

1 **Anthropogenic dust as a significant source of ice-nucleating particles in the urban**
2 **environment**

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14 **Key Points:**

- 15 • The coarse ice nucleating particles (INPs) contribute ~95.2% of the total INPs in the
16 urban atmosphere.
- 17 • Anthropogenic dust such as traffic-influenced road dust proved to be a major source of
18 heat-resistant INPs in the urban environment.
- 19 • Coarse particles are strongly correlated to anthropogenic dust INPs and can be used to
20 predict their concentration.
21

22 **Abstract**

23 Anthropogenic dust is an important constituent of airborne particles in the urban environment
24 but its ice nucleation activity remains uninvestigated. Here, we studied the sources and ice
25 nucleating properties of size-resolved particles in the urban atmosphere under mixed-phase
26 cloud conditions. The heat-resistant ice nucleating particles (INPs) unexpectedly contributed
27 ~70% of the INPs in coarse mode at temperatures below $-15\text{ }^{\circ}\text{C}$. Detailed size-resolved particle
28 chemical composition analysis showed that these INPs were contributed by anthropogenic dust,
29 such as traffic-influenced road dust. A parameterization based on coarse particles was
30 developed to predict the anthropogenic dust INP concentration, due to their correlations on
31 concentration and similarity in chemical compositions. The parameterization can be used for
32 further evaluating the anthropogenic dust contribution to INPs on a global scale. We suggest
33 anthropogenic dust associated with rapid urbanization will become an important factor for
34 urban climate change by altering the cloud microphysics.

35 **Plain Language Summary**

36 Anthropogenic dust (dust particles generated by human activities) prevails in the urban
37 atmosphere, but its ability to nucleate ice is not well understood. Combing the chemical
38 composition analysis and ice-nucleating particle measurements of urban aerosols, we found
39 anthropogenic dust, such as traffic-influenced road dust, is an important source of INPs in the
40 urban atmosphere. On basis of the number concentration of coarse particles and INPs, a new
41 parameterization is developed to predict the INP concentration contributed by anthropogenic
42 dust. We propose that anthropogenic dust will become more important for the future urban
43 climate by serving as INPs associated with the increasing anthropogenic dust loading caused
44 by rapid urbanization.

45 **1 Introduction**

46 The heterogeneous ice nucleation process aided by ice nucleating particles (INPs) are
47 important for ice crystal formation in clouds and is still an unsolved issue in aerosol-cloud
48 interaction [Fan *et al.*, 2016]. The INP concentrations and their ice nucleation activities in the
49 atmosphere would regulate the microphysics and radiative properties of clouds by influencing
50 the number concentration and size of ice crystals, as a result, indirectly influencing the global
51 climate (IPCC report 2021) [Legg, 2021]. A fundamental understanding of the abundance and
52 sources of atmospheric INPs is still not achieved due to the rarity and complexity of INPs [Kanji
53 *et al.*, 2017]. In the urban atmosphere, aerosols originate from both natural and anthropogenic
54 emissions and undergo a number of aging and transformation processes. Such complexity adds
55 to the difficulties to identify the major sources of INPs and quantify the respective
56 contributions. This can be a missing piece for the cloud and climate model to predict the aerosol-
57 ice cloud interaction in urban regions and its following impact on the urban-climate system.

58 The ice nucleation activities of the bulk urban aerosol and its constituents have been
59 studied under mixed-phase cloud conditions [Bi *et al.*, 2019; Chen *et al.*, 2018a; Chen *et al.*,
60 2021b; Hasenkopf *et al.*, 2016; Pereira *et al.*, 2021; Yadav *et al.*, 2019; Zhang *et al.*, 2022]. The
61 biological INPs prevail in the urban atmosphere and are considered efficient INPs at
62 temperatures above $-15\text{ }^{\circ}\text{C}$ [Pereira *et al.*, 2021; Yadav *et al.*, 2019]. The INP concentrations
63 can be extremely enhanced when the urban atmosphere is influenced by the long-range
64 transport dust plume originating from desert regions [Bi *et al.*, 2019; Chen *et al.*, 2021a]. A
65 recent study by Tian *et al.* [2022] shows the contribution of organic aerosols (heat-resistant at
66 $350\text{ }^{\circ}\text{C}$) to immersion-freezing INPs at $-30\text{ }^{\circ}\text{C}$ in the urban region. However, this study was
67 unable to exclude or evaluate the impact of local dust emissions on these refractory INPs, which
68 is a big contributor to aerosols in urban regions [Han *et al.*, 2005]. The poor ice nucleation
69 performance of black carbon (BC) or soot particles from fossil fuel combustion under mixed-

70 phase cloud conditions has been verified by both laboratory studies [Kanji *et al.*, 2020; Vergara-
71 *Temprado et al.*, 2018] and field measurements [Chen *et al.*, 2018; Zhang *et al.*, 2022]. The
72 INP concentrations did not show an obvious increase or decrease during heavily polluted
73 periods in Beijing, indicating the secondary aerosols formed from the increasingly emitted
74 anthropogenic precursors were not a major source of the observed INP concentrations [Bi *et al.*,
75 2019; Chen *et al.*, 2018; Zhang *et al.*, 2022]. This conclusion is also confirmed by many field
76 studies on a global scale either conducted in highly polluted regions [Hasenkopf *et al.*, 2016;
77 *Pereira et al.*, 2021; Yadav *et al.*, 2019] or in regions occasionally influenced by air pollution
78 [Creamean *et al.*, 2018; Wex *et al.*, 2015], where a negligible effect on INP production from
79 the non-dust air pollution was observed.

80 These studies imply that the INPs in the urban atmosphere are unlikely contributed by
81 anthropogenic pollutants under mixed-phase cloud conditions as far as dust air pollution is not
82 considered. Instead, the natural biological particles and the long-range transported desert dust
83 are still considered common INP sources and there may exist unidentified INP sources in non-
84 dust days. However, note that the atmospheric dust loading can also be changed
85 anthropogenically due to land use changes by human activity, the so-called anthropogenic dust
86 (as defined in Zender *et al.* [2004]). Anthropogenic dust accounts for ~30 to 70% of total dust
87 concentrations in urban in recent decades [Chen *et al.*, 2018; Huang *et al.*, 2015]. On a global
88 scale, anthropogenic contribution to atmospheric dust loads today is between ~90 and 2000 Mt.
89 year⁻¹ [Webb and Pierre, 2018]. It is an important constituent of airborne particles in the urban
90 atmosphere and can be emitted by construction works, traffic-generated turbulence, and
91 agricultural and industrial activities [Haynes *et al.*, 2020; Philip *et al.*, 2017]. The representative
92 anthropogenic dust includes traffic-influenced road dust [Xia *et al.*, 2022] and soil particles
93 from disturbed soil [Wang *et al.*, 2018].

94 Despite the high mass loading of anthropogenic dust in the urban atmosphere and its
95 important direct and indirect effects on the urban climate system [Philip *et al.*, 2017; Xia *et al.*,
96 2022], limited studies have investigated the ice nucleation properties of this dust species
97 compared to those of other pollutants. Many studies investigated the ice nucleation activities of
98 ground-based collected soil dust. They showed that soil dust can serve as INPs in a wide
99 temperature range (-35 to -6 °C) [Hill *et al.*, 2016; O'Sullivan *et al.*, 2014; Pereira *et al.*, 2022;
100 *Steinke et al.*, 2016; *Tobo et al.*, 2014] and its ice nucleation activity is constrained by the
101 containing biological compounds [Conen *et al.*, 2011] and the organic matters (OM) [Pereira
102 *et al.*, 2022; *Tobo et al.*, 2014]. The ice nucleation activity of anthropogenic soil dust from the
103 emission of disturbed soils is less studied. Studies pointed out that the atmospheric INPs in
104 South America can come from airborne agricultural dust [Gong *et al.*, 2022; Testa *et al.*, 2021].
105 *Corbin et al.* [2012] showed dust particles were riched in the ice residues activated by INPs (at
106 a temperature of -34 °C and relative humidity of 95% with respect to water) in Toronto. The
107 major source of the detected dust is likely vehicular resuspension from nearby roads, implying
108 the potential contribution of road dust to INPs [Corbin *et al.*, 2012].

109 In the present study, the size-resolved ambient particles were collected in the urban
110 atmosphere when no impacts of natural desert dust were observed. The ice nucleation activities
111 and the chemical compositions of the collected particles were investigated and evidence that
112 anthropogenic dust makes a significant contribution to INPs in the urban environment.
113 Regarding the continuous rapid increase of urbanization in the future, we suggest that
114 anthropogenic dust emissions due to off-road vehicles and urban construction may play an
115 important role in affecting the global INP budget.

116 2 Materials and Methods

117 2.1 Aerosol Sampling and Characterization

118 The aerosol samples were collected at Peking University Urban Atmospheric
 119 Environment Monitoring Station (PKUERS) in the summer from June 22 to July 21, 2020.
 120 PKUERS is located on the campus of Peking University and is 20 meters above the ground.
 121 This site is a representative urban site affected by multiple anthropogenic emissions, such as
 122 transport emissions and fossil fuel combustion [Chen *et al.*, 2018a; Zhang *et al.*, 2022].

123 Aerosols were collected onto the 47 mm diameter polycarbonate filters (Whatman,
 124 111107) by a Micro-Orifice Uniform Deposit Impactor (MOUDI, MSP Corporation, USA) with
 125 a flow rate of 30 L min⁻¹. MOUDI allows the particles to be classified and collected with
 126 different aerodynamic diameters (AD). Each sample set (i.e., each set including filter samples
 127 collected particles with different cut-off sizes) was collected for 24 hours, and the detailed
 128 sampling information was listed in Table S1. Particles with the cut-off size of 0.56 μm, 1.0 μm,
 129 1.8 μm, 3.2 μm, and 5.6 μm were collected for each set, an aerosol population that is of interest
 130 in the INP parameterizations. All reported sizes in the present study are the 50% cut-off AD
 131 (D_{50}), which corresponds to the AD of particles trapped with an efficiency of 50% at a given
 132 stage. In total, 8 sets of filter samples were collected.

133 An aerodynamic particle sizer (APS, model 3021, TSI) measured the number
 134 concentration of particles with AD ranging from 0.542 μm to 19.81 μm ($N_{>*}$ is defined as the
 135 number concentration of particles larger than * μm, N_* is the number concentration of particle
 136 with the size of * μm). Note that APS and MOUDI both detect the particle aerodynamic
 137 diameter. Elements of samples include Na, Mg, Al, K, Ca, Mn, Fe, Zn, As, Ba, Pb, and others
 138 (V, Co, Se, Sr, Mo, Tl, Bi, Th, U, Cd, Ni, Cu, Ti, Cr, and P) were measured by an Inductively
 139 Coupled Plasma-Mass Spectrometry (ICP-MS, Bruker, aurora M90). The mass concentrations
 140 of the organic carbon (OC) and element carbon (EC) were measured by the Sunset ECOC
 141 analyzer (Sunset Lab 4) using quartz filters collected parallelly by the aerosol sampler with
 142 PM_{2.5} cut-off size.

143 2.2 Ice Nucleation Measurement

144 The ice nucleation measurement was performed by the Peking University Ice
 145 Nucleation Array (PKU-INA), which is a cold-stage-based device to measure the freezing
 146 ability of droplets under mixed-phase cloud conditions [Chen *et al.*, 2018b]. In this study, each
 147 filter sample was first immersed in distilled water of 7 mL and shaken by the vortex for 40
 148 minutes to wash particles off. The resulting suspension was then pipetted onto the cold stage to
 149 form 90 droplets with a volume of 1 μL. Droplets were separated by a spacer and then sealed
 150 by a cover glass to avoid the Wegener-Bergeron-Findeisen process. Droplets were cooled down
 151 to -32 °C with a cooling rate of 1 °C/min. Meanwhile, the status of all the droplets was recorded
 152 every 6 seconds by a high-speed camera (Q-imaging MicroPublisher 5.0 RTV, QImaging,
 153 Surrey, BC, Canada) mounted on the top of the cold stage. The obtained images were then input
 154 into the developed MATLAB program to identify the freezing temperatures of the droplets
 155 according to the brightness change of each droplet upon its phase transition.

156 The frozen fraction (f_{ice}) of the droplets at each temperature can be obtained using Eq.
 157 (1) by assuming a time-independent ice nucleation process of droplets:

$$158 \quad f_{ice}(T) = \frac{N_f}{N_t} \quad (1)$$

159 where N_f is the number of frozen droplets at a given temperature and N_t is the total
 160 number of droplets (i.e., 90 in this study).

161 The number concentration of INP (N_{INP}) per unit volume of sampled air collected on
 162 each sample filter (i.e., at each cut-off size) is calculated based on the f_{ice} and the total volume
 163 of the sampled air that was collected into each droplet (V_{air}) using Eq. (2):

$$164 \quad N_{\text{INP}}(T) = \frac{-\ln(1 - f_{\text{ice}}(T))}{V_{\text{air}}} \quad (2)$$

165 and the total number concentration of INPs each day was derived by integrating the INP number
 166 concentration with different cut-off sizes.

167 The cumulative number concentration of the ice active sites per unit surface area of
 168 INPs (n_s), as derived in many studies to describe the ice nucleation ability of particles [Connolly
 169 *et al.*, 2009; Hiranuma *et al.*, 2015; Niemand *et al.*, 2012], is calculated according to Eq. (3):

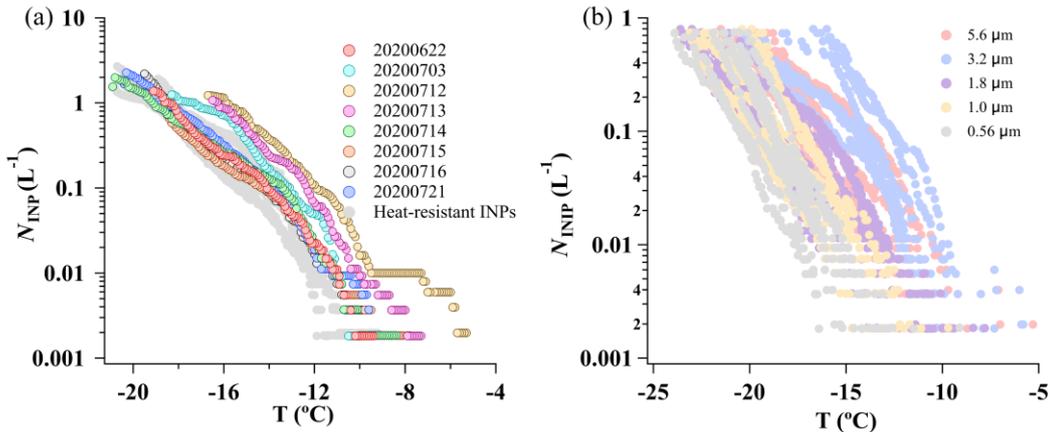
$$170 \quad n_s(T) = \frac{N_{\text{INP}}(T)}{A} \quad (3)$$

171 For particles collected at each cut-off size, A ($A = \frac{\pi N_* D_p^2}{4}$) is the total surface area of the
 172 particles per unit volume of air sampled at each stage. A is calculated based on the number
 173 concentration of particles per unit volume of air with different aerodynamic sizes (D_p) measured
 174 by APS and assuming the particles are spherical.

175 3 Results and Discussion

176 3.1 Overview of the INP Concentration

177 Figure 1(a) shows the total number concentration of INPs (N_{INP}) (the sum of size-
 178 resolved N_{INP}) detected on different days as a function of temperatures. The temperature
 179 dependencies of total N_{INP} are similar and the variations of N_{INP} are less than one magnitude
 180 from -21 °C to -5 °C. This implies the INPs on different days may originate from similar
 181 aerosol sources and no specific strong sources of INPs presented on one particular day during
 182 the sampling period.



183
 184 **Figure 1.** The total (a) and size-dependent (b) number concentration of INPs (N_{INP}) as
 185 a function of temperatures; The gray circles in (a) represent the number concentration of heat-
 186 resistant INPs after being heated at 95 °C for 20 mins. The cut-off size in (b) corresponds to the
 187 aerodynamic diameter of particles trapped with an efficiency of 50% at a given stage.

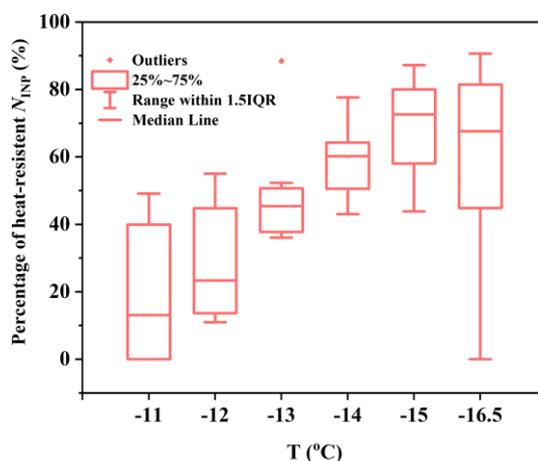
188 The size-dependent N_{INP} (with $D_{50}=0.56$ μm, 1 μm, 1.8 μm, 3.2 μm, and $D_{50}=5.6$ μm)
 189 is shown in Fig. 1(b). Each line represents the N_{INP} with one cut-off size collected in one day.
 190 On average, the N_{INP} in size larger than 1 μm explained $95.2\% \pm 4\%$ of the total N_{INP} over the

191 temperature range from $-20\text{ }^{\circ}\text{C}$ to $-10\text{ }^{\circ}\text{C}$, meaning that coarse INPs dominate the INP
 192 concentration in the urban atmosphere of Beijing. The prevalence and importance of the coarse
 193 INPs are pointed out by other studies conducted in the global atmosphere as well [Mason *et al.*,
 194 2016; Gong *et al.*, 2020; Mitts *et al.*, 2021]. In the urban atmosphere, the coarse particles can
 195 originate from dust particles, tire debris, and biological particles (see Wu and Boor [2021] and
 196 the references therein). The concentration of coarse particles has a strong correlation with the
 197 INP concentrations detected in the urban environment, thus can be an important source of INPs
 198 [Che *et al.*, 2021; Chen *et al.*, 2021; Jiang *et al.*, 2023]. The sources of the detected coarse
 199 particles and INPs will be discussed in the following sections.

200 3.2 The Contribution of Anthropogenic Dust to INPs

201 The extracted sample solutions were heated to $95\text{ }^{\circ}\text{C}$ for 20 mins. The total N_{INP}
 202 decreased after heat treatment (Fig. 1), indicating that proteinaceous biological INPs
 203 contributed to the observed INPs (refer to [Christner *et al.*, 2008]). Fig. 2 furtherly shows the
 204 percentage of heat-resistant N_{INP} (the ratio of N_{INP} measured after and before heat treatment) at
 205 different temperatures. The contribution of heat-sensitive proteinaceous biological particles to
 206 INPs only becomes important at temperatures above $-14.3\text{ }^{\circ}\text{C}$ ($\sim 50\%$). The percentage of heat-
 207 resistant N_{INP} increases with decreasing temperatures. At temperatures below $-15\text{ }^{\circ}\text{C}$, the heat-
 208 resistant INPs account for $\sim 70\%$ of the total INPs.

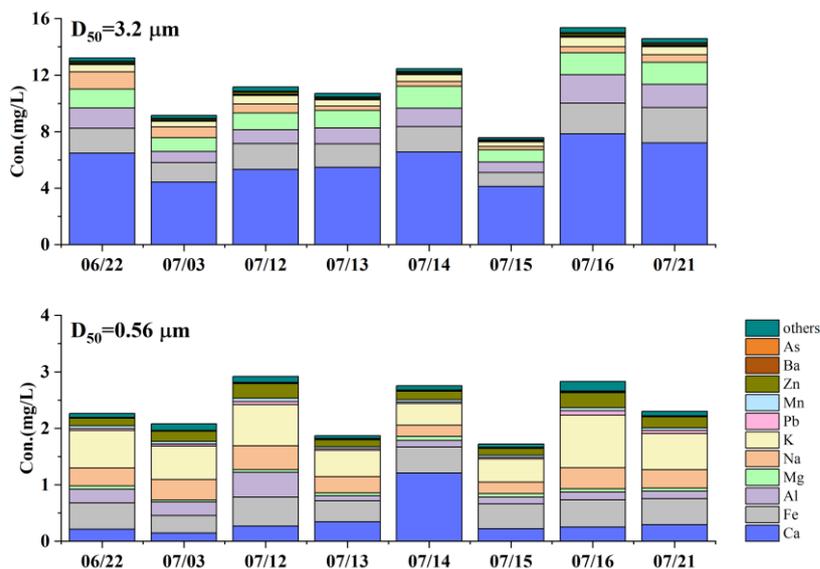
209 The heat-resistant INPs can originate from dust particles (refer to [Hill *et al.*, 2016;
 210 O'Sullivan *et al.*, 2014; Perkins *et al.*, 2019]) and some unidentified organic matter (OM)
 211 [McCluskey *et al.*, 2018; Tian *et al.*, 2022]. The impacts of soot particles and other inorganic
 212 components (inorganic salts) on the heat-resistant INPs are excluded due to their poor ice
 213 nucleation activities under mixed-phase cloud conditions [Chen *et al.*, 2018a; Kanji *et al.*, 2020;
 214 Zhang *et al.*, 2022].



215
 216 **Figure 2.** The percentages of heat-resistant N_{INP} vs. temperatures; only values at
 217 temperatures above $-16.5\text{ }^{\circ}\text{C}$ were calculated where droplets in all samples did not finish frozen.

218 The source of aerosols and heat-resistant INPs was explored based on the chemical
 219 analysis of collected particles. The contribution of heat-resistant INPs from OM is considered
 220 to be minor under the determined conditions, as a poor correlation between the detected heat-
 221 resistant INPs and organic carbon was found ($R^2=0.19$, Fig. S1). The chemical elements of the
 222 coarse particles ($D_{50}=3.2\text{ }\mu\text{m}$) (Fig. 3 and Table S3) show that crustal elements including Ca
 223 ($50.46\%\pm 2.27\%$), Mg ($10.98\%\pm 0.82\%$), Fe ($14.89\%\pm 1.48\%$) and Al ($10.42\%\pm 1.45\%$) account
 224 for a major mass percentage ($88.64\%\pm 2.92\%$) of the total element mass of particles,
 225 demonstrating the strong contribution from dust particles. Differently, the crustal elements in

226 fine particles ($D_{50}=0.56\ \mu\text{m}$) are much less, while a large mass percentage of non-dust related
 227 elements (Zn, Pb, K, and Na) are found ($48.53\%\pm 10.61\%$) (Fig. 3 and Table S3). This indicates
 228 that dust particles contribute to heat-resistant INPs in the coarse particles, while fine particles
 229 may be influenced by non-dust anthropogenic sources (see detailed discussion in Fig. S2 and
 230 Text S1). This conclusion is also supported by the different ice nucleation activities (presented
 231 as n_s , calculated based on Eq.(3)) obtained for coarse and fine INPs (Fig. S3), which comes
 232 from different aerosol sources.



233

234

Figure 3. The element compositions of particles in different sizes.

235 Dust particles in the urban atmosphere can either be long-range transported desert dust
 236 or anthropogenic dust [Han *et al.*, 2005; Zender *et al.*, 2004]. Here, we confirmed natural desert
 237 dust has negligible impact on the heat-resistant INPs while anthropogenic dust is a major
 238 source, as supported by the following evidence. No obvious enhancement in the number
 239 concentration of coarse particles and in n_s has been observed during the sampling time (Fig.
 240 S4(a)), which typically occurred during Asian dust storm events (Fig. S4(b)) [Chen *et al.*,
 241 2021a]. The n_s obtained in this study ($3*10^3$ to $2*10^5\ \text{m}^{-2}$ at $-15\ ^\circ\text{C}$) were two magnitudes lower
 242 than those obtained during the spring Asian dust events (10^5 to $10^7\ \text{m}^{-2}$ at $-15\ ^\circ\text{C}$) at the same
 243 site [Chen *et al.*, 2021a]. The analysis of 72-hours backward trajectories (calculated every 6
 244 hours at 1:00, 7:00, 13:00, and 19:00 UTC at the height of 500 m) (Fig. S5) shows that air
 245 masses did not pass over the desert regions. Asian dust events occur especially during spring
 246 months and not in the summer of Beijing [Shao and Dong, 2006], while the anthropogenic dust
 247 from local emissions can contribute $\sim 80\%$ of the total dust [Han *et al.*, 2005].

248 The impact of one representative anthropogenic dust species on the collected aerosols
 249 and INPs, the traffic-influenced road dust, is proved by the good correlation between the hour-
 250 mean concentration of the coarse particles and nitric oxide (NO) during the sampling time
 251 ($R^2=0.53$, Fig. S6). Coarse particles and NO are indicators of the dust particles and vehicle
 252 primary emissions, respectively. Good correlations between heat-resistant N_{INP} at $-16\ ^\circ\text{C}$ and
 253 the mass percentage of Ba and Zn were observed in coarse particles ($R^2=0.39$ and $R^2=0.60$, Fig.
 254 S7). Ba and Zn are two tracers for the road dust particles [Gietl *et al.*, 2010; Harrison and
 255 Alghamdi, 2023; Peltier *et al.*, 2011], which proves the contribution of road dust to heat-
 256 resistant INPs again. Other anthropogenic dust species, such as those generated from
 257 agricultural and construction activities may contribute to the collected coarse particles as well
 258 but cannot be validated in the present study due to lacking reliable tracers. Since the dust
 259 generated from construction is commonly found in urban regions [Azarov *et al.*, 2019; Yan *et*
 260 *al.*, 2020], we expect its contribution to the observed anthropogenic dust INPs. The impact of

261 soil dust from agricultural activities is hard to be evaluated here. However, the n_s values of the
 262 obtained heat-resistant INPs are 3 orders of magnitudes lower than those from soil dust (original
 263 or H₂O₂ treated soil dust) measured by *Tobo et al.* [2014] (Fig. S8) and than the n_s of inorganic
 264 INPs measured by *Testa et al.* [2021] which was presumably from land surface emission. This
 265 comparison implies that the ice nucleation activity of the airborne anthropogenic dust cannot
 266 be explained alone by the collected ground-based and near-source soil dust.

267 3.3 Parameterization of Anthropogenic Dust INPs

268 Based on the aforementioned results, the coarse heat-resistant N_{INP} can be referred to as
 269 anthropogenic dust N_{INP} in the urban atmosphere. The anthropogenic N_{INP} is highly correlated
 270 to the number concentration of coarse particles ($N_{>1\ \mu\text{m}}$, $R^2=0.67$, Table S2), which can be
 271 foreseen from the strong impact of dust on the coarse particles (section 3.2). Thus, $N_{>1\ \mu\text{m}}$ is
 272 used to predict the concentration of anthropogenic dust INPs. For comparison, the correlations
 273 between the anthropogenic N_{INP} and the number concentration of particles with other size
 274 ranges ($N_{>500\ \text{nm}}$ and $N_{>1.8\ \mu\text{m}}$) are listed in Table S2.

275 **Table 1** The parameterizations to predict the anthropogenic dust INPs

| | $N_{\text{INP}} = (N_{>1\ \mu\text{m}})^{(\alpha(-T)+\beta)} * \exp(\gamma(-T) + \delta)$ | | | | | |
|-------------------------|-------------------------------------------------------------------------------------------|---------|------------|----------|-------------|-------|
| | α | β | γ | δ | R^2 | |
| Anthropogenic dust INPs | - 1.835*10 ⁻⁵ | 4 | 0.029 2 | 0.653 | -12.51 8 | 0.927 |

276 The parametrization to predict the anthropogenic dust INPs at a temperature range from
 277 $-21^\circ\text{C} \sim -7^\circ\text{C}$ was developed based on $N_{>1\ \mu\text{m}}$ following the form of *DeMott et al.* [2015]
 278 parameterization (hereafter D15) (listed in Table 1). The coefficient of determination (R^2) of
 279 the resulting parameterization is 0.95 with 95% confidence. The comparison between the results
 280 from observation and prediction was shown in Fig. S9. The ratios of the observed to the
 281 predicted values are within a factor of 3, indicating a good prediction from the parameterization.
 282 To be compared, the heat-resistant N_{INP} was estimated using D15 as well, which was developed
 283 based on $N_{>500\ \text{nm}}$ (Fig. S10). An overestimation of heat-resistant/total INPs about 1~2
 284 magnitudes was observed, indicating poor prediction. This is because the particles larger than
 285 500 nm cannot represent the exact size range of the measured INP species, the anthropogenic
 286 dust INPs here (evidenced by the different chemical compositions and ice nucleation activities
 287 of coarse and fine particles). The overestimation of N_{INP} in the urban environment by D15 was
 288 confirmed by *Chen et al.* [2018a] and *Bi et al.* [2019] as well.

289 Overall, the parameterization developed based on $N_{>1\ \mu\text{m}}$ showed a good performance
 290 on the prediction of the anthropogenic dust INPs. This parameterization can then be used in
 291 regional or climate models to predict the INPs contributed by anthropogenic dust and evaluate
 292 its further impact on cloud formation and the urban climate. Note that using heat-resistant N_{INP}
 293 as the proxy of anthropogenic dust N_{INP} in the urban atmosphere can cause a bias. This
 294 assumption cannot exclude that some of the measured heat-resistant INPs are from the heat-
 295 resistant OM [*Hill et al.*, 2016; *O'Sullivan et al.*, 2014; *Tobo et al.*, 2014], although a poor
 296 correlation between OM and heat-resistant INPs and significant dust influence has been found
 297 here. On the other hand, some dust species can significantly lose their ice nucleation activity
 298 when undergoing the heating process [*Daily et al.*, 2020], indicating the possibility of
 299 underestimating the anthropogenic dust N_{INP} based on the heat-lability of particles. These
 300 uncertainties might influence the quantification of anthropogenic dust INPs and need more
 301 investigation on measuring the contribution of anthropogenic dust to aerosols and atmospheric
 302 INPs in the urban environment.

303 **4 Conclusions**

304 The abundance and sources of INPs remain unclear in the urban atmosphere. In the
305 present study, the ice nucleation activities of size-resolved particles were investigated in a
306 typical urban environment. INPs in coarse-mode account for $95.2\% \pm 4\%$ of the detected N_{INP}
307 and are significantly contributed by heat-resistant INPs ($\sim 70\%$) at temperatures below -15°C .
308 The further size-resolved chemical composition analysis indicates that anthropogenic dust such
309 as traffic-influenced road dust is a major source of these heat-resistant INPs. The number
310 concentration of coarse particles is strongly correlated with those of the anthropogenic dust
311 INPs, thus was used to develop the parameterization for predicting anthropogenic dust INP
312 concentrations. The new parameterization predicts the observed N_{INP} within a factor of 3,
313 showing a good performance. This study highlighted the importance of anthropogenic dust
314 particles as a significant source of INPs in urban environments, which has not been extensively
315 studied before.

316 Projections of past and future atmospheric dust loads suggest that the contribution of
317 anthropogenic emissions is increased in the last decades and it will continue to have a profound
318 role in the Earth's future climate [Mahowald and Luo, 2003; Stanelle *et al.*, 2014]. Due to the
319 good ice nucleation activity of the anthropogenic dust and its large contributions to the
320 atmospheric INPs reported by the present study, we suggest that anthropogenic dust will
321 become an important factor to connect with urban climate change by modifying INP
322 concentrations and cloud microphysics.

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325 Science Foundation of China and the Swedish Foundation for International Cooperation
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327

328 **Conflict of interest**

329 The authors declare no conflicts of interest relevant to this study.
330

331 **Open Research**

332 The data that support the findings of this study are available at this site:
333 <https://doi.org/10.5281/zenodo.7788005>.
334

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