findings largely support our results, demonstrating lags with GPP, temperature, and water level, but with temporal synchrony with VPD and energy fluxes (Delwiche et al. 2021). For example, in a multi-site synthesis study of freshwater sites, median seasonal lags associated with water level extended 17±11 days, 8±16, and 5±14 days for Tair and Tsoil, respectively, and were synchronous for Patm while at the diel scale, GPP was lagged up to 4 hours and energy fluxes and VPD were synchronous (<1 hour) (Knox et al. 2021). A similar study from a forested/shrub wetland found CH<sub>4</sub> lags associated with GPP extending up to 60 days (Turner et al. 2021), potentially indicating different time dependencies based on vegetation type and structure. There is overwhelming evidence supporting that temporal lags exist between multiple environmental variables and

CH<sub>4</sub> fluxes across different types of wetlands. Thus, it is important to include methods such as CCM to identify such lags in the context of causality to properly represent the complexity of how biophysical drivers control CH<sub>4</sub> fluxes

## 4.2 Seasonal CH<sub>4</sub> Dynamics

Seasonal CH<sub>4</sub> dynamics are strongly controlled by temperature components, GPP, and DO, indicating reliance on biological factors. Previous studies showed that predictors such as Level, Patm, and VPD are important in a linear multivariate context but rank lower in causality in the current study (Vázquez-Lule and Vargas 2021, Hill and Vargas 2022). This is likely because these variables are less strongly coupled with CH<sub>4</sub> activity at this scale due to: a) limited occurrence as with transient low-pressure systems, which operate synchronously with CH<sub>4</sub> emissions and rarely exert influence on scales extending beyond several days (Fig. 2b) (Mønster, Kjeldsen, and Scheutz 2019, Tokida et al. 2007, Knox et al. 2021); or b) indirect mechanisms as with Patm and pressurized gas flow (Björn et al. 2022, Zhang and Ding 2011), and VPD limiting stomatal conductance and the possible decrease in plant-mediated transport. At the seasonal scale, temperature is a strong predictor in temperate ecosystems because it regulates the window of active plant growth and belowground microbial metabolic rates (Zhu et al. 2019, Yvon-Durocher et al. 2014), influencing below-ground storage pools and, ultimately, sediment-atmosphere fluxes (Reid et al. 2013, Capooci and Vargas 2022). GPP also has strong coupling and may suggest there is connectivity or flow between plant photoassimilates and methanogens in the rhizosphere, which increases with seasonal biomass development. We observed a 32-day median lag time between GPP and effects on CH<sub>4</sub> flux, similar to the lag identified for temperature (35 days), which may indicate structural changes in the vegetation are responsible for the observed seasonal increase in CH<sub>4</sub> flux. Weaker couplings were associated with physical variables such as Patm and lag times for these variables were much shorter. Thus, it appeared that biological vs. physical processes operate at different temporal scales.

Salinity has previously been identified as an important daily-scale predictor by linear modeling but had no direct causal influence on seasonal CH<sub>4</sub> fluxes in this study. While salinity causes direct inhibitory effects for methanogens by limiting the availability of sulfate substrates via microbial competition in most wetlands (Reddy and DeLaune 2008), there is evidence for the presence of methylophilic bacteria in salt marshes which produce CH<sub>4</sub> from non-competitive substrates resulting in elevated emissions despite the presence of sulfate (Seyfferth et al. 2020, Capoori and Vargas 2022). Evidence to support this phenomenon exists from a variety of wetland studies that either lack *S. alterniflora* and thus more closely follow the paradigm of CH<sub>4</sub> inhibition (Chmura, Kellman, and Guntenspergen 2011, Krauss et al. 2016, Poffenbarger et al. 2011), or contain *S. alterniflora* and emit methane above or at the high-end of a widely accepted CH<sub>4</sub>-sulfate threshold (Poffenbarger et al. 2011, Oremland et al. 1982, Huertas et al. 2019, Rosentreter et al. 2021). We propose that the elevated emissions, especially during senescence, are tied to the release of secondary plant compounds (osmolytes) associated with *S. alterniflora*, including glycine betaine and dimethyl sulfoniopropionate which are broken down via fermentation to trimethylamine and dimethyl sulfide and are exclusively utilized by methylophilic methanogens (Husband and Kiene 2007, Jones et al. 2019). It is also important to note that this marsh rarely experiences salinity extremes over 20ppt, thus sites influenced by higher salinity may experience different characteristics.

#### 4.3 Diurnal CH<sub>4</sub> Dynamics

Diurnal CH<sub>4</sub> dynamics during the growing phenoperiod are controlled by multiple factors with many interdependencies among variables, yet the most important causal variables are Level, Patm, Sal, DO, and Tsoil. This indicates that CH<sub>4</sub> fluxes over the diurnal course are controlled by both biochemical factors likely tied to plant and microbial activity (Sal, DO, Tsoil) and physical factors likely tied to emissions (Level, Patm) (Moore and Roulet 1993, Reid, et al. 2015, Knox et al. 2021, Liu et al. 2019). While several studies have made attempts at reducing this level of complexity by applying parsimonious models that only incorporate the most important drivers (via linear methods) (Levy et al. 2012, Abdul-

Aziz et al. 2018, Baird et al. 2019), this may only be useful for generating coarse-scale information. Other studies have applied more complex process-based models, although these require an extensive mechanistic understanding and often experience difficulties when moving across study sites (Zhang et al. 2020, Li et al. 2016), where interacting variables and driving processes can differ considerably (Krauss et al. 2016, Rosentreter et al. 2021). This can limit the range of potential applications and may be less relevant for coastal wetlands where relationships can differ greatly from site to site.

The dormancy phenoperiod is comparatively simple, with the most important causal variables consisting of Level, DO, and Sal. This is also supported by considerably less interconnection between variables. Considering the lack of active vegetation, we can assume mechanisms related to these predictors change in relation to the growing or year-round daily averaging periods (Bansal et al. 2020, Reid et al. 2013). Although the relationship is weak, coupling between CH<sub>4</sub> and Sal increases following the end of senescence and the preceding seasonal peak in emissions. Because vegetation is no longer present, the flow of photoassimilates stops, and availability of osmolyte byproducts likely decreases within sediments. Thus, Sal may become more relevant during dormancy as competition between sulfate-reducing bacteria and methanogens likely increases (Derby et al. 2021, Poffenbarger et al. 2011), while the spatial separation between production zones at depth may be less exposed to sulfate fluctuations from tidal exchange (Koebsch et al. 2019). Further, plant-mediated transport is curtailed, leading to greater importance of conditions at the sediment interface (Wang and Han 2005). This is supported by the lack of causal influence from associated variables such as GPP, LE, and H<sub>2</sub>O flux. This also likely explains the importance of DO in tidal waters, which reaches a concentration peak during dormancy, likely influencing CH<sub>4</sub> oxidation rates. In either case, the top hourly predictor for both growing and dormancy phenoperiods is Level that oscillates between low and high tides at six-hour intervals but also shifts on a 14-day spring/neap tide and is highly tied to fluctuations of DO and Sal (Vázquez-Lule and Vargas 2021). Our results show that while all these variables are related and often highly cross-correlated with linear methods, each of these variables exerts a unique influence on CH<sub>4</sub> dynamics, yet there is also strong

interdependence with how these variables change in time, highlighting the importance of incorporating true nonlinear analysis methods.

#### 4.4 Relevance of Time Lags

On a seasonal/multiday scale, almost all causally significant variables were associated with lag times where influence is exerted across several consecutive days or weeks. The most important causal variables (temperature components and GPP) had medium lag times (i.e., “slow”) near 3-5 weeks, likely associated with the possible threshold effect with the temperature-driven initiation of plant growth (O'Connell, Alber and Pennings 2020), or increased production and transport of photoassimilates to the rhizosphere (Knox et al. 2021). This lag time (3 weeks) was also associated with the effect of salinity on GPP, likely indicating increases in plant osmolyte production coinciding with initial spring biomass accumulation (Husband and Kiene 2007, Mulholland and Otte 2000). It is important to recognize that we are measuring CH<sub>4</sub> fluxes (the physical process of release); thus, these observed lags are also likely associated with the disconnection between production and observed fluxes. The shortest lags exert immediate (<1 week) (i.e., “fast”) but also prolonged impact extending up to 1 week for Level, Patm, and RH and up to 4 weeks for DO. This effect from Level is likely a result of high/low water levels tied to spring/neap tide cycles, which oscillate on a 6-7 day period. Yet, there are also fluctuations at shorter time periods, both of which impact soil water content and biogeochemical conditions of the sediments (Seyfferth et al. 2020). Tidal waters also directly influence salinity and DO levels within sediments, but the effect of DO is prolonged and likely tied to its strong influence on GPP, as shown by our results (Fig. 4). The effects of Patm and RH are likely tied to mesoscale circulation and air mass transit which modify the air column and provide a physical forcing mechanism (Tokida et al. 2007). The longest lag times were associated with LE, H<sub>2</sub>O flux, and PAR, which likely represents a major seasonal oscillation driven by solar angle and changes in available radiation (Guo et al. 2010). While only weakly causal, there were no lags associated with WSpd or Precip, indicating an effect exclusively at the sub-daily level as, in most cases, the occurrence of wind or rain rarely extends into multi-day events.

During the growing phenoperiod, top causal variables identified by CCM analysis (Level, Patm, Sal, DO, Tsoil) all operated at different peak lag times, highlighting the complexity of observing and measuring fluxes during this time. The most important causal variable (Level) exerts influence over the entire 24-hour period yet peaks in causal effect at 0-5 hours and 20-24 hours, near when Sal (moderately casual) was influential, which is likely tied to semi-diurnal tidal cycling and water levels which are typically higher during nighttime high tide (Forbrich and Giblin 2015). Influence over the entire 24-hour period may be due to a smearing-out effect as the tide shifts its range an hour each day. Patm and DO (moderately causal) were associated with a zero-lag time but the effect of Patm extended up to 5 hours. These lags are likely related to fast (immediate) oxidation activity across sediment, daily pressure fluctuations, and convective flow-through transport of CH<sub>4</sub> via plant-mediated transport driven by solar activity (Kim, Verma and Billesbach 1999, Ding and Cai 2007). Although the overall causality was weaker and distributed across variables, many of the fast variables during the growing phenoperiod are related to energy exchange dynamics (LE and H) and tied to changes in evapotranspiration (H<sub>2</sub>O flux, VPD, RH). This is likely explained by combined evaporation from saturated sediments where CH<sub>4</sub> has accumulated in porewater, evaporation from areas of open water that receive a lateral influx of dissolved CH<sub>4</sub>, and plant-mediated transport factors reliant on stomatal activity, which channel dissolved CH<sub>4</sub> in the transpiration stream (Megonigal and Guenther 2008, Sun et al. 2012, Trifunovic et al. 2020). GPP is both a fast and slow variable indicating a supply of photoassimilates is triggered upon initiation of morning photosynthesis but with lags extending up to 8 hours when photosynthesis gears down in the evening.

During the dormancy phenoperiod, top causal variables identified by CCM analysis (Level, Sal, DO) all had immediate influence but were influential at lag times extending up to 5-10 hours, highlighting the importance of tidal activity and modulation of associated variables (Sal and DO). While the remaining variables were only weakly causal, several interesting patterns exist. In contrast to the slower effect of Tair during the growing phenoperiod, Tair was a fast variable with short lag times in dormancy, while the effect of Tsoil had longer lags, up to 9 hours, indicating a connection to microbial activity and temperature thresholds associated with CH<sub>4</sub> production (Chadburn et al. 2020). All top causal variables

were related to tidal activity (Level, DO, Sal), indicating that hydrological factors and the riverine-ocean connectivity become crucial during the dormancy phenoperiod. The shorter lag times during this phenoperiod may also indicate more direct coupling with physical factors and less complex interactions involving the presence of vegetation, such as substrate supply and plant-mediated transport (Vázquez-Lule and Vargas 2021). Taken together, these results highlight the multi-mechanistic activities and processes which regulate CH<sub>4</sub> fluxes.

#### **4.5 Multivariate EDM Modeling**

This study demonstrates high complexity in the biophysical controls of CH<sub>4</sub> fluxes in tidal wetlands. These results are based on high-frequency measurements and long-term observations (5 years) that contribute to our understanding of CH<sub>4</sub> dynamics. The multivariate EDM mode provided predictions within 1% of the CH<sub>4</sub> carbon sum measured from the EC benchmark. This model captured most of the higher magnitude fluctuations and high emission peaks during the growing phenoperiod, yet omitted higher magnitude fluctuations during senescence. Despite good performance, this model still considerably underpredicted during senescence and overpredicted during green-up. Consequently, these transitions represent a challenge as rapid metabolic and structural changes happen in the plant canopy (Vazquez-Lule et al 2022). Although we included the full suite of identified causal variables from CCM, this could indicate that additional variables omitted or not measured (e.g., sediment redox potential, biomass, sediment substrate availability) are essential to explain these unusually high fluxes during senescence.

#### **5. Conclusion**

We demonstrate that nonlinear state space methods uncover the complex dynamics and interdependencies among predictors and their influence on CH<sub>4</sub> flux. The technique can tease apart not only the strength of causal variables but also identify lags, and complex interconnections among the full suite of measured variables as past information about driving variables becomes embedded within the presently observed CH<sub>4</sub> time series. Overall, CCM was able to identify and rank multiple causal variables while providing a holistic picture of variable interactions and lagged relationships, which is critical for an



accurate mechanistic understanding of system behavior. We highlight that CCM is a powerful technique that provides mechanistic information regarding causality, lags, and interconnectivity among an extensive collection of potentially important variables. Therefore, by complementing CCM with multivariate EDM we can generate interpretable empirical models that provide insights into controlling mechanisms of ecosystem processes.

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## **Open Research**

Meteorological (station: delsjmet-p) and water quality (station: Aspen Landing) data are available from the National Estuarine Research Reserve's Centralized Data Management Office (CDMO) at <https://cdmo.baruch.sc.edu/>. Phenological data are available from the PhenoCam network (site: stjones) at <https://phenocam.sr.unh.edu/webcam/sites/stjones/>. Eddy covariance data are available via AmeriFlux (US-StJ; <https://ameriflux.lbl.gov/sites/siteinfo/US-StJ>).

## 655    **References**

- 656    Abbasi, T., T. Abbasi, C. Luithui & S. A. Abbasi (2019) Modelling Methane and Nitrous Oxide  
657    Emissions from Rice Paddy Wetlands in India Using Artificial Neural Networks (ANNs). *Water*, 11.
- 658    Abdul-Aziz, O. I., K. S. Ishtiaq, J. W. Tang, S. Moseman-Valtierra, K. D. Kroeger, M. E. Gonneea, J.  
659    Mora & K. Morkeski (2018) Environmental Controls, Emergent Scaling, and Predictions of Greenhouse  
660    Gas (GHG) Fluxes in Coastal Salt Marshes. *Journal of Geophysical Research-Biogeosciences*, 123, 2234-  
661    2256.
- 662    Al-Haj, A. N. & R. W. Fulweiler (2020) A synthesis of methane emissions from shallow vegetated  
663    coastal ecosystems. *Global Change Biology*, 26, 2988-3005.
- 664    Baird, A., S. Green, E. Brown & G. Dooling (2019) Modelling time-integrated fluxes of CO<sub>2</sub> and CH<sub>4</sub> in  
665    peatlands: A review. *Mires and Peat*, 24.
- 666    Baldocchi, D. D. (2003) Assessing the eddy covariance technique for evaluating carbon dioxide exchange  
667    rates of ecosystems: past, present and future. *Global Change Biology*, 9, 479-492.
- 668    Bansal, S., O. F. Johnson, J. Meier & X. Zhu (2020) Vegetation Affects Timing and Location of Wetland  
669    Methane Emissions. *Journal of Geophysical Research: Biogeosciences*, 125, e2020JG005777.
- 670    Barba, J., A. Cueva, M. Bahn, G. A. Barron-Gafford, B. Bond-Lamberty, P. J. Hanson, A. Jaimes, L.  
671    Kulmala, J. Pumpanen, R. L. Scott, G. Wohlfahrt & R. Vargas (2018) Comparing ecosystem and soil  
672    respiration: Review and key challenges of tower-based and soil measurements. *Agricultural and Forest*  
673    *Meteorology*, 249, 434-443.
- 674    Barton, K. & M. K. Barton (2015) Package ‘mumin’. *Version*, 1, 439.
- 675    Bridgman, S. D., H. Cadillo-Quiroz, J. K. Keller & Q. L. Zhuang (2013) Methane emissions from  
676    wetlands: biogeochemical, microbial, and modeling perspectives from local to global scales. *Global*  
677    *Change Biology*, 19, 1325-1346.
- 678    Capooci, M. & R. Vargas (2022) Trace gas fluxes from tidal salt marsh soils: implications for carbon-  
679    sulfur biogeochemistry. *Biogeosciences*, 19, 4655-4670.
- 680    Chadburn, S. E., T. Aalto, M. Aurela, D. Baldocchi, C. Biasi, J. Boike, E. J. Burke, E. Comyn-Platt, A. J.  
681    Dolman, C. Duran-Rojas, Y. Fan, T. Friborg, Y. Gao, N. Gedney, M. Göckede, G. D. Hayman, D. Holl,  
682    G. Hugelius, L. Kutzbach, H. Lee, A. Lohila, F.-J. W. Parmentier, T. Sachs, N. J. Shurpali & S.  
683    Westermann (2020) Modeled Microbial Dynamics Explain the Apparent Temperature Sensitivity of  
684    Wetland Methane Emissions. *Global Biogeochemical Cycles*, 34, e2020GB006678.
- 685    Chang, C.-W., M. Ushio & C.-h. Hsieh (2017) Empirical dynamic modeling for beginners. *Ecological*  
686    *Research*, 32, 785-796.
- 687    Christensen, T. R., A. Ekberg, L. Ström, M. Mastepanov, N. Panikov, M. Öquist, B. H. Svensson, H.  
688    Nykänen, P. J. Martikainen & H. Oskarsson (2003) Factors controlling large scale variations in methane  
689    emissions from wetlands. *Geophysical Research Letters*, 30.

690 Chu, X. J., G. X. Han, Q. H. Xing, J. Y. Xia, B. Y. Sun, J. B. Yu & D. J. Li (2018) Dual effect of  
691 precipitation redistribution on net ecosystem CO<sub>2</sub> exchange of a coastal wetland in the Yellow River  
692 Delta. *Agricultural and Forest Meteorology*, 249, 286-296.

693 Dean, J. F., J. J. Middelburg, T. Röckmann, R. Aerts, L. G. Blauw, M. Egger, M. S. M. Jetten, A. E. E. de  
694 Jong, O. H. Meisel, O. Rasigraf, C. P. Slomp, M. H. in't Zandt & A. J. Dolman (2018) Methane  
695 Feedbacks to the Global Climate System in a Warmer World. *Reviews of Geophysics*, 56, 207-250.

696 Delwiche, K. B., S. H. Knox, A. Malhotra, E. Fluet-Chouinard, G. McNicol, S. Feron, Z. Ouyang, D.  
697 Papale, C. Trotta, E. Canfora, Y. W. Cheah, D. Christianson, M. C. R. Alberto, P. Alekseychik, M.  
698 Aurela, D. Baldocchi, S. Bansal, D. P. Billesbach, G. Bohrer, R. Bracho, N. Buchmann, D. I. Campbell,  
699 G. Celis, J. Chen, W. Chen, H. Chu, H. J. Dalmagro, S. Dengel, A. R. Desai, M. Detto, H. Dolman, E.  
700 Eichelmann, E. Euskirchen, D. Famulari, K. Fuchs, M. Goeckede, S. Gogo, M. J. Gondwe, J. P.  
701 Goodrich, P. Gottschalk, S. L. Graham, M. Heimann, M. Helbig, C. Helfter, K. S. Hemes, T. Hirano, D.  
702 Hollinger, L. Hörtnagl, H. Iwata, A. Jacotot, G. Jurasinski, M. Kang, K. Kasak, J. King, J. Klatt, F.  
703 Koebisch, K. W. Krauss, D. Y. F. Lai, A. Lohila, I. Mammarella, L. Belelli Marchesini, G. Manca, J. H.  
704 Matthes, T. Maximov, L. Merbold, B. Mitra, T. H. Morin, E. Nemitz, M. B. Nilsson, S. Niu, W. C.  
705 Oechel, P. Y. Oikawa, K. Ono, M. Peichl, O. Peltola, M. L. Reba, A. D. Richardson, W. Riley, B. R. K.  
706 Runkle, Y. Ryu, T. Sachs, A. Sakabe, C. R. Sanchez, E. A. Schuur, K. V. R. Schäfer, O. Sonnentag, J. P.  
707 Sparks, E. Stuart-Haëntjens, C. Sturtevant, R. C. Sullivan, D. J. Szutu, J. E. Thom, M. S. Torn, E. S.  
708 Tuittila, J. Turner, M. Ueyama, A. C. Valach, R. Vargas, A. Varlagin, A. Vazquez-Lule, et al. (2021)  
709 FLUXNET-CH<sub>4</sub>: a global, multi-ecosystem dataset and analysis of methane seasonality from freshwater  
710 wetlands. *Earth Syst. Sci. Data*, 13, 3607-3689.

711 Deyle, E. R., May, R. M., Munch, S. B., & Sugihara, G. (2016). Tracking and forecasting ecosystem  
712 interactions in real time. *Proc Biol Sci*, 283(1822).

713 Derby, R. K., B. A. Needelman, A. A. Roden & J. P. Megonigal (2021) Stratifying by Vegetation and  
714 Hydrology Improves Tidal Marsh Methane Emission Accounting.

715 Fan, H., Y. Meng, D. Wang, Y. Zhao & C. Zhao. 2020.. In *2020 35th Youth Academic Annual*  
716 *Conference of Chinese Association of Automation (YAC)*, 347-352. IEEE.

717 Filippa, G., E. Cremonese, M. Migliavacca, M. Galvagno, M. Forkel, L. Wingate, E. Tomelleri, U. Morra  
718 di Cella & A. D. Richardson (2016) Phenopix: A R package for image-based vegetation phenology.  
719 *Agricultural and Forest Meteorology*, 220, 141-150.

720 Gao, G. F., P. F. Li, Z. J. Shen, Y. Y. Qin, X. M. Zhang, K. Ghoti, X. Y. Zhu & H. L. Zheng (2018)  
721 Exotic *Spartina alterniflora* invasion increases CH<sub>4</sub> while reduces CO<sub>2</sub> emissions from mangrove wetland  
722 soils in southeastern China. *Scientific Reports*, 8.

723 Geoghegan, E. K., J. S. Caplan, F. N. Leech, P. E. Weber, C. E. Bauer & T. J. Mozdzer (2018) Nitrogen  
724 enrichment alters carbon fluxes in a New England salt marsh. *Ecosystem Health and Sustainability*, 4,  
725 277-287.

726 Glagolev, M. V., O. R. Kotsyurbenko, A. F. Sabrekov, Y. V. Litti & I. E. Terentieva (2021)  
727 Methodologies for Measuring Microbial Methane Production and Emission from Soils-A Review.  
728 *Microbiology*, 90, 1-19.

729 Granger, C. W. (1969) Investigating causal relations by econometric models and cross-spectral methods.  
730 *Econometrica: journal of the Econometric Society*, 424-438.

731 Granse, D., J. Titschack, M. Ainouche, K. Jensen & K. Koop-Jakobsen (2022) Subsurface aeration of  
732 tidal wetland soils: Root-system structure and aerenchyma connectivity in *Spartina* (Poaceae). *Science of*  
733 *The Total Environment*, 802, 149771.

734 Guo, H. Q., B. Zhao, J. Q. Chen, Y. E. Yan, B. Li & J. K. Chen (2010) Seasonal Changes of Energy  
735 Fluxes in an Estuarine Wetland of Shanghai, China. *Chinese Geographical Science*, 20, 23-29.

736 Guo, Z., M. Hao, B. Yu & B. Yao (2022) Detecting delay propagation in regional air transport systems  
737 using convergent cross mapping and complex network theory. *Transportation Research Part E: Logistics*  
738 *and Transportation Review*, 157, 102585.

739 Helfter, C., M. Gondwe, M. Murray-Hudson, A. Makati, M. F. Lunt, P. I. Palmer & U. Skiba (2022)  
740 Phenology is the dominant control of methane emissions in a tropical non-forested wetland. *Nature*  
741 *Communications*, 13, 133.

742 Hill, A. & R. Vargas. 2021. Methane and Carbon Dioxide Fluxes in a Temperate Salt Marsh:  
743 Comparisons Between Plot, Ecosystem and Component Measurements. B25C-1465.

744 Hill, A. C. & R. Vargas (2022) Methane and Carbon Dioxide Fluxes in a Temperate Tidal Salt Marsh:  
745 Comparisons Between Plot and Ecosystem Measurements. *Journal of Geophysical Research:*  
746 *Biogeosciences*, 127, e2022JG006943.

747 Hsieh, C.-h., S. M. Glaser, A. J. Lucas & G. Sugihara (2005) Distinguishing random environmental  
748 fluctuations from ecological catastrophes for the North Pacific Ocean. *Nature*, 435, 336-340.

749 Huertas, I. E., M. de la Paz, F. F. Perez, G. Navarro & S. Flecha (2019) Methane Emissions From the Salt  
750 Marshes of Donana Wetlands: Spatio-Temporal Variability and Controlling Factors. *Frontiers in Ecology*  
751 *and Evolution*, 7.

752 Husband, J. D. & R. P. Kiene (2007) Occurrence of dimethylsulfoxide in leaves, stems, and roots of  
753 *Spartina alterniflora*. *Wetlands*, 27, 224-229.

754 Jackson, R. B., M. Saunio, P. Bousquet, J. G. Canadell, B. Poulter, A. R. Stavert, P. Bergamaschi, Y.  
755 Niwa, A. Segers & A. Tsuruta (2020) Increasing anthropogenic methane emissions arise equally from  
756 agricultural and fossil fuel sources. *Environmental Research Letters*, 15, 071002.

757 Jones, H. J., Kröber, E., Stephenson, J., Mausz, M. A., Jameson, E., Millard, A., Purdy, K. J., & Chen, Y.  
758 (2019) A new family of uncultivated bacteria involved in methanogenesis from the ubiquitous osmolyte  
759 glycine betaine in coastal saltmarsh sediments. *Microbiome*, 7(1), 120.

760 Kim, D. S. (2007) Greenhouse Gas (CH<sub>4</sub>, CO<sub>2</sub>, N<sub>2</sub>O) Emissions from Estuarine Tidal and Wetland and  
761 Their Characteristics. *Journal of Korean Society for Atmospheric Environment*, 23, 225-241.

762 Kim, Y., M. S. Johnson, S. H. Knox, T. A. Black, H. J. Dalmagro, M. Kang, J. Kim & D. Baldocchi  
763 (2020) Gap-filling approaches for eddy covariance methane fluxes: A comparison of three machine  
764 learning algorithms and a traditional method with principal component analysis. *Global Change Biology*,  
765 26, 1499-1518.

766 Knox, S. H., S. Bansal, G. McNicol, K. Schafer, C. Sturtevant, M. Ueyama, A. C. Valach, D. Baldocchi,  
767 K. Delwiche & A. R. Desai (2021) Identifying dominant environmental predictors of freshwater wetland  
768 methane fluxes across diurnal to seasonal time scales. *Global change biology*.

769 Knox, S. H., J. H. Matthes, C. Sturtevant, P. Y. Oikawa, J. Verfaillie & D. Baldocchi (2016) Biophysical  
770 controls on interannual variability in ecosystem-scale CO<sub>2</sub> and CH<sub>4</sub> exchange in a California rice paddy.  
771 *Journal of Geophysical Research-Biogeosciences*, 121, 978-1001.

772 Koebisch, F., M. Winkel, S. Liebner, B. Liu, J. Westphal, I. Schmiedinger, A. Spitzzy, M. Gehre, G.  
773 Jurasinski, S. Köhler, V. Unger, M. Koch, T. Sachs & M. E. Böttcher (2019) Sulfate deprivation triggers  
774 high methane production in a disturbed and rewetted coastal peatland. *Biogeosciences*, 16, 1937-1953.

775 Krauss, K. W., G. O. Holm, B. C. Perez, D. E. McWhorter, N. Cormier, R. F. Moss, D. J. Johnson, S. C.  
776 Neubauer & R. C. Raynie (2016) Component greenhouse gas fluxes and radiative balance from two  
777 deltaic marshes in Louisiana: Pairing chamber techniques and eddy covariance. *Journal of Geophysical*  
778 *Research-Biogeosciences*, 121, 1503-1521.

779 Kuhn, M. (2008) Building Predictive Models in R Using the caret Package. *Journal of Statistical*  
780 *Software*, 28, 1 - 26.

781 Lakomiec, P., J. Holst, T. Friberg, P. Crill, N. Rakos, N. Kljun, P. O. Olsson, L. Eklundh, A. Persson & J.  
782 Rinne (2021) Field-scale CH<sub>4</sub> emission at a subarctic mire with heterogeneous permafrost thaw status.  
783 *Biogeosciences*, 18, 5811-5830.

784 Levy, P. E., A. Burden, M. D. Cooper, K. J. Dinsmore, J. Drewer, C. Evans, D. Fowler, J. Gaiawyn, A.  
785 Gray & S. K. Jones (2012) Methane emissions from soils: synthesis and analysis of a large UK data set.  
786 *Global Change Biology*, 18, 1657-1669.

787 Li, C. (2007) Quantifying greenhouse gas emissions from soils: Scientific basis and modeling approach.  
788 *Soil Science and Plant Nutrition*, 53, 344-352.

789 Li, T. T., B. H. Xie, G. C. Wang, W. Zhang, Q. Zhang, T. Vesala & M. Raivonen (2016) Field-scale  
790 simulation of methane emissions from coastal wetlands in China using an improved version of  
791 CH<sub>4</sub>MOD(wetland). *Science of the Total Environment*, 559, 256-267.

792 Li, Y., D. Wang, Z. Chen, J. Chen, H. Hu & R. Wang (2021) Methane Emissions during the Tide Cycle  
793 of a Yangtze Estuary Salt Marsh. *Atmosphere*, 12, 245.

794 Liu, L. J., D. Q. Wang, S. Chen, Z. J. Yu, Y. K. Xu, Y. Li, Z. M. Ge & Z. L. Chen (2019) Methane  
795 Emissions from Estuarine Coastal Wetlands: Implications for Global Change Effect. *Soil Science Society*  
796 *of America Journal*, 83, 1368-1377.

797 Martin, R. M. & S. Moseman-Valtierra (2017) Different short-term responses of greenhouse gas fluxes  
798 from salt marsh mesocosms to simulated global change drivers. *Hydrobiologia*, 802, 71-83.

799 McKenna, T., J. Callahan, C. Medlock & N. Bates (2018) Creation of improved accuracy LiDAR-based  
800 digital elevation models for the St. Jones River and Blackbird Creek Watersheds. *Newark: Delaware*  
801 *Geologic Survey, Technical Report*, 1-30.

802 Mitra, B., K. Minick, G. Miao, J.-C. Domec, P. Prajapati, S. G. McNulty, G. Sun, J. S. King & A.  
803 Noormets (2020) Spectral evidence for substrate availability rather than environmental control of methane  
804 emissions from a coastal forested wetland. *Agricultural and Forest Meteorology*, 291, 108062.

805 Mønster, J., P. Kjeldsen & C. Scheutz (2019) Methodologies for measuring fugitive methane emissions  
806 from landfills – A review. *Waste Management*, 87, 835-859.

807 Moore, T. R. & N. T. Roulet (1993) Methane flux: Water table relations in northern wetlands.  
808 *Geophysical Research Letters*, 20, 587-590.

809 Morin, T. H. (2019) Advances in the Eddy Covariance Approach to CH<sub>4</sub> Monitoring Over Two and a  
810 Half Decades. *Journal of Geophysical Research-Biogeosciences*, 124, 453-460.

811 Morin, T. H., G. Bohrer, L. Naor-Azrieli, S. Mesi, W. T. Kenny, W. J. Mitsch & K. V. R. Schafer (2014)  
812 The seasonal and diurnal dynamics of methane flux at a created urban wetland. *Ecological Engineering*,  
813 72, 74-83.

814 Morin, T. H., G. Bohrer, K. C. Stefanik, A. C. Rey-Sanchez, A. M. Matheny & W. J. Mitsch (2017)  
815 Combining eddy-covariance and chamber measurements to determine the methane budget from a small,  
816 heterogeneous urban floodplain wetland park. *Agricultural and Forest Meteorology*, 237, 160-170.

817 Mulholland, M. M., & Otte, M. L. (2000) Effects of varying sulphate and nitrogen supply on DMSP and  
818 glycine betaine levels in *Spartina anglica*. *Journal of Sea Research*, 43(3-4), 199-207.

819 O'Connell, J. L., M. Alber & S. C. Pennings (2020) Microspatial Differences in Soil Temperature Cause  
820 Phenology Change on Par with Long-Term Climate Warming in Salt Marshes. *Ecosystems*, 23, 498-510.

821 Oikawa, P. Y., G. D. Jenerette, S. H. Knox, C. Sturtevant, J. Verfaillie, I. Dronova, C. M. Poindexter, E.  
822 Eichelmann & D. D. Baldocchi (2017) Evaluation of a hierarchy of models reveals importance of  
823 substrate limitation for predicting carbon dioxide and methane exchange in restored wetlands. *Journal of*  
824 *Geophysical Research-Biogeosciences*, 122, 145-167.

825 Paparella, F. (2005) Filling gaps in chaotic time series. *Physics Letters A*, 346, 47-53.

826 Pinheiro, J., D. Bates, S. DebRoy, D. Sarkar, S. Heisterkamp, B. Van Willigen & R. Maintainer (2017)  
827 Package 'nlme'. *Linear and nonlinear mixed effects models, version*, 3.

828 Poffenbarger, H. J., B. A. Needelman & J. P. Megonigal (2011) Salinity Influence on Methane Emissions  
829 from Tidal Marshes. *Wetlands*, 31, 831-842.

830 Reichstein, M., E. Falge, D. Baldocchi, D. Papale, M. Aubinet, P. Berbigier, C. Bernhofer, N. Buchmann,  
831 T. Gilmanov, A. Granier, T. Grünwald, K. Havránková, H. Ilvesniemi, D. Janous, A. Knohl, T. Laurila,  
832 A. Lohila, D. Loustau, G. Matteucci, T. Meyers, F. Miglietta, J.-M. Ourcival, J. Pumpanen, S. Rambal, E.  
833 Rotenberg, M. Sanz, J. Tenhunen, G. Seufert, F. Vaccari, T. Vesala, D. Yakir & R. Valentini (2005) On  
834 the separation of net ecosystem exchange into assimilation and ecosystem respiration: review and  
835 improved algorithm. *Global Change Biology*, 11, 1424-1439.

836 Reid, M. C., D. S. Pal & P. R. Jaffe (2015) Dissolved gas dynamics in wetland soils: Root-mediated gas  
837 transfer kinetics determined via push-pull tracer tests. *Water Resources Research*, 51, 7343-7357.

838 Reid, M. C., R. Tripathy, K. V. R. Schafer & P. R. Jaffe (2013) Tidal marsh methane dynamics:  
839 Difference in seasonal lags in emissions driven by storage in vegetated versus unvegetated sediments.  
840 *Journal of Geophysical Research-Biogeosciences*, 118, 1802-1813.

841 Rey-Sanchez, A. C., T. H. Morin, K. C. Stefanik, K. Wrigton & G. Bohrer (2018) Determining total  
842 emissions and environmental drivers of methane flux in a Lake Erie estuarine marsh. *Ecological*  
843 *Engineering*, 114, 7-15.

844 Rosentreter, J. A., A. N. Al-Haj, R. W. Fulweiler & P. Williamson (2021) Methane and Nitrous Oxide  
845 Emissions Complicate Coastal Blue Carbon Assessments. *Global Biogeochemical Cycles*, 35,  
846 e2020GB006858.

847 Schiecke, K., B. Pester, M. Feucht, L. Leistritz & H. Witte. 2015. Convergent Cross Mapping: Basic  
848 concept, influence of estimation parameters and practical application. In *2015 37th Annual International*  
849 *Conference of the IEEE Engineering in Medicine and Biology Society (EMBC)*, 7418-7421.

850 Seyfferth, A. L., F. Bothfeld, R. Vargas, J. W. Stuckey, J. Wang, K. Kearns, H. A. Michael, J. Guimond,  
851 X. Yu & D. L. Sparks (2020) Spatial and temporal heterogeneity of geochemical controls on carbon  
852 cycling in a tidal salt marsh. *Geochimica et Cosmochimica Acta*, 282, 1-18.

853 Stavert, A. R., M. Saunio, J. G. Canadell, B. Poulter, R. B. Jackson, P. Regnier, R. Lauerwald, P. A.  
854 Raymond, G. H. Allen, P. K. Patra, P. Bergamaschi, P. Bousquet, N. Chandra, P. Ciais, A. Gustafson, M.  
855 Ishizawa, A. Ito, T. Kleinen, S. Maksyutov, J. McNorton, J. R. Melton, J. Müller, Y. Niwa, S. Peng, W. J.  
856 Riley, A. Segers, H. Tian, A. Tsuruta, Y. Yin, Z. Zhang, B. Zheng & Q. Zhuang (2022) Regional trends  
857 and drivers of the global methane budget. *Global Change Biology*, 28, 182-200.

858 Strom, L., J. M. Falk, K. Skov, M. Jackowicz-Korczynski, M. Mastepanov, T. R. Christensen, M. Lund &  
859 N. M. Schmidt (2015) Controls of spatial and temporal variability in CH<sub>4</sub> flux in a high arctic fen over  
860 three years. *Biogeochemistry*, 125, 21-35.

861 Sugihara, G., R. May, H. Ye, C.-h. Hsieh, E. Deyle, M. Fogarty & S. Munch (2012) Detecting causality  
862 in complex ecosystems. *science*, 338, 496-500.

863 Sugihara, G. & R. M. May (1990) Nonlinear forecasting as a way of distinguishing chaos from  
864 measurement error in time series. *Nature*, 344, 734-741.

865 Sun, X., W. Fang, X. Gao, S. An, S. Liu & T. Wu (2021) Time-varying causality inference of different  
866 nickel markets based on the convergent cross mapping method. *Resources Policy*, 74, 102385.

867 Swain, C. K., P. Bhattacharyya, A. K. Nayak, N. R. Singh, S. Neogi, D. Chatterjee & H. Pathak (2018)  
868 Dynamics of net ecosystem methane exchanges on temporal scale in tropical lowland rice. *Atmospheric*  
869 *Environment*, 191, 291-301.

870 Takens, F. 1981. Detecting strange attractors in turbulence. In *Dynamical systems and turbulence*,  
871 *Warwick 1980*, 366-381. Springer.

872 Tokida, T., T. Miyazaki, M. Mizoguchi, O. Nagata, F. Takakai, A. Kagemoto & R. Hatano (2007) Falling  
873 atmospheric pressure as a trigger for methane ebullition from peatland. *Global Biogeochemical Cycles*,  
874 21.

875 Tong, C., W. Q. Wang, C. S. Zeng & R. Marrs (2010) Methane (CH<sub>4</sub>) emission from a tidal marsh in the  
876 Min River estuary, southeast China. *Journal of Environmental Science and Health Part a-*  
877 *Toxic/Hazardous Substances & Environmental Engineering*, 45, 506-516.

878 Tsonis, A. A., E. R. Deyle, H. Ye & G. Sugihara. 2018. Convergent Cross Mapping: Theory and an  
879 Example. In *Advances in Nonlinear Geosciences*, ed. A. A. Tsonis, 587-600. Cham: Springer  
880 International Publishing.

881 Turner, J., A. R. Desai, J. Thom & K. P. Wickland (2021) Lagged Wetland CH<sub>4</sub> Flux Response in a  
882 Historically Wet Year. *Journal of Geophysical Research: Biogeosciences*, 126, e2021JG006458.

883 Ushio, M. & K. Kawatsu. 2020. Forecasting Ecological Time Series Using Empirical Dynamic Modeling:  
884 A Tutorial for Simplex Projection and S-map. In *Diversity of Functional Traits and Interactions:  
885 Perspectives on Community Dynamics*, ed. A. Mougi, 193-213. Singapore: Springer Singapore.

886 Vázquez-Lule, A. & R. Vargas (2021) Biophysical drivers of net ecosystem and methane exchange across  
887 phenological phases in a tidal salt marsh. *Agricultural and Forest Meteorology*, 300, 108309.

888 Vázquez-Lule, A., A. L. Seyfferth, M. A. Limmer, P. Mey, M. Guevara, and R. Vargas. 2022.  
889 Hyperspectral reflectance for measuring canopy-level nutrients and photosynthesis in a salt marsh.  
890 *Journal of geophysical research. Biogeosciences*, 127.

891 Villa, J. A., Y. Ju, T. Stephen, C. Rey-Sanchez, K. C. Wrighton & G. Bohrer (2020) Plant-mediated  
892 methane transport in emergent and floating-leaved species of a temperate freshwater mineral-soil wetland.  
893 *Limnology and Oceanography*, 65, 1635-1650.

894 Villa, J. A., Y. Ju, C. Vines, C. Rey-Sanchez, T. H. Morin, K. C. Wrighton & G. Bohrer (2019)  
895 Relationships Between Methane and Carbon Dioxide Fluxes in a Temperate Cattail-Dominated  
896 Freshwater Wetland. *Journal of Geophysical Research-Biogeosciences*, 124, 2076-2089.

897 Wang, D. Q., Z. L. Chen & S. Y. Xu (2009) Methane emission from Yangtze estuarine wetland, China.  
898 *Journal of Geophysical Research-Biogeosciences*, 114.

899 Wang, Y., J. Yang, Y. Chen, P. De Maeyer, Z. Li & W. Duan (2018) Detecting the Causal Effect of Soil  
900 Moisture on Precipitation Using Convergent Cross Mapping. *Scientific Reports*, 8, 12171.

901 Wang, Z. P. & X. G. Han (2005) Diurnal variation in methane emissions in relation to plants and  
902 environmental variables in the Inner Mongolia marshes. *Atmospheric Environment*, 39, 6295-6305.

903 Ward, N. D., J. P. Megonigal, B. Bond-Lamberty, V. L. Bailey, D. Butman, E. A. Canuel, H.  
904 Diefenderfer, N. K. Ganju, M. A. Goñi, E. B. Graham, C. S. Hopkinson, T. Khangaonkar, J. A. Langley,  
905 N. G. McDowell, A. N. Myers-Pigg, R. B. Neumann, C. L. Osburn, R. M. Price, J. Rowland, A.  
906 Sengupta, M. Simard, P. E. Thornton, M. Tzortziou, R. Vargas, P. B. Weisenhorn & L. Windham-Myers  
907 (2020) Representing the function and sensitivity of coastal interfaces in Earth system models. *Nature  
908 Communications*, 11, 2458.

909 Wuebbles, D. J. & K. Hayhoe (2002) Atmospheric methane and global change. *Earth-Science Reviews*,  
910 57, 177-210.

911 Yang, B., X. Li, S. Lin, C. Jiang, L. Xue, J. Wang, X. Liu, M. Espenberg, J. Pärn & Ü. Mander (2021)  
912 Invasive *Spartina alterniflora* changes the Yangtze Estuary salt marsh from CH<sub>4</sub> sink to source.  
913 *Estuarine, Coastal and Shelf Science*, 252, 107258.

914 Ye, A. Clark, E. Deyle & G. Sugihara (2016) rEDM: an R package for empirical dynamic modeling and  
915 convergent cross-mapping. *cran. r-project. org*.

916 Ye, H., R. J. Beamish, S. M. Glaser, S. C. H. Grant, C.-h. Hsieh, L. J. Richards, J. T. Schnute & G.  
917 Sugihara (2015a) Equation-free mechanistic ecosystem forecasting using empirical dynamic modeling.  
918 *Proceedings of the National Academy of Sciences*, 112, E1569-E1576.



919 Ye, H., E. R. Deyle, L. J. Gilarranz & G. Sugihara (2015b) Distinguishing time-delayed causal  
920 interactions using convergent cross mapping. *Scientific Reports*, 5, 14750.

921 Yuan, J. J., W. X. Ding, D. Y. Liu, H. Kang, J. Xiang & Y. X. Lin (2016) Shifts in methanogen  
922 community structure and function across a coastal marsh transect: effects of exotic *Spartina alterniflora*  
923 invasion. *Scientific Reports*, 6.

924 Yuan, J. J., D. Y. Liu, Y. Ji, J. Xiang, Y. X. Lin, M. Wu & W. X. Ding (2019) *Spartina alterniflora*  
925 invasion drastically increases methane production potential by shifting methanogenesis from  
926 hydrogenotrophic to methylotrophic pathway in a coastal marsh. *Journal of Ecology*, 107, 2436-2450.

927 Yvon-Durocher, G., A. P. Allen, D. Bastviken, R. Conrad, C. Gudas, A. St-Pierre, N. Thanh-Duc & P.  
928 A. del Giorgio (2014) Methane fluxes show consistent temperature dependence across microbial to  
929 ecosystem scales. *Nature*, 507, 488-491.

930 Zaki, M. T. & O. I. Abdul-Aziz (2022) Predicting greenhouse gas fluxes in coastal salt marshes using  
931 artificial neural networks. *Wetlands*, 42.

932 Zhang, Q., L. Ting-Ting, Q. Zhang, W. Guo-Cheng, Y. Li-Jun, G. Bin & H. Peng-Fei (2020) Accuracy  
933 analysis in CH4MODwetland in the simulation of CH4 emissions from Chinese wetlands. *Advances in*  
934 *Climate Change Research*, 11, 52-59.

935 Zhang, Z., N. E. Zimmermann, A. Stenke, X. Li, E. L. Hodson, G. F. Zhu, C. L. Huang & B. Poulter  
936 (2017) Emerging role of wetland methane emissions in driving 21st century climate change. *Proceedings*  
937 *of the National Academy of Sciences of the United States of America*, 114, 9647-9652.

938 Zhao, M. L., G. X. Han, J. Y. Li, W. M. Song, W. D. Qu, F. Eller, J. P. Wang & C. S. Jiang (2020)  
939 Responses of soil CO2 and CH4 emissions to changing water table level in a coastal wetland. *Journal of*  
940 *Cleaner Production*, 269.

941 Zhu, X. D., L. X. Meng, Y. H. Zhang, Q. H. Weng & J. Morris (2019) Tidal and Meteorological  
942 Influences on the Growth of Invasive *Spartina alterniflora*: Evidence from UAV Remote Sensing. *Remote*  
943 *Sensing*, 11.

944 Zhu, X. D., Q. L. Zhuang, Z. C. Qin, M. Glagolev & L. L. Song (2013) Estimating wetland methane  
945 emissions from the northern high latitudes from 1990 to 2009 using artificial neural networks. *Global*  
946 *Biogeochemical Cycles*, 27, 592-604.