

1 **Configuration and validation of an oceanic physical and**
2 **biogeochemical model to investigate coastal**
3 **eutrophication in the Southern California Bight**

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11 **Key Points:**

- 12 • The model reproduces the gradients of physical and biogeochemical properties that
13 can be traced to the influence of coastal inputs.
14 • The model reproduces nutrient enrichment via subsurface wastewater outfall plumes
15 and riverine runoff.
16 • The model has been developed to disentangle natural variability, climate change, and
17 local anthropogenic forcing.

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Abstract

The Southern California Bight (SCB), an eastern boundary upwelling system, is impacted by global warming, acidification and oxygen loss, and receives anthropogenic nutrients from a coastal population of 20 million people. We describe the configuration, forcing, and validation of a realistic, submesoscale resolving ocean model as a tool to investigate coastal eutrophication. This modeling system represents an important achievement because it strikes a balance of capturing the forcing by U.S. Pacific Coast-wide phenomena, while representing the bathymetric features and submesoscale circulation that affect the transport of nutrients from natural and human sources. Moreover, the model allows to run simulations at timescales that approach the interannual frequencies of ocean variability, making the grand challenge of disentangling natural variability, climate change, and local anthropogenic forcing a tractable task in the near-term. The model simulation is evaluated against a broad suite of observational data throughout the SCB, showing realistic depiction of the mean state and its variability with satellite and *in situ* measurements of state variables and biogeochemical rates. The simulation reproduces the main structure of the seasonal upwelling front, the mean current patterns, the dispersion of wastewater plumes, as well as their seasonal variability. Furthermore, it reproduces the mean distributions of key biogeochemical and ecosystem properties and their variability. Biogeochemical rates reproduced by the model, such as primary production and nitrification, are also consistent with measured rates. Results of this validation exercise demonstrate the utility of fine-scale resolution modeling in support of management decisions on local anthropogenic nutrient discharges to coastal zones.

Plain Language Summary

We applied and validated an ocean numerical model to investigate the effects of land-based and atmospheric nutrient loading on coastal eutrophication and its effects on carbon, nitrogen and oxygen cycles of the Southern California Bight, an upwelling-dominated marine embayment on the U.S. West Coast. The model is capable of high resolution, multi-year hindcast simulations, which enable investigations to disentangle natural variability, climate change, and local human pressures that accelerate land-based and atmospheric nutrient loads. The model performance assessment illustrates that it faithfully reproduces monitored ocean properties related to algal blooms, oxygen and water acidity, among others, that can be traced to land-based and atmospheric inputs of nutrients and carbon from human activities. The model performance assessment helps to constrain uncertainties in predictions to support ongoing conversations on approaches to reduce the effects of climate change, including considerations of management of local nutrient and carbon inputs.

1 Introduction

Human-driven eutrophication has resulted in profound impacts to coastal ecosystems around the world. These impacts are arguably the best studied in estuaries and enclosed bays (e.g. Chesapeake Bay; Cerco and Cole (1993); Boesch et al. (2001)) and semi-enclosed seas such as the Baltic Sea (Savchuk & Wulff, 2007; Cederwall & Elmgren, 1990), the Mediterranean Sea (Arhonditsis et al., 2000), and the Gulf of Mexico (Justić et al., 2005; Laurent et al., 2018). To date, few investigations of coastal eutrophication have occurred in Eastern Boundary Upwelling systems (EBUS). While strong upwelling and vigorous surface currents would generally limit the extent to which coastal eutrophication could occur (Fennel & Testa, 2019), such investigations have also been limited by coupled physical biogeochemical numerical modeling approaches that can adequately resolve fine-resolution bathymetry and the complexities of submesoscale circulation (McWilliams, 2016; Dauhajre et al., 2019), while simulating a sufficient duration (several years) to distinguish oceanic versus terrestrial forcing. These submesoscale circulation features, including fine scale eddies and filaments < 5 km in horizontal resolution, strongly control the magnitude and variability of nearshore

68 upwelling and associated nutrient transport. Thus, high resolution, submesoscale-resolving
69 numerical models are a necessary prerequisite for mechanistic modeling studies and source
70 attribution of oceanic versus terrestrial drivers of coastal eutrophication in EBUS. Inad-
71 equate modeling system and lack of numerical model validation have been identified as
72 significant barriers to effective, evidence-based solutions to coastal eutrophication (Boesch,
73 2019).

74 All the necessary ingredients are present to motivate a numerical modeling investigation
75 of the role of coastal eutrophication in driving ocean acidification and oxygen loss in the
76 Southern California Bight (SCB), a large marine open embayment found in the California
77 Current System (CCS) on the U.S. Pacific Coast. First, the SCB is a biologically-productive
78 region, and thus of high economic and ecological importance. Seasonal upwelling of nutrient-
79 rich deep water maintains high rates of biological productivity over broad scales. At the
80 same time, upwelling draws water masses that are naturally low in dissolved oxygen, pH,
81 and carbonate saturation state (Ω_{Ar}) onto the shelf and into the photic zone (Sutton et al.,
82 2017). Second, the SCB has one of the most spatially comprehensive and longest-running
83 coastal observational systems in the world. Several physical and biogeochemical variables
84 are sampled regularly and extensively, creating an ideal setting for model-data comparisons.
85 Third, the SCB is home to one of the most densely populated coastal regions in North Amer-
86 ica, where the discharges of primary or secondary treated wastewater from a population of
87 20 million people are released to the coastal zone via ocean outfalls, along with the urban
88 and agricultural runoff from 75 rivers. These nutrient sources rival natural upwelling in
89 magnitude (Howard et al., 2014), roughly doubling available nitrogen to nearshore coastal
90 waters. Intensifying ocean acidification, oxygen loss and harmful algal blooms have moti-
91 vated California policy makers to consider reducing anthropogenic nutrients as a climate
92 change mitigation strategy (Ocean Protection Council, 2018), but wastewater treatment
93 plant upgrades and methods to increase control or reduce non-point sources would cost
94 billions. A numerical modeling approach is needed to disentangle the effects of natural
95 upwelling and climate change from anthropogenic nutrient loading from land-based and
96 atmospheric sources.

97 To support such investigations, the regional oceanic model system (ROMS, Shchepetkin
98 and McWilliams (2005)) coupled to the biogeochemical elemental cycling model (BEC,
99 Moore et al. (2004)) has been recently adapted for the CCS (Renault, McWilliams, et
100 al., 2020; Deutsch et al., 2020). A downscaled model domain was established, scaling from
101 a 4 km horizontal resolution configuration spanning the entire CCS, to a 1 km resolution
102 grid covering the much of the California coast (latitude $< 40.25^\circ\text{N}$), to a 0.3 km grid in the
103 Southern California Bight (SCB), where investigations of local anthropogenic inputs were
104 focused. Modeling experiments investigating submesoscale transport (captured at model
105 resolutions ≤ 1 km) have demonstrated an up to ten-fold increase in the magnitude of in-
106 stantaneous vertical N fluxes (Kessouri, Bianchi, et al., 2020) relative to mesoscale transport
107 represented by a 4 km model (Section 2.2). Furthermore, a finer horizontal resolution of
108 bathymetry improves the representation of coastal currents, submesoscale circulation, and
109 coast-offshore connectivity (Dauhajre et al., 2019). For this reason, investigations of coastal
110 eutrophication are simulated here at 0.3 km horizontal resolution. Simulations conducted
111 with the 4 km ROMS-BEC model domain have been validated for regional-scale atmospheric
112 forcing, physics, and biogeochemistry, including O_2 , carbonate saturation state, primary
113 productivity, and hydrographic parameters, demonstrating that the model captures broad
114 patterns of critical properties in the CCS (Renault, McWilliams, et al., 2020; Deutsch et
115 al., 2020). However, additional focused validation of nearshore, anthropogenically-enhanced
116 gradients in nutrients, primary production, oxygen and pH in model simulations conducted
117 at 0.3 km resolution are needed to gauge model utility to investigate the impacts of coastal
118 eutrophication on ocean acidification and oxygen loss.

119 We employed this downscaled, submesoscale-resolving physical-biogeochemical model
120 to investigate the effects of land-based and atmospheric nutrient inputs in driving coastal

121 eutrophication and ocean acidification and oxygen loss. The aim of this manuscript is to: 1)
 122 document the SCB ROMS-BEC model configuration, including the effects of land-based and
 123 atmospheric inputs of nutrients and organic carbon, intended to support investigations of
 124 coastal eutrophication, and 2) present a validation of SCB ROMS-BEC simulations against
 125 available observations, focusing on anthropogenically-enhanced gradients in nutrients, pri-
 126 mary production, oxygen, and pH.

127 **2 SCB coupled physical and biogeochemical model description, configu-** 128 **ration and forcing**

129 **2.1 Model description**

130 *2.1.1 Ocean hydrodynamics*

131 Ocean hydrodynamics is modeled with the Regional Oceanic Modeling System (ROMS)
 132 (Shchepetkin & McWilliams, 2005), a free-surface, terrain-following coordinate model with
 133 3-D curvilinear coordinates that solves the primitive equations with split-explicit time steps.
 134 It contains state-of-art numerical algorithms that provide an accurate and stable representa-
 135 tion of physical processes down to scales of tens of meters, and allows for offline downscaling
 136 of high-resolution sub-domains within larger domains. The offline downscaling is based
 137 on the Orlanski scheme for the baroclinic mode (Marchesiello et al., 2001) and a modified
 138 Fletcher scheme for the barotropic mode (Mason et al., 2010). Vertical mixing in the bound-
 139 ary layers is represented by a K-profile parameterization (W. G. Large et al., 1994). The
 140 U.S. West Coast hindcast model has been successfully run over two decades at 1 and 4 km
 141 horizontal resolution using high-resolution spatial and temporal atmospheric forcing that
 142 represent the effects of wind drop-off, the current feedback on the surface stress, and high-
 143 frequency wind fluctuations (Renault, Hall, & McWilliams, 2016a; Renault, Molemaker,
 144 McWilliams, et al., 2016). For this study, we further downscale to 0.3 km resolution to
 145 capture submesoscale processes, and run the model for 4 consecutive years between January
 146 1997 and December 2000.

147 *2.1.2 Biogeochemistry*

148 Ocean biogeochemical modeling approaches can have a broad range of complexities,
 149 ranging from few functional groups (e.g. NPZD models, Fasham (1993)), to multiple bio-
 150 geochemical cycles (e.g. C, N, O) and plankton functional groups. To provide a repre-
 151 sentation of biogeochemical cycles, ROMS is dynamically coupled to the Biogeochemical
 152 Elemental Cycling (BEC) model (Moore et al., 2004; Gruber, 2004; Gruber et al., 2011;
 153 Deutsch et al., 2020). A schematic of BEC is shown in Fig. 1(b). BEC is a multi-element
 154 (C, N, P, O, Fe, Si) and multiplankton model that includes three explicit phytoplankton
 155 functional groups (picoplankton, silicifying diatoms, N-fixing diazotrophs), one zooplankton
 156 group, and dissolved and sinking organic detritus. The impacts of calcifying phytoplankton
 157 (coccolithophores) on the carbon system is represented implicitly. Remineralization of sink-
 158 ing organic material follows the multi-phase mineral ballast parameterization of Armstrong
 159 et al. (2001).” and “Sedimentary processes have also been expanded. Particulate organic
 160 matter reaching the sediment is accumulated and slowly remineralized with a timescale of
 161 330 days, to provide a buffer between particle deposition and nutrient release. Nitrogen loss
 162 to the sediment is parameterized according to the empirical diagenetic model for sediment
 163 denitrification of Middelburg et al. (1996). Water column denitrification is only active when
 164 oxygen concentrations fall below 5 mmol m⁻³. Sedimentary release of Fe is based on the
 165 benthic chamber measurements of (Severmann et al., 2010) for the California-Oregon coast,
 166 and increases as bottom water oxygen concentrations decrease. Atmospheric dust deposition
 167 follows the parameterization by Mahowald et al. (2006) and provides an additional source
 168 of iron at the surface, although of minor importance compared to sedimentary iron release
 169 in the region (Deutsch et al., 2020). The ecosystem is linked to a carbon system module
 170 that tracks dissolved inorganic carbon (DIC) and alkalinity, and an air-sea gas exchange

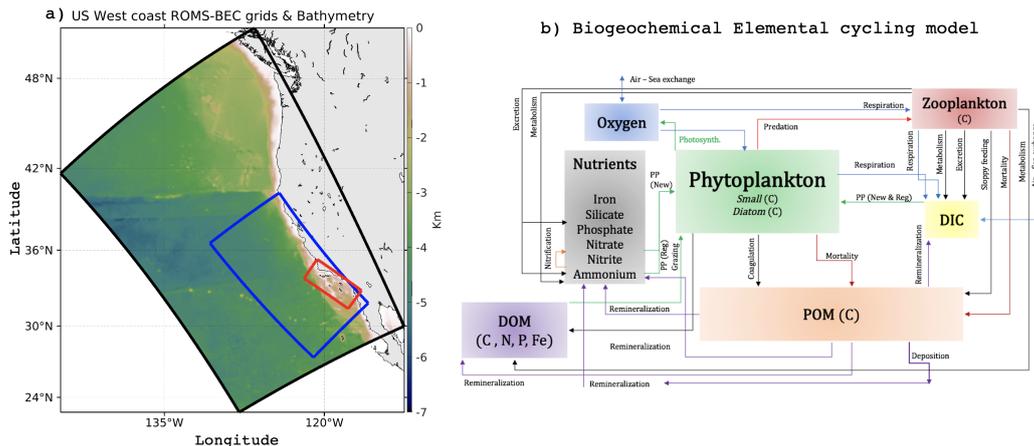


Figure 1: a) ROMS-BEC model configurations. $dx = 4$ km is the black box, $dx = 1$ km is the blue box, $dx = 0.3$ km is the red box. Background color contours show the topography from $dx = 4$ km. b) Schematic of the biogeochemical elemental cycling model. The schematic shows state variables (boxes) and biogeochemical rates and feedback (arrows).

171 module that allows realistic representation of dissolved gases (e.g. O_2 , CO_2 and nitrous
 172 oxide), based on the formulation of Wanninkhof (1992).

173 **2.1.3 Model configuration**

174 The SCB model domain extends along a 450 km stretch of the coast, from Tijuana to
 175 Pismo Beach, and about 200 km offshore. This grid, shown in Fig. 1a), is composed of 1400
 176 x 600 grid-points, with a nominal resolution of $dx = 0.3$ km. The grid has 60 σ -coordinate
 177 vertical levels using the stretching function described in Shchepetkin and McWilliams (2009),
 178 with the following stretching parameters: $\theta_s = 6$, $\theta_b = 3$, and $h_c = 250$ m. The model is
 179 run with a time step of 30 seconds, and output is saved as 1-day averages.

180 The oceanic forcing of the 0.3 km domain originates from multi-level offline downscal-
 181 ing. A 4 km simulation is initialized and forced at the open boundaries by a preexisting
 182 North-east Pacific-wide ROMS solution at 12 km resolution (Renault, McWilliams, et al.,
 183 2020), initialized and forced on the boundaries by the global model Mercator Glorys2V3
 184 (<http://www.myocean.eu>), and is run for the period 1995–2010, after a spin-up of 2 years.
 185 A 1 km simulation is initialized and forced from the 4 km model, starting in October 1996
 186 and ending in December 2007. The 0.3 km simulation is initialized and forced at its bound-
 187 aries by the 1 km simulation starting from January 1997 and ending in December 2000. The
 188 bathymetry used in this configuration comes from the Southern California Coastal Oceanic
 189 Observation System (SCCOOS) 3 Arc-Second Coastal Relief Model Development (90 m
 190 horizontal resolution).

191 The oceanic model is forced by hourly outputs from the atmospheric uncoupled Weather
 192 Research and Forecast model (WRF06; Skamarock and Klemp (2008)). Using bulk formu-
 193 lae (W. B. Large, 2006), WRF06 provides heat, surface evaporation, momentum and
 194 atmospheric data and is run at 6 km resolution over a domain similar to the 4 km (Fig. 1
 195 and used for Renault, Hall, and McWilliams (2016b)), and includes a wind-current coupling
 196 parameterization necessary to attain more realistic simulations of the oceanic eddy kinetic
 197 energy (EKE) and circulation (Renault, Molemaker, McWilliams, et al., 2016; Renault,
 198 Masson, et al., 2020).

199 Model simulations were conducted from 1997-2000, a period chosen to capture the
 200 effects of all three phases of the El Niño–Southern Oscillation (ENSO); it also captures the
 201 beginning of the "modern" state of point source management in the SCB, where several large
 202 Publicly Owned Treatment Plants (POTW) were in transition from primary to secondary
 203 treatment. (We will refer to submarine point sources outfalls from the treatment plants as
 204 "POTW" hereafter.)

205 2.2 Importance of submesoscale circulation

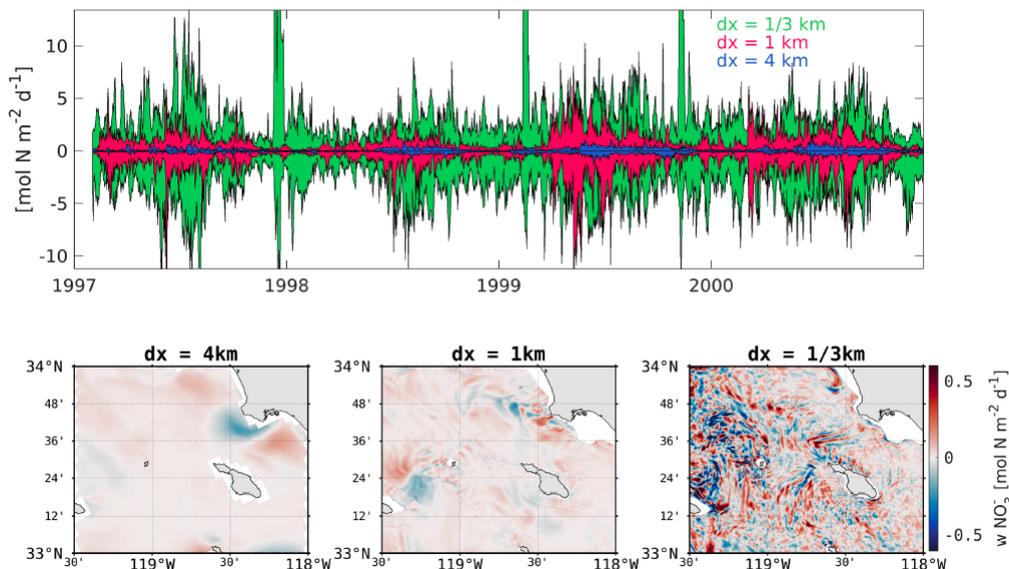


Figure 2: (Upper panel) Timeseries (1997-2001) of the vertical eddy flux of nitrate at 40 m depth calculated as follow: $\overline{wN} = \overline{wN} + \overline{w'N'}$, where the overbar represents a monthly average, and the prime the deviation from this average, for region covering the entire Southern California Bight. The minimum and maximum values (i.e. the envelope) of the flux are shown in blue for the 4 km solution, in red for the 1 km solution and in green for the 1/3 km. (Lower panel) Snapshot of the vertical flux of nitrate in spring at 40 m off the coast of Palos Verdes that shows higher magnitudes and enhanced variability as resolution increases.

206 Downscaling to $dx = 0.3$ km allows the model to represent ocean circulation that
 207 includes baroclinic and barotropic eddies and turbulence generated at the submesoscale
 208 (Capet, Campos, & Paiva, 2008). Resolving submesoscale eddies dramatically increases
 209 the variability of vertical fluxes of biogeochemical tracers and other material properties,
 210 eventually allowing a more accurate representation of chemical and biological constituents.
 211 Figure 2(upper panel) shows the temporal variability and horizontal distribution of vertical
 212 eddy fluxes of nitrate at 40 m from 3 different resolutions with the ROMS model (see
 213 section 2.1.3). Submesoscale dynamics increase instantaneous fluxes by more than one
 214 order of magnitude, with more frequent and vigorous fine-scale structures (Fig. 2(bottom
 215 panels)) when increasing the resolution from 4 km to 1 km, and similarly another order
 216 of magnitude when increasing resolution from 1 km to 0.3 km. Intensification of vertical flux of
 217 nitrate at the euphotic depth has previously been shown in idealized models (Mahadevan,
 218 2016; Lévy et al., 2012) and in realistic simulations in the central California upwelling
 219 system (Kessouri, Bianchi, et al., 2020), but has never been modeled in the SCB at this
 220 resolution. Submesoscale eddies have been associated with increased productivity in the

221 oligotrophic ocean (Mahadevan, 2016) and decreased productivity in the upwelling region
 222 (Kessouri, Bianchi, et al., 2020). Our submesoscale-resolving simulation at $dx = 0.3$ km is an
 223 opportunity to quantify the balances of nitrogen, dissolved oxygen, carbon and productivity
 224 using a more realistic representation of the physical circulation, as well as a representation
 225 of urban anthropogenic inputs to the ocean.

226 Inclusion of submesoscale dynamics energizes frontogenesis by mesoscale straining and
 227 mixed layer instabilities (Capet, Klein, et al., 2008; Capet, Campos, & Paiva, 2008; Capet,
 228 McWilliams, et al., 2008). Oceanic fronts are a driver of significant nutrient supply to the
 229 upper ocean. They have also been recognised as areas of enhanced biomass in many regions
 230 of the global ocean (Woodson & Litvin, 2015), as well as important locations for fisheries (e.g.
 231 (Galarza et al., 2009)). In our set of simulations, we show that the increased number of fronts
 232 and submesoscale instabilities promote intense variability of nitrate transport as shown
 233 in figure 2, as well as increased heterogeneity at the subsurface chlorophyll *a* maximum.
 234 However, surface phytoplankton biomass is only intensified if the timescale of the enrichment
 235 is sufficiently long and maintained in these small scale features. Modeling at this scale
 236 allows for a more accurate simulation of biogeochemical tracers and rates, as described in
 237 subsequent sections.

238 **2.3 Terrestrial and atmospheric forcing of freshwater, nutrients and carbon**

239 Model simulations were forced with a monthly time series of spatially-explicit inputs
 240 (Fig. 3, upper), including freshwater flow, nitrogen, phosphorus, silica, iron, and organic
 241 carbon representing natural and anthropogenic sources (Sutula et al., 2021b). These data
 242 include POTW ocean outfalls and riverine discharges (1997-2017) and spatially-explicit
 243 modeled estimates of atmospheric deposition. POTW effluent data were compiled from permit
 244 monitoring databases and communication with sanitary agencies. Monthly time series
 245 of surface water runoff from 75 rivers are derived from model simulations and monitoring
 246 data (Sutula et al., 2021b). Direct atmospheric deposition is derived from the Community
 247 Multi-scale Air Quality (CMAQ) model (Byun et al., 2006), and follows the implementation
 248 of Deutsch et al. (2020). In this paper, we discuss in detail the formulation of the river and
 249 wastewater outfall inputs.

250 **2.4 Configuration of river and wastewater outfall forcing in the model**

Ocean outfalls and coastal rivers are modeled as mass sources into the ocean (Fig. 3, upper). To accomplish this, we add explicit volume fluxes to the otherwise divergence-free flow in the ocean. The inclusion of these fluxes makes it possible to account for associated sources of tracers, while satisfying conservation laws. Specifically, our approach allows for the proper influx of fresh water in the ocean, without resorting to a ‘virtual salt’ flux, which is a common approach in larger scale ocean models (Kang et al., 2017). Since we explicitly include known volume fluxes for both rivers and outfall pipes, specification of tracer concentration is sufficient to correctly model the source terms. The tracer evolution equations that are used in ROMS are implemented by using control volumes (Shchepetkin & McWilliams, 2005) where for each tracer concentration $C = C(x, y, z, t)$,

$$\frac{\partial \iiint C dV}{\partial t} = \iint u_n C dA. \quad (1)$$

where $V = V(x, y, z, t)$ is the volume of the entire domain, u_n is the normal velocity into the volume and $A = A(x, y)$ is the total area of grid cells source is being input. Additionally, we enforce mass conservation which implies;

$$\frac{\partial V}{\partial t} = \iint u_n dA. \quad (2)$$

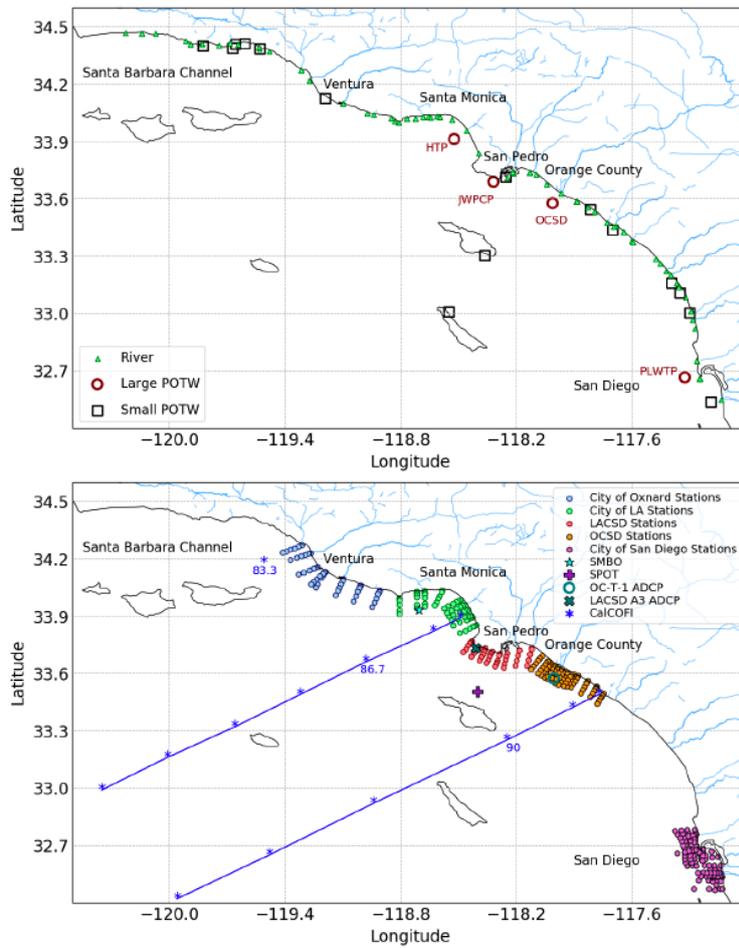


Figure 3: (Upper panel) Location of rivers and POTW outfalls along the SCB. (Lower panel) Location of monitoring stations used for the validation, including POTW quarterly monitoring surveys, CalCOFI seasonal observations, showing the line numbers, Santa Monica Bay Observatory (SMBO), and San Pedro Oceanographic Timeseries (SPOT), mooring.

251 In absence of rivers and outfalls, the flow is volume conservative, and the integral on the
 252 right hand side of Equation 2 is zero. Using Equations 1 and 2, it is easy to see that the mean
 253 concentration of a tracer can be lowered if the average concentration of the flux entering the
 254 control volume is less than the mean concentration in that volume. In this manner, fresh
 255 water rivers will lower the salinity of the water in which they enter. All 75 rivers and 23
 256 POTW pipes that are considered in this study are implemented in this manner.

Each individual source is based on the following equation:

$$S(x, y, z, t) = \frac{W(x, y, z) Q_s(t) C_s(t)}{V_s} \quad (3)$$

257 With:

- 258 $S(x, y, z, t)$: volume source of contaminant ($\text{mmol m}^{-3} \text{s}^{-1}$).
 259 $W(x, y, z)$: non-dimensional shape function (with values between 0 and 1).
 260 $Q_s(t)$: water volume flux from the source ($\text{m}^3 \text{s}^{-1}$).
 261 $C_s(t)$: concentration of the tracer C in the source water (mmol m^{-3}).
 262 V_s : effective volume of the source (m^3).
 263

For each source, $Q_s(t)$ and $C_s(t)$ are prescribed as time series. The shape function $W(x, y, z)$ distributes the tracer spatially and in the water column, representing non-resolved mixing and dilution effects. Its values represent the relative intensity of the *in situ* tracer injection, with values between 0 and 1. Tracer concentration C is distributed in the water column as $C(x, y, z, t) = W(x, y, z)C_s(t)$ The effective 3D volume of the source is calculated from the shape function $W(x, y, z)$ as:

$$V_s = \iiint W(x, y, z) dV \quad (4)$$

where the integral is over the model domain. For convenience, we assume that $W(x, y, z)$ can be separated into a horizontal shape function $A(x, y)$, multiplied by a vertical shape function $H(z)$ (both non-dimensional and with values between 0 and 1), such that:

$$V_s = \iint A(x, y) dx dy \int H(z) dz = A_s H_s \quad (5)$$

264 Here, A_s represents the effective source surface area (m^2), and H_s the effective source
 265 thickness (m). The functions $A(x, y)$ and $H(z)$ are defined differently for POTW and rivers.
 266 They are assumed to be fixed in time; a time-dependent generalization (for example to
 267 mimic variations in the depth of the POTW buoyant plume) is straightforward. For POTW
 268 inputs, at each main diffuser, the horizontal distribution $A(x, y)$ of the source is shown in
 269 Fig. S1. This method of weighting the plume in different cells allows the effluent to be
 270 properly diluted vertically and horizontally at this resolution and prevents the model from
 271 developing numerical instabilities.

Each large treatment plant has specialized outfall configurations that are taken into account for representation in the model (Fig. S1). The flow is divided in two at Hyperion Treatment Plant (HTP) located 6km off Marina Del Rey (Santa Monica Bay) (Fig. S1A) and Point Loma Wastewater Treatment Plant (PLWTP) in San Diego coast (Fig. S1D) to account for their Y-shaped diffuser, partitioning 50% of the flow to each diffuser. Orange County Sanitation District (OCSD) located 6km off Huntington Beach (Fig. S1C) has one flow through its L-shaped diffuser. Joint Water Pollution Control Plant (JWPCP) in Palos Verdes shelf (Los Angeles) (Fig. S1B) has three diffusers, the Y-shape northern typically discharges 17.5% of the flow for each leg of the Y-diffuser, and the southern L-shape diffuser discharges 65% of the flow. The vertical profile of the POTW sources is defined by a Gaussian function centered at a height z above the bottom (h_b), to mimic a buoyant plume, so that $H(z)$ is given by:

$$H(z) = e^{-z^2/d_s^2} \quad (6)$$

272 Where $z = -h_b + h_s$, with
 273 h_b : bottom depth (m).
 274 h_s : depth of the buoyant plume above the bottom (m).
 275 d_s : vertical scale of the POTW plume (m).
 276 We further assume $h_s = 20$ m and $d_s = 10$ m, as in Uchiyama et al. (2014).

277 We distribute the SCB rivers on one horizontal grid point (0.3 km wide), where we
 278 assume $A(x, y) = 1$, and similarly distribute the source vertically, with the Gaussian function
 279 centered at the surface. h_s here is simply the water column depth to put the maximum input
 280 at the surface. Because in ROMS the thickness of vertical grid cells varies in time, to ensure
 281 tracer conservation the calculation of the input source volume V_s must be done at each time
 282 step, even in the case of a time-independent source shape function $W(x, y, z)$. Effectively,
 283 only $H_s = H(z)$ needs to be recalculated at each time step.

284 3 Model performance assessment approach

285 The conceptual approach for model performance assessment is comprised of three com-
 286 ponents, addressing different aspects of skill assessment: 1) statistical comparison of model
 287 output to observational data for state variables by region and season; 2) comparison of model
 288 output to observational data for biogeochemical rates; 3) evaluation of model behavior com-
 289 pared to expected biogeochemical dynamics for coastal zones. Comparison of model output
 290 to observational data by region and season is designed to document model skill at reproduc-
 291 ing the statistics (e.g., mean values and variability) of ocean physical and biogeochemical
 292 parameters at the spatio-temporal scales more relevant for evaluating human impacts on the
 293 coastal environment. Comparison of model output to observational data for biogeochemi-
 294 cal rates assures that model is capturing the appropriate transformations in nutrients and
 295 carbon that structure the ecosystem response to eutrophication. Finally, the evaluation of
 296 model behavior compared to the expected physical and biogeochemical dynamics for coastal
 297 zones is a more qualitative evaluation of model performance to document that the model
 298 broadly reproduces oceanographic phenomena in a way that reflects our understanding of
 299 nearshore ocean environments.

300 3.1 Description of Observational Datasets

301 3.1.1 Ship-Based Ocean Monitoring

302 The SCB is home to a suite of long-running monitoring programs that make it one of
 303 the best observed coastal ecosystems in the world (3, lower). Among them, the Califor-
 304 nia Cooperative Oceanic Fisheries Investigations (CalCOFI) program (McClatchie, 2016),
 305 initiated in the 1950s, samples the SCB quarterly each year, collecting hydrographic and
 306 biogeochemical measurements in coordination with the Southern California Coastal Ocean
 307 Observing System (SCCOOS). These observations are augmented nearshore by quarterly
 308 surveys of nearshore water column and benthic parameters conducted collaboratively since
 309 1990 by POTW agencies as a part of their regulatory monitoring requirements (Howard
 310 et al., 2014; McLaughlin et al., 2018; Booth et al., 2014; Nezlin et al., 2018). These pro-
 311 grams provide good temporal and geographical coverage of both the offshore (CalCOFI) and
 312 nearshore (POTW) areas, coinciding with the model period, and include publicly available
 313 water quality data for targeted sites measured quarterly. We validated model output against
 314 observed temperature, dissolved oxygen, nitrate, ammonium, chlorophyll, carbon-system
 315 parameters (pH and aragonite saturation state), primary production, and nitrification.

316 *In situ* measurements have inherent uncertainty, due to a combination of measurement
 317 sensitivity and sampling frequency and intensity, making them an imperfect “truth” with
 318 which to compare to model output. However, this uncertainty is not the same for all param-
 319 eters. Both temperature and dissolved oxygen are collected using high resolution probes,
 320 though the two programs used in this study incorporate slightly different calibration pro-

321 protocols for dissolved oxygen, lending greater confidence to data-model comparisons for these
 322 datasets. Chlorophyll is measured on discrete bottle samples in the CalCOFI program,
 323 a high quality measurement, but inferred from *in situ* fluorescence measurements in the
 324 POTW monitoring program, adding uncertainty to these measurements. Nitrate and am-
 325 monium concentrations are measured on discrete bottle samples for both programs, but the
 326 detection limits are more sensitive in the CalCOFI program. Furthermore, nutrients are not
 327 measured with the same sampling density in POTW monitoring programs as sensor data.
 328 Similarly, primary production is measured at a subset of locations in the CalCOFI program
 329 and as a short-term special study in Southern California Bight Regional Marine Monitoring
 330 Program (Bight Program). Details on measurements and sample collection protocols for the
 331 CalCOFI program can be found on their website (<https://calcofi.org> ; McClatchie (2016))
 332 and for the POTW monitoring programs in Howard et al. (2014). We also use selected
 333 nutrient observations from the Santa Monica Bay Observatory (SMBO) mooring located in
 334 the Santa Monica Bay (Leinweber et al., 2009). Figure 3 shows a map of all monitoring
 335 stations used in this study. The repository of data can be found in Kessouri, McLaughlin,
 336 et al. (2020).

337 *3.1.2 High Frequency Radar and Acoustic Doppler Current Profilers*

338 High Frequency Radar (HF) data from the database of the University of California, San
 339 Diego (<https://hfrnet-tds.ucsd.edu/thredds/catalog.html>) provide surface currents along the
 340 west coast of the United States, including the SCB. Seasonally averaged data from 2012-2020
 341 were used to analyze trends of surface currents in the Bight compared to the model. Acous-
 342 tic Doppler Current Profilers (ADCP) provide current data in the water column. ADCP
 343 measurement data from Orange County Sanitation District (OCSD) for the period June
 344 1999 to June 2000 and Los Angeles County Sanitation District (LACSD) during November
 345 2000 to June 2007 were used to validate vertical profiles of currents.

346 *3.1.3 Remote sensing observations*

347 Satellite ocean color measurements for chlorophyll were used to characterize horizontal
 348 gradients at finer scales and higher density than possible with the ship-based monitoring.
 349 We use monthly averaged surface chlorophyll concentration from the period 1997 to 2000
 350 derived from the SeaWiFS sensor at 4 km spatial resolution. Large gaps in the dataset
 351 can occur because of dense cloud cover that occurs in late spring and early summer. The
 352 products of the Vertically Generalized Production Model (VGPM) net primary production
 353 algorithm (Behrenfeld & Falkowski, 1997) were also considered for this validation. Despite
 354 limitations, satellite data provide a valuable representation of the spatial distribution of
 355 chlorophyll, temperature, and primary production at seasonal scales over the region.

356 **3.2 Performance Statistics**

357 Our approach to a statistical assessment of agreement between model predictions versus
 358 observations reflect the fact that the hydrodynamic model, under the influence of realistic
 359 forcings (e.g. wind fields) and without data assimilation, develops its own intrinsic vari-
 360 ability in circulation, e.g. submesoscale eddies (McWilliams, 2007). The resulting modeled
 361 state variables would not necessarily overlap with observations on a point-by-point basis,
 362 but would be comparable to observations when averaged over appropriate spatio-temporal
 363 scales. We assessed a suite of statistics and metrics, following the methodology of Allen et
 364 al. (2007), to assess how well the model reproduces the magnitude and gradients of selected
 365 state variables, whether the model agreement has an apparent bias, and how well the model
 366 reproduces natural variability. We calculated six metrics, defined in the following, where N
 367 is the total number of appropriate observational data, D represents each individual observa-
 368 tional datum, M is the model estimate representing
 369 an observation, and \bar{M} is the mean of the model estimate. The metrics considered include:

The Pearson correlation coefficient, reflecting the degree of linear correlation between the observed and model variable, and the statistical significance (p-value) of this correlation:

$$r_{xy} = \frac{\sum_{n=1}^N (D_n - \bar{D})(M_n - \bar{M})}{\sqrt{\sum_{n=1}^N (D_n - \bar{D})^2} \sqrt{\sum_{n=1}^N (M_n - \bar{M})^2}}; \quad (7)$$

The Cost Function (CF), which gives a non-dimensional value indicative of the “goodness of fit” between two sets of data, quantifying the difference between model results and measurement data:

$$CF = \frac{1}{N} \sum_{n=1}^N \frac{|D_n - M_n|}{\sigma_D} \quad (8)$$

370 where σ_D is the standard deviation of the observations;

The Percentage Bias (PB) (the sum of model error normalized by the data):

$$PB = \frac{\sum (D - M)}{\sum D} * 100; \quad (9)$$

The Ratio of the Standard Deviations (RSD):

$$RSD = \frac{\sigma_D}{\sigma_M} \quad (10)$$

371 where σ_M is the standard deviation of model outputs;

The Nash-Sutcliffe Model Efficiency (ME) (Nash & Sutcliffe, 1970), a measure of the ratio of the model error to the variability of the data:

$$ME = 1 - \frac{\sum (D_n - M_n)^2}{\sum (D - \bar{D})^2}; \quad (11)$$

And the two-sample t-test, or Welch’s t-test (Welch, 1947; Derrick et al., 2016):

$$H = (\bar{D} - \bar{M}) / \sqrt{\frac{\sigma_D^2}{N} + \frac{\sigma_M^2}{N}}. \quad (12)$$

372 We score the model performance following Table 1 per the methodology of Allen et al.
373 (2007).

Statistic	Excellent	Good	Reasonable	Poor
Cost Function (Moll & Radach, 2003)	<1	1-2	2-3	>3
Nash Sutcliff Model Efficiency (Nash & Sutcliffe, 1970)	>0.65	0.65-0.5	0.5-0.2	<0.2
Percentage Bias (Maréchal, 2004)	< 0.1	0.1-0.2	0.2-0.4	> 0.4
H (Welch, 1947)	0			1
Correlation Coefficient	1-0.9	0.9-0.8	0.8-0.6	<0.6
p-value	<0.05			>0.05
Ratio of Standard Deviations	1-0.9, 1-1.1	0.9-0.8, 1.1-1.2	0.8-0.6, 1.2-1.4	<0.6, >1.4

Table 1: Model performance

374 4 Model performance assessment findings

375 4.1 Ocean circulation

376 The SCB is situated at the confluence of water masses from the subarctic Pacific via the
 377 California Current, and from the eastern tropical North Pacific via the California Undercur-
 378 rent, which all interact with the local topography, the coast, and the atmosphere to sustain
 379 variability in circulation on inter-annual, seasonal, and intraseasonal time scales (Dong et
 380 al., 2009; Bograd et al., 2015). The effects of this variability in circulation has profound
 381 consequences for coastal ocean biogeochemistry (Gruber et al., 2011; Bograd et al., 2015;
 382 Nagai et al., 2015; Nezlin et al., 2018), and is therefore critical that the model accurately
 383 simulates spatial and temporal variability in circulation patterns.

384 Figure 4 shows the hydrodynamic characteristics of the SCB in the model compared to
 385 data. In the northern SCB, the model shows similar qualitative and quantitative patterns
 386 for the horizontal circulation compared to HF data (Fig. 4(a)-(b)) and as seen in Dong et
 387 al. (2009). The circulation in the SCB is characterized by northward currents in the first 20
 388 km of the coast and cyclonic circulation in the middle of the SCB that is stronger in summer
 389 and weaker in winter. The model successfully reproduces observed current patterns, with
 390 similar current magnitudes. The intensity of the northward coastal branch of the current is
 391 on average about $0.15\text{-}0.3\text{ m s}^{-1}$ in summer versus $0.05\text{-}0.15\text{ m s}^{-1}$ in winter. The offshore
 392 southward branch is generally about 0.3 m s^{-1} all year round (Fig. 4(a)-(b)). The dominant
 393 current in the coastal band (15 km from coast) of the SCB flows northward, and follows the
 394 topography along isobaths on the shelf (Fig. 4(g)-(h)).

395 The simulated June 1999–June 2000 variability of the current in depth is shown in
 396 the vertical profiles extracted off the coast of Palos Verdes and Orange county compared
 397 to the ADCP data at the same locations (Fig. 4(c)-(f)). The location of both of these
 398 profiles are a few kilometers from the continental slope and therefore capture a suite of
 399 physical processes, including mesoscale and submesoscale eddies, fronts, jets, and internal
 400 tides (Capet, McWilliams, et al., 2008; Kim et al., 2011; Dong et al., 2009). The model
 401 generally reproduces the means and range of the variability shown in these close to shore
 402 horizontal currents, which demonstrates that ROMS at $dx = 0.3\text{ km}$ resolution captures the
 403 submesoscale variability described in Section 2.2.

404 In the northern SCB, cyclonic vortices are generated inside the Santa Barbara Chan-
 405 nel (Fig. 4(i)) when the northward current that flows along the Ventura coast meets the
 406 eastern side of the Channel Islands, with higher intensity in summer (Fig. 4(a) versus (b))
 407 (Winant et al., 2003). Submesoscale eddies are particularly prominent in this region, in
 408 particular persistent cyclonic eddies that drive an upward doming of isopycnals (Fig. 4(j))
 409 (McGillicuddy Jr, 2016), which supplies nutrients to the euphotic layer. The model correctly
 410 reproduces this vertical transport, described in Brzezinski and Washburn (2011), and the
 411 high concentrations of nitrate and other nutrients in the upper layers of the Santa Barbara
 412 Channel, as further detailed in Section 4.3.1.

413 In the central and southern SCB (latitude $< 34.7^\circ\text{N}$), the model successfully captures
 414 flow regimes around the large POTW outfalls, indicating that it can appropriately represent
 415 the dispersal of wastewater plumes in these regions. In the Santa Monica and San Pedro
 416 Bays, topography drives the circulation of currents inside the Bays, converging back to
 417 the main current offshore (Fig. 4(g)-(h)). On top of the Hyperion and JWPCP outfalls
 418 (in the Santa Monica Bay and offshore of the Palos Verdes peninsula, respectively), the
 419 current is mostly south-eastward. Near the OCSO outfall, the current direction varies in
 420 winter between south-eastward and north-westward, but is primarily southward in summer
 421 (Fig. 4(a)-(b), (e)-(f)). At the PLWTP outfall, the current is narrow, with a dominant
 422 south-eastern direction, parallel to the coast, demonstrated by both model and HF radar
 423 data.

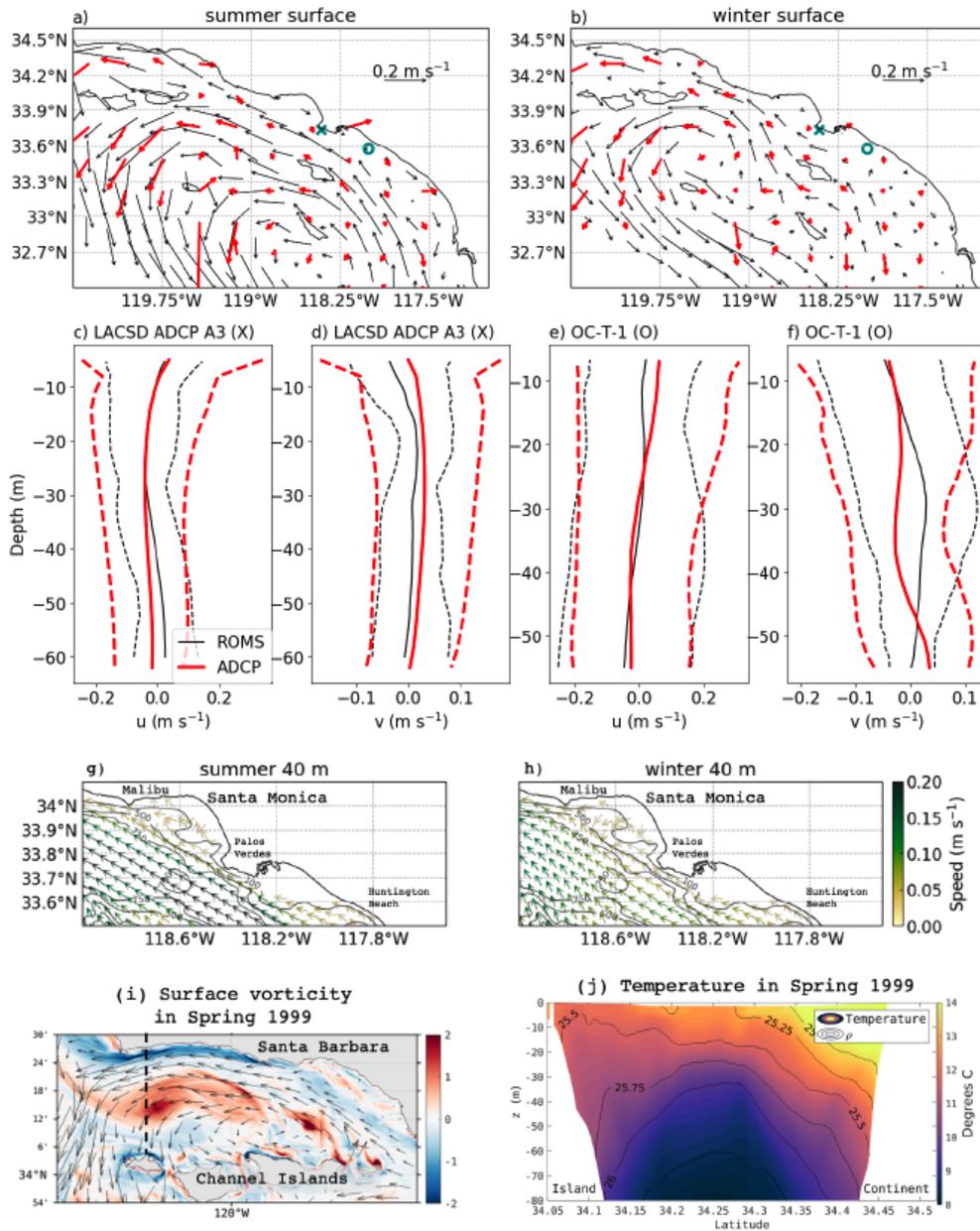


Figure 4: a) Mean surface currents in the Southern California Bight from HF data during 2012-2020 (thick red arrows) and model during 1999-2000 (black arrows) in summer and b) winter. c)-f) Vertical profiles of horizontal velocity components from ADCP instruments (thick red lines) and model (thinner black lines). The two dashed lines indicate the 5th and 95th percentile current values. c)-d) ADCP data come from the LACSD mooring A3 stationed at the teal ‘X’ in a)-b) and e)-f) come from the OCSD mooring OC-T-1 located at the teal ‘O’. g) Mean model current direction and speed (colored) at 40 m depth with bathymetry contoured in summer and h) winter. i) Surface model vorticity normalized by f in spring in Santa Barbara Channel showing cyclonic eddies. j) Cross-section of temperature and density isopycnals as drawn by the dashed line in (i) from model to show eddy-driven uplifting of the isopycnals in the center of Santa Barbara Channel.

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4.2 Vertical gradients and seasonal variability of temperature and mixed layer depth

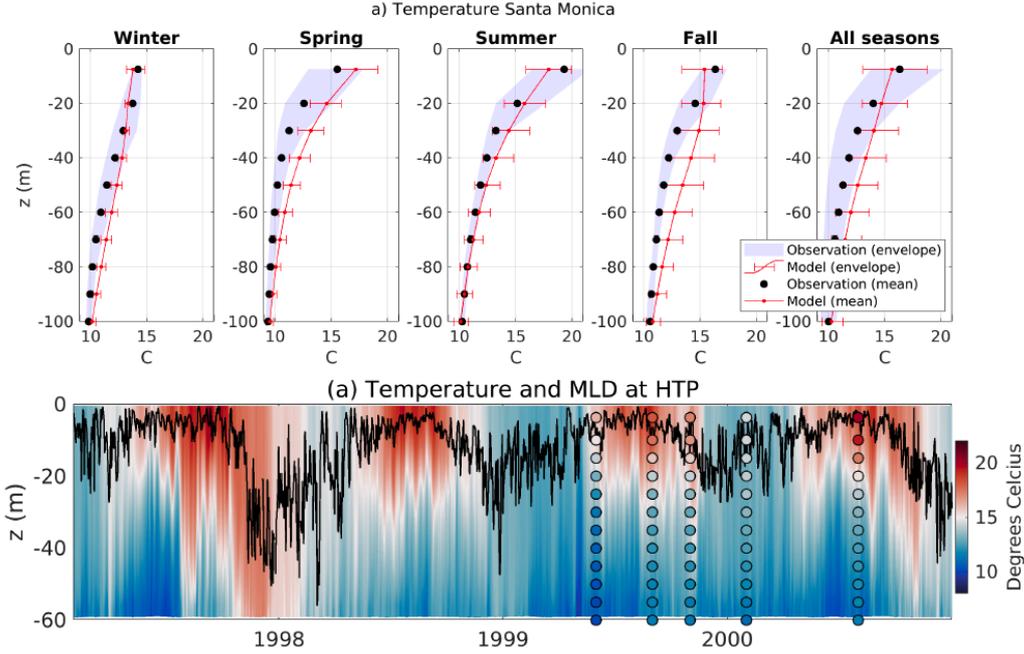


Figure 5: (a) Average seasonal profiles of temperature in the Santa Monica Bay. The red lines and red bars show the spatio-temporal mean and the variability from the model respectively. The black dots and the gray shading show the spatio-temporal mean and the variability from *in situ* data (City of LA stations), respectively. (b) Hovmöller diagram of temperature at the location of the Hyperion POTW outfall (HTP) in the Santa Monica Bay issued from the model. The black line shows the simulated time-series of mixed layer depth. The deepest mixing occurs during El Niño 1998 (>40 m). Colored dots are average concentrations from *in situ* measurements.

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The model successfully reproduces the three-dimensional and seasonal variability of physical tracers, here exemplified by temperature. Temperature is the parameter in which we have the highest confidence in the observational record, because observations are abundant, and sensors are accurate and precise, regularly calibrated, and with negligible drifts. The greatest source of observational uncertainty is temporal under-sampling, but some sources of model bias may also be important (e.g., from atmospheric forcing, wind, or shortwave detailed in Renault, McWilliams, et al. (2020)). Quantitative statistical analysis indicates that model performance is ‘*excellent*’ or ‘*good*’ for nearly all metrics for all regions and seasons (see Table 2). The lowest performance of the model is characterized as ‘*reasonable*’ for certain sub-regions (Palos Verdes, Orange County, and San Diego) in spring and fall (Palos Verdes only) (see Supporting Information Table S2). As noted above, this may be due to under-sampling during these months, which can be highly variable because the region is shifting between a well-mixed to a more stratified ocean regime. Detailed information on the other sub-regions and their statistical comparison can be found in the Supporting Information, Tables S1 to S4.

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Following common practices (de Boyer Montégut et al., 2004), we define the mixed layer depth (MLD) as the depth at which temperature decreases from its surface value by more than 0.2°C . On average, the MLD deepens from the coast to offshore, and varies with season

444 (e.g. in Santa Monica Bay in Fig. 5b). The model successfully simulates the seasonal cycle
 445 of MLD along the coast. For example, the model recreates the observed seasonal deepening
 446 of the mixed layer in the Santa Monica Bay to depths greater than 16-20 m (the typical
 447 depth of the upper signature of the POTW plumes, see Section 4.3.2) nearly every winter
 448 (black line in the Fig. 5b).

449 Regular winter shows a homogeneous upper layer of $< 14^{\circ}\text{C}$ temperature, and a mixed
 450 layer located at 18-20 m in the coastal region and 40-60 m offshore. The surface ocean is
 451 colder around the Channel Islands ($\text{SST} < 12^{\circ}\text{C}$) (see Fig. 19). In the open ocean, the model
 452 reproduces the de-stratification with deepening of the thermocline to about 70m and a MLD
 453 at about 40m (Fig. 6c and d). In summer, stratification is the strongest, reflecting an intense
 454 vertical temperature gradient, and the MLD (both in the model and in the observations) is
 455 found few meters below the surface (approximately 10 m). Temperature varies rapidly from
 456 more than 20°C at the surface in the southern domain ($16\text{-}17^{\circ}\text{C}$ in the northern domain)
 457 to less than 12°C at 50m depth over the entire SCB (see also Fig. 19). In the open ocean,
 458 the model succeeds in reproducing the stratification that brings the seasonal thermocline to
 459 50m and the MLD to 15m (Fig. 6c and e)).

460 The model reproduces interannual variability in MLD under the influence of El Niño-
 461 Southern Oscillation (ENSO, hereafter referred to as El Niño, i.e., the period from fall 1997
 462 to spring 1998 in Fig. 5b), when the MLD reached 40 m. We show that during winter of
 463 El Niño year, the entire water column of the SCB is warmer than on average, and surface
 464 temperature is more homogeneous, varying between 15.5 and 17°C (Fig. 6a). In the open
 465 ocean, during El Niño, with warmer upper layer than regular winters, the model shows
 466 good performance in reproducing the deepening of the seasonal thermocline (>120 m) and
 467 of the MLD (>50 m) (e.g. offshore Santa Monica Bay in Fig. 6a and b). These patterns of
 468 variability in temperature are consistent with regional observations of El Niño in the SCB
 469 (Todd et al., 2011).

470 4.3 Dissolved Inorganic Nitrogen

471 4.3.1 *Spatial patterns and seasonality of nitrate*

472 Nitrate observations are only broadly available in the offshore CalCOFI dataset, so
 473 only large-scale regional patterns in nitrate concentration can be validated. There is a clear
 474 seasonality of nitrate, where surface concentrations are higher in spring and summer, and
 475 decrease in fall and winter (Fig. 7). The model reproduces the average seasonal patterns
 476 observed in the *in situ* nitrate data across multiple regions. The model also captures along-
 477 shore variability in coastal nitrate concentrations, reproducing values greater than 25 mmol
 478 N m^{-3} off Santa Barbara, 20 mmol N m^{-3} off Los Angeles, and 15 mmol N m^{-3} off San
 479 Diego.

480 The model also reproduces observed patterns in the depth of the nitracline (Mantyla
 481 et al., 2008; Nezlin et al., 2018), which tends to follow sloping density surfaces in the
 482 region. These patterns include: the high values at the euphotic depth limit ($\sim 50\text{m}$ below
 483 the surface) along the Santa Barbara coast in spring; the doming of the nitracline in the
 484 center of the Santa Barbara Channel (Fig. 7b); the 20 to 30 m deep nitracline along the
 485 Los Angeles coast; and the deepening of the nitracline from about 30 m at the coast to
 486 more than 60 m offshore in San Diego. In the offshore region of the SCB, the model is
 487 consistent with observations showing high nitrate (>20 mmol N m^{-3}) around the Channel
 488 Islands (not shown) as compared to less than 5 mmol N m^{-3} farther offshore. This pattern
 489 is strongest in winter and summer, when the offshore regions are particularly oligotrophic
 490 (surface $\text{NO}_3^- < 1$ mmol m^{-3}) throughout the SCB.

Santa Monica: Temperature

	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.96 E	7E-06 E	0.05 E	-0.04 E	1.10 G	0.81 E	716
Spring	0 E	0.98 E	8E-07 E	0.10 E	-0.10 G	0.78 R	0.51 G	716
Summer	0 E	0.97 E	9E-06 E	0.04 E	-0.02 E	1.07 E	0.93 E	712
Fall	0 E	0.89 G	3E-06 E	0.09 E	-0.08 E	0.98 E	0.51 G	718
All Seasons	0 E	0.95 E	3E-05 E	0.08 E	-0.07 E	1.02 E	0.73 E	2862

Table 2: Statistical comparison between *in situ* data and model outputs for temperature profile in Santa Monica Bay (City of LA stations). Letters next to numbers indicate model performance: E = Excellent, G = good, R = reasonable, P = Poor.

Santa Monica: Ammonium

	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.94 E	0.06 P	0.54 E	0.24 R	1.86 P	0.68 E	20
Spring	0 E	0.85 G	0.14 P	0.58 E	-0.57 P	0.69 R	-0.61 P	21
Summer	0 E	0.58 P	0.42 P	0.72 E	0.19 G	1.76 P	0.29 R	21
Fall	0 E	0.91 E	0.09 P	0.42 E	0.07 E	1.47 P	0.80 E	21
All Seasons	0 E	0.81 G	0.10 P	0.36 E	-0.03 E	1.23 R	0.60 G	83

Table 3: Statistical comparison between *in situ* data and model outputs for ammonium profile in Santa Monica Bay.

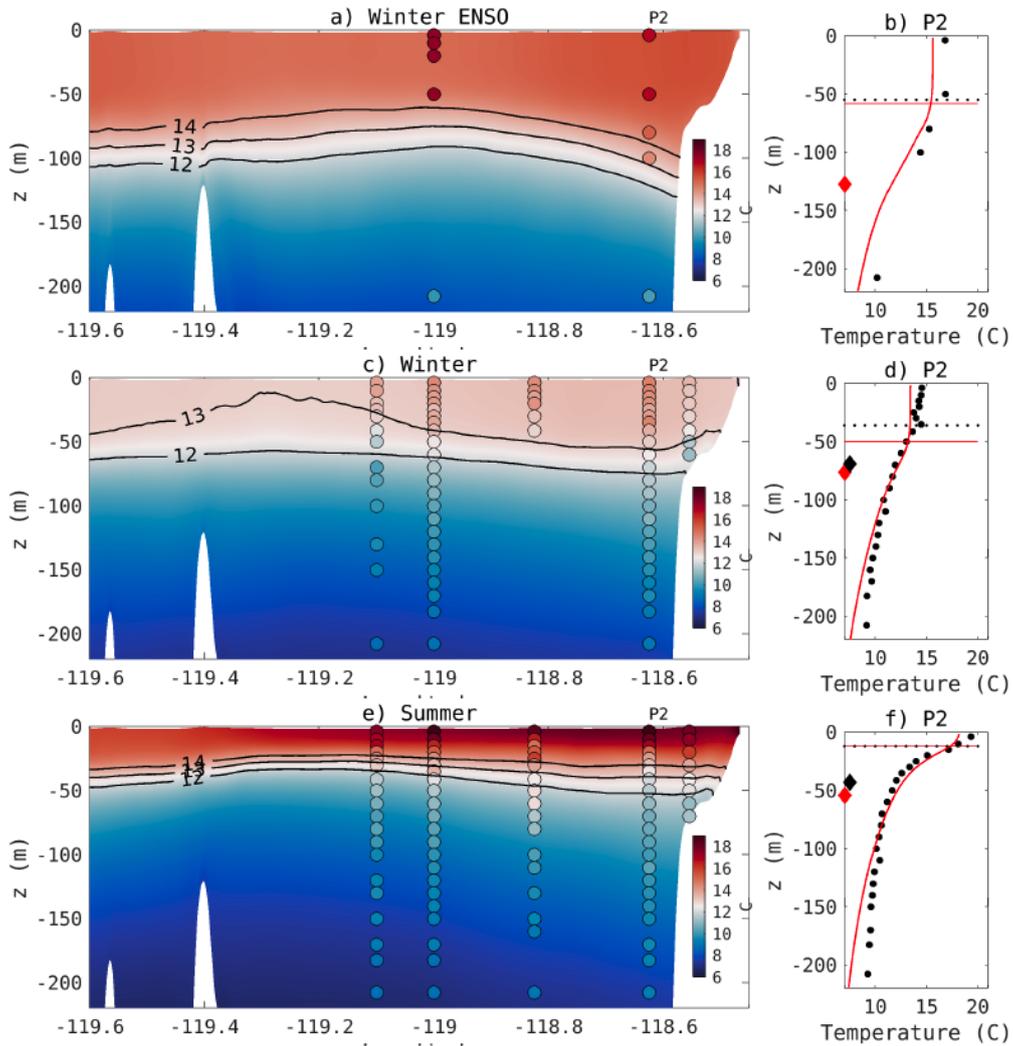


Figure 6: (a) Cross section of average temperature following line 86.7 from CalCOFI monitoring stations during an El Niño winter (12/1997 to February 1998). (b) Profile at station P2. Black dots are CalCOFI *in situ* data, red line is the simulated profile. The horizontal line is the MLD (black is CalCOFI, red is simulated). Diamonds (black is CalCOFI, red is simulated) is the depth of the maximum gradient to estimate the depth of the seasonal thermocline at 12°C. (c-d) are similar to (a-b) for average winter, and (e-f) are for average summer.

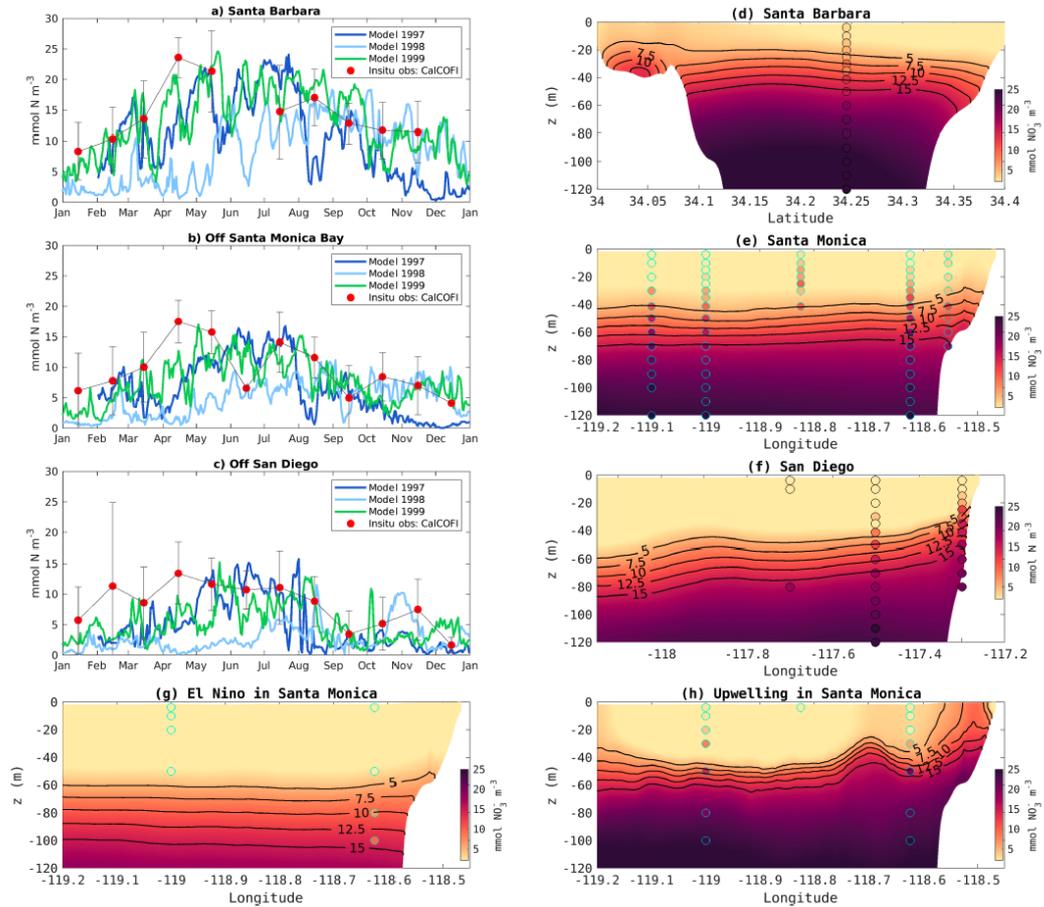


Figure 7: (a-c) Time series of nitrate concentration at 50 m depth in three different locations of the SCB: (a) is near the center of Santa Barbara Channel, (b) is offshore the Santa Monica Bay, and (c) is offshore San Diego. Model outputs are represented by the lines for three different years, with the dots showing mean values from *in situ* measurement from CalCOFI, and gray bars the standard deviation from the mean. The time-series show prominent interannual variability in addition to seasonal variability. While the years 1997 and 1999 show similar nitrate distributions, the El Niño period between the end of 1997 to 1998 is significantly different, showing nearly uniform concentrations between November 1997 through May 1998. This is caused by the deepening of the thermocline during El Niño, which depresses the nutricline. (d-f) Cross sections showing the average springtime nitrate concentration in (d) the Santa Barbara region (e) the SM region, and (f) the SD region. Background are model outputs and dots are CalCOFI *in situ* measurements. Model and *in situ* data agree on the vertical and seasonal patterns in the three regions. They highlight the main differences in these three regimes, consisting of a shallower nitracline in the Santa Barbara Channel, and a deeper nitracline in southern waters. (g-h) Comparison of nitrate concentrations during (g) winter El Niño (January-March 1998) and (h) during an upwelling event (the first week of May 1999) to illustrate the ability of the model (vs. *in situ* CalCOFI data) to simulate the vertical displacement of the nitracline during these specific events.

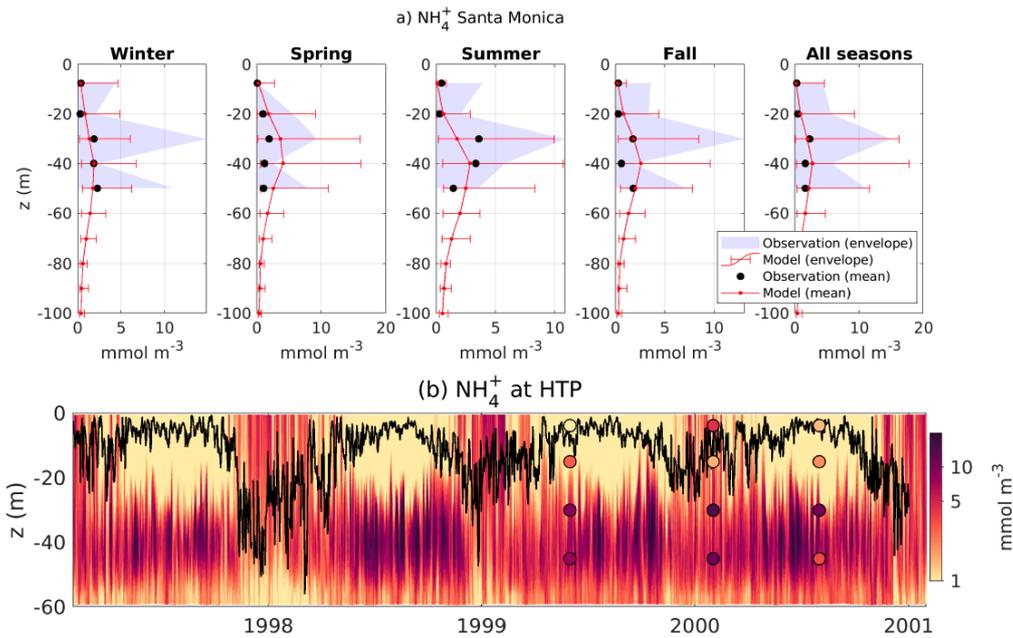


Figure 8: As for Fig. 5, but for ammonium concentration. These profiles are showing agreement on intensity, seasonality and shape of the vertical profile with exceptionally high concentrations at mid-depth.

4.3.2 Vertical gradients and seasonal variability of ammonium

Ammonium concentrations above a natural background concentration of 1 mmol N m^{-3} are indicative of POTW wastewater plumes. The model reproduces the observed average vertical profile of ammonium in the Santa Monica Bay, falling within the range of observed variability (Fig. 8a). Similar figures for other regions are shown in the Supporting Information (Fig. S6-Fig. S9). All regions show a similar maximum concentration between 30 to 45 m below the surface, in all seasons. The highest concentrations are seen in summer, when stratification is stronger, while lower concentrations in winter likely reflect increased dilution by seasonal mixing from the deepening of the mixed layer (Fig. 8b). Near ocean outfalls, both model and observations show mid-depth peaks of ammonium concentration, occasionally exceeding 10 mmol m^{-3} , which considerably overshadow values observed away from outfalls. In the model, these high ammonium concentrations are caused by wastewater plumes.

The main source of uncertainty in data-model comparisons is the limited spatial and temporal coverage of measurements. Ammonium is typically measured near ocean outfalls and is therefore biased towards high concentrations, but the dataset is highly variable. Methodological difficulties exist with the measurement of ammonium in seawater, and as such, we excluded non-detectable ammonium values in our analyses. Near the submarine outfalls, ammonium concentrations are likely extremely heterogeneous due to buoyant plume filaments, as observed in DiGiacomo et al. (2004) and in Warrick et al. (2007) in the Santa Monica Bay, as well as in other regions (e.g. Florida, in Marmorino et al. (2010)). These plume filaments are caused by horizontal advection and straining of the discharged effluent by currents. As a result, the under-sampling of ammonium may have led to poor statistical agreement between observations and model output. The model shows high to moderate agreement for the shape of the profile and the mean concentration (Table 3). However,

516 p-values for the correlations were not always significant. Similarly, there were often biases
 517 and low performance regarding variability statistics. This low model performance can be
 518 explained by the following two reasons: (1) spatial sampling is likely missing plume filaments,
 519 for example observational data points with high ammonium values that are capturing the
 520 plume are recorded next to very low or non-detectable values; and (2) the resolution of
 521 the model (0.3 km), as well as model averaging over the day, season, and depth range
 522 causes plume filaments to appear more uniformly spread near the outfalls. Because plume
 523 filaments are lost in this averaging, the model represents plumes as cloud-like distributions
 524 around outfalls; nevertheless, the average ammonium concentration of wastewater plumes
 525 is reasonably well represented. Detailed information on the other sub-regions and their
 526 statistical comparison can be found in the Supporting Information Tables S1 to S4.

527 **4.3.3 Horizontal gradients of ammonium**

528 Both *in situ* observations (dots in Fig. 9, Fig. 8a) and model output (background
 529 colours in Fig. 9 and red line in Fig. 8a) show high concentrations of ammonium in the
 530 subsurface layer below the thermocline (Fig. 9c), which we refer to as "high-ammonium
 531 plume". This high-ammonium plume can extend from Huntington Beach to South Ventura,
 532 encompassing three of the four major wastewater treatment plant outfalls in the SCB (See
 533 Section 2.4). Both model and observations show that the width and strength of the high-
 534 ammonium plume are greatest in summer compared to other seasons. The Santa Monica
 535 Bay Observatory mooring (SMBO, Leinweber et al. (2009)) located 17 km north-west of the
 536 submarine pipe Hyperion in Santa Monica Bay (Fig. 9g) frequently recorded concentrations
 537 higher than 2 mmol m^{-3} , and up to 4 mmol m^{-3} at mid-depth (Fig. 9e), consistent with the
 538 model (Fig. 9f). The depth of the maximum variability is at 40 m in the model, and slightly
 539 shallower in the SMBO data, possibly because of a mismatch in the time period (1997-2000
 540 for the model, and 2004-2010 for the SMBO). During winter, the model indicates vertical
 541 mixing and dilution of the plume at the surface. Accordingly, ammonium concentrations
 542 decrease slightly at depth (Fig. 9a) and increase at the surface, reaching values up to $2\text{-}6$
 543 mmol m^{-3} , also consistent with observations around the outfall pipes (Fig. 8a).

544 **4.3.4 Spatial patterns in rates of nitrogen transformation**

545 Although we had no *in situ* nitrogen transformation rates with which to compare model
 546 output during the simulation period, several datasets exist for the region that can serve
 547 as a test for whether the model is simulating reasonable patterns in rates via the right
 548 mechanisms. We found that modeled rates do agree with observed nitrogen transformation
 549 rates. Nitrification rates, the sequential oxidation of NH_4^+ to NO_3^- via NO_2^- , have been
 550 observed to be higher within wastewater plumes in the SCB (McLaughlin et al., 2021), a
 551 pattern driven by high ammonium concentrations in the discharges (McLaughlin, Nezhlin, et
 552 al., 2017). In both observations and the model, nitrification predominately occurs below the
 553 euphotic layer. Modeled vertically-integrated nitrification rates vary between 0.15 and 1.5
 554 $\text{mmol N m}^{-2}\text{d}^{-1}$, consistent with observations within the SCB and in the California Current
 555 (Table 5). The model also reproduces higher nitrification rates within wastewater plumes
 556 (See Supporting Information Fig. S22). There is also good agreement between observed and
 557 modeled rates of nitrate and ammonium uptake by phytoplankton communities (McLaughlin
 558 et al., 2021) and (Kudela et al., 2017). Modeled nitrate uptake rates vary between 2 and 11
 559 $\text{mmol N m}^{-2}\text{d}^{-1}$ and ammonium uptake rates vary between 6 and $51 \text{ mmol N m}^{-2}\text{d}^{-1}$
 560 in the Los Angeles and Orange County coasts, consistent with observations in the SCB (Table
 561 5).

562 **4.4 Chlorophyll concentrations**

563 In general, the model was found to reproduce vertical and horizontal gradients in chloro-
 564 phyll concentration in different subregions (Fig. 12). The timing of blooms was consistent

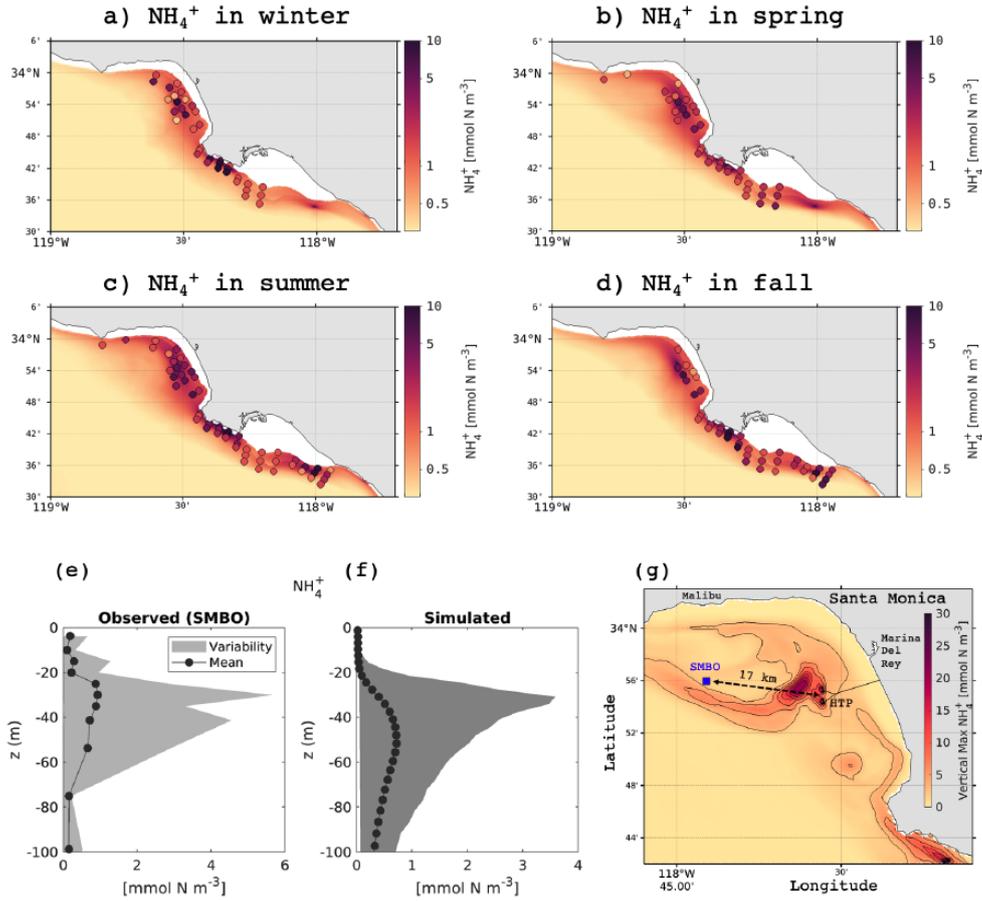


Figure 9: (a-d) Colors show seasonal average ammonium concentration between 30 and 45 m depth from the model, and dots from observations. High values highlight the movement and dispersion of subsurface wastewater plumes along the Orange and Los Angeles counties. The highest concentrations are located within a narrow coastal band of about 10 - 15 km width, and are carried along the topography by the mean currents. (e-f) show a statistical comparison of the vertical profiles of ammonium at the SMBO mooring and the same location in the model. The anthropogenic ammonium plume signature is apparent, albeit intermittently, 17 km away from the Hyperion outfall. (g) shows the simulated vertical maximum concentration of NH_4^+ averaged during a representative day to illustrate the dispersal of the effluent toward SMBO originating from the 2 diffusers of Hyperion Treatment Plant (HTP).

565 with changes in mixing and nutrient delivery in the SCB. We present three different subre-
 566 gions characterized by distinct hydrodynamic regimes: the Santa Barbara Channel, the Los
 567 Angeles coast, and San Diego coast.

568 There are several sources of uncertainty in the chlorophyll, primary production, phyto-
 569 plankton growth, and grazing rates observational records. For chlorophyll, bottle measure-
 570 ments are accurate and precise, but measure a limited portion of the water column. Sensors
 571 are accurate and precise in their measurement of fluorescence and have a rapid response
 572 time, providing vertically resolved profiles; however, the algorithm to convert fluorescence
 573 to chlorophyll concentration is inaccurate for the SCB. As a result, a correction factor has
 574 been applied to Bight data which adds uncertainty to the observational dataset (Nezlin et
 575 al., 2018). Satellite measurements of chlorophyll are inferred from ocean color (Kahru et
 576 al., 2009). This method works well offshore, but breaks down nearshore where terrestrially-
 577 derived colored dissolved organic matter creates uncertainty in reported satellite chlorophyll
 578 estimates on the order of 100% or greater (Zheng & DiGiacomo, 2017). For primary
 579 production, the incubation method to derive the rates is sensitive and precise (Cullen, 2001),
 580 though measured rates are subject to bottle effects and there is some ambiguity as to whether
 581 the experiments measure net primary production or gross primary production (Regaudie-de
 582 Gioux et al., 2014). Phytoplankton growth and zooplankton grazing are also determined
 583 experimentally, and duplicate measurements indicate that these methods are not very pre-
 584 cise, with differences between duplicates ranging from 80% to 200% (Landry et al., 2009; Li
 585 et al., 2011). For all three measurements, spatial and temporal under-sampling, particularly
 586 during seasons with high variability, adds uncertainty to the data-model comparison.

587 *4.4.1 Horizontal gradients in chlorophyll*

588 Despite the uncertainties outlined above, the model successfully simulates horizontal
 589 gradients in chlorophyll in the three subregions (Santa Barbara, Los Angeles and San Diego).
 590 The model captures the early, wide-spread spring bloom in the Santa Barbara Channel,
 591 which occurs as a combination of a coastal bloom driven by spring upwelling, followed by a
 592 bloom in the central and southwestern regions of the Channel (near the islands) in spring
 593 and summer (Fig. 10). The latter is driven by the strengthening of the cyclonic circulation
 594 in the Channel, which transports nutrients to the upper layers, and is regularly observed
 595 in the region (Brzezinski & Washburn, 2011). The model captures the strong seasonality
 596 in chlorophyll, wherein concentrations change from near zero in winter to up to 8 mg Chl
 597 m^{-3} in spring. Of the three regions, the blooms off Santa Barbara extends further into
 598 late summer and fall, where the average concentration is approximately 1-2 mg Chl m^{-3} , a
 599 pattern replicated in both model and observations .

600 In the Los Angeles subregion, the model predicts broad patterns in chlorophyll concen-
 601 trations, including a persistent bloom in the San Pedro Bay, consistent with observations
 602 (Nezlin et al., 2012), and validated by comparison with remote sensing (Fig. 11). Both
 603 satellite-derived and modeled chlorophyll show concentrations in the San Pedro Bay consis-
 604 tently higher than 3 mg Chl m^{-3} year-round, often extending into the Santa Monica Bay.
 605 The model also reproduces the strong offshore gradients in chlorophyll, where across less
 606 than 15 km offshore surface concentrations are reduced 3 to 4 fold (<1 mg Chl m^{-3}) further
 607 decreasing towards the open ocean. The model also reproduces the timing and magnitude of
 608 the blooms in the Santa Monica and San Pedro Bays. The difference in timing of maximum
 609 chlorophyll concentrations between the Santa Monica and San Pedro Bays likely reflects
 610 differences in nutrient supply. Nutrients, in particular ammonium, are available near the
 611 surface during winter (see Section 4.3.2), reflecting more vigorous mixing of the wastewater
 612 plume and land-based nutrient supply by rivers (in particular in the San Pedro Bay) during
 613 winter storms (Lyon & Stein, 2009). Storms and winter mixing events have been connected
 614 to phytoplankton blooms in the region (Nezlin et al., 2012; Mantyla et al., 2008). Fur-
 615 ther offshore in the Los Angeles region, the model recreates the weak seasonality of surface
 616 chlorophyll, with higher concentrations during winter and spring, and lower concentrations

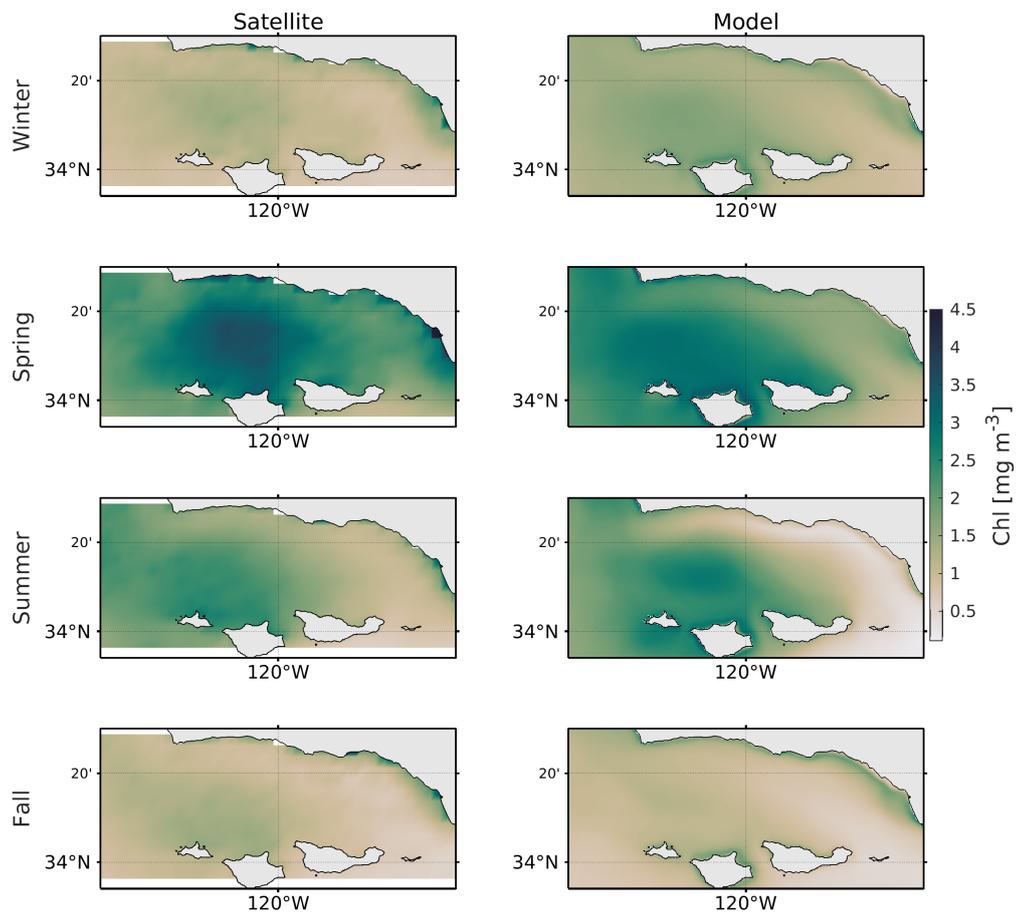


Figure 10: Comparison of seasonally-averaged surface chlorophyll between SeaWiFS remote sensing data (left panels) and the model (right panels) in the Santa Barbara Channel, where an important seasonal bloom is observed.

617 in summer and fall. In the offshore region of the Santa Monica Bay, the seasonal cycle is
 618 marked by the increase of surface phytoplankton between March and May as shown in Fig.
 619 12b. Mean chlorophyll values reach up to 3 to 4 mg Chl m^{-3} in April and May, although
 620 concentrations below 2 mg Chl m^{-3} are more common, consistent with observations over
 621 the same period.

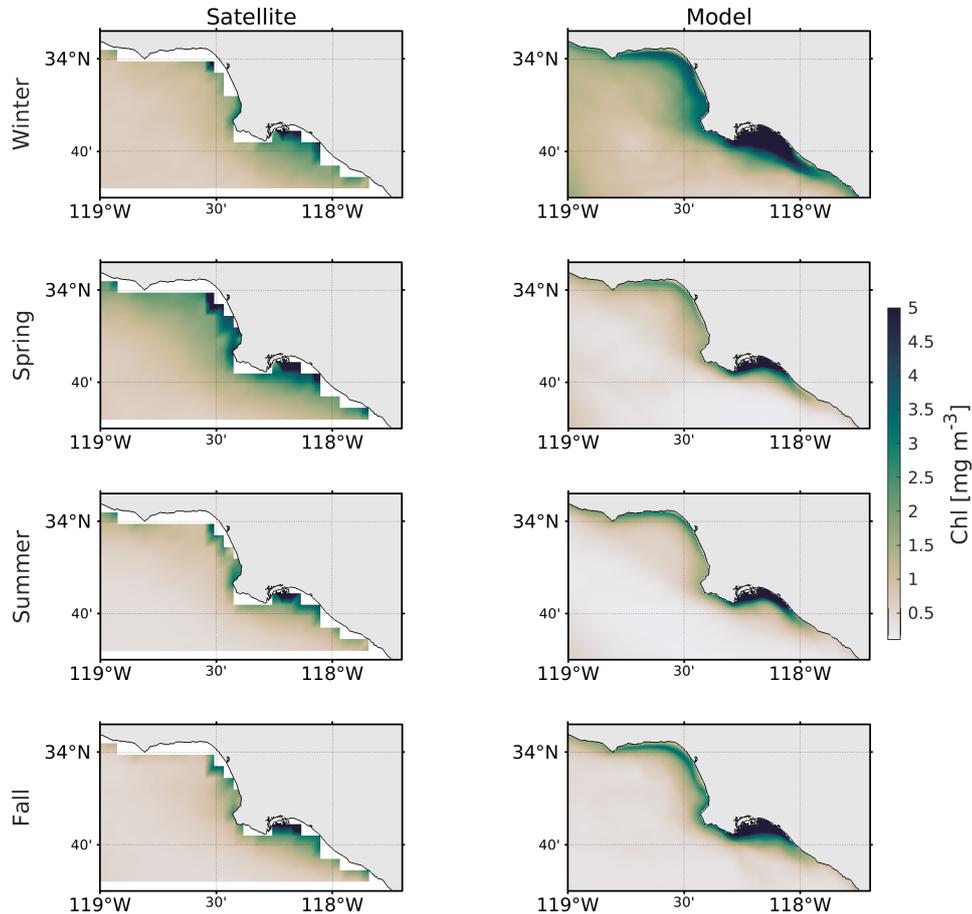


Figure 11: Comparison of seasonally-averaged surface chlorophyll between SeaWiFS remote sensing data (left panels) and the model (right panels) for years 1998-2000 in the Santa Monica and San Pedro Bays, where major POTW outfalls are found. The figure highlights the persistent coastal phytoplankton bloom, and the sharp inshore-offshore gradients.

622 Offshore of the San Diego coast, the model recreates a slight increase in surface chloro-
 623 phytoplankton in March; however, concentrations are generally below 1 mg Chl m^{-3} year-round (Fig
 624 12(c)). The oligotrophic conditions of the southern Bight (Nezlin et al., 2012; Mantyla
 625 et al., 2008) have been attributed to a deeper nitracline, which in turns supports a deep
 626 chlorophyll maximum layer (Mantyla et al., 2008). This feature is well represented in the
 627 model, which reproduces relatively high concentrations of chlorophyll in subsurface layers
 628 (generally between 20 and 90 m depth in the region).

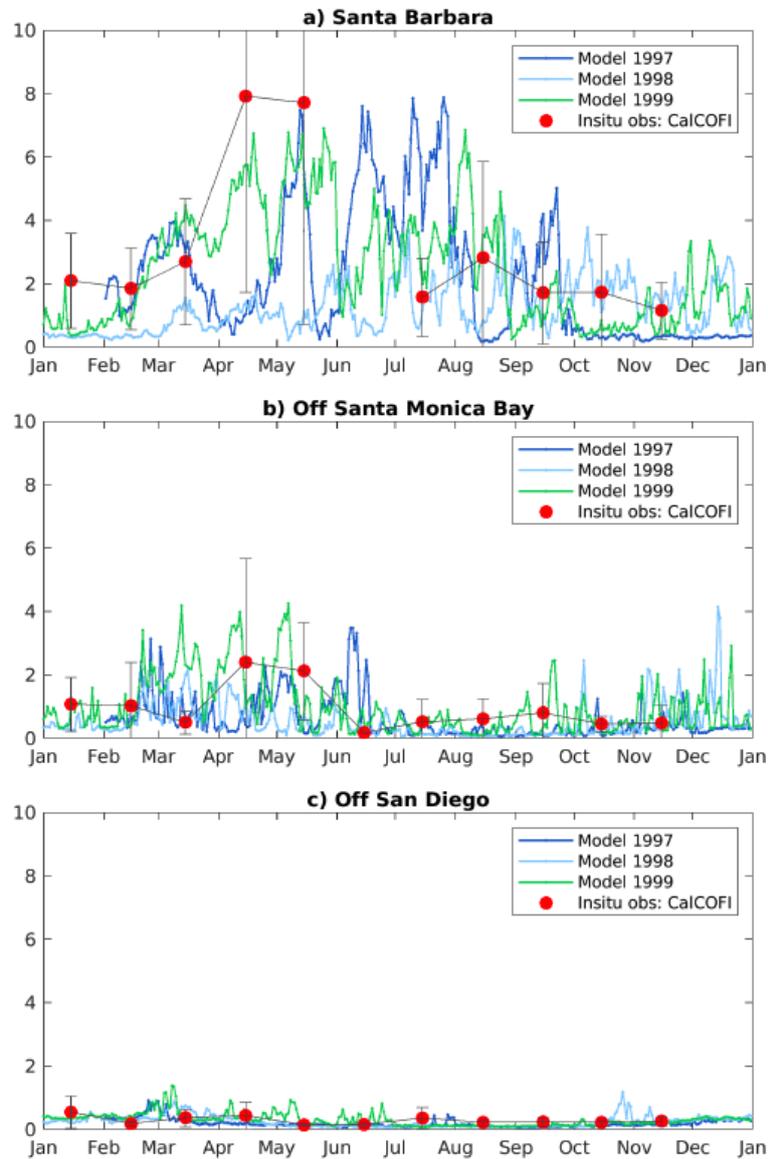


Figure 12: Comparison of surface chlorophyll concentration between different years of model output, and a climatology from CalCOFI *in situ* data. (a) is near the center of Santa Barbara Channel, (b) is offshore the Santa Monica Bay, and (c) is offshore San Diego. The model reproduces different productivity regimes across the Southern California Bight, with highly productive waters in the northern region, where average concentrations greater than 3 mg m^{-3} are observed for more than half of the year, and oligotrophic southern regions, where average surface concentrations rarely exceed 1 mg m^{-3} .

629

4.4.2 Vertical gradients and seasonal variability of chlorophyll

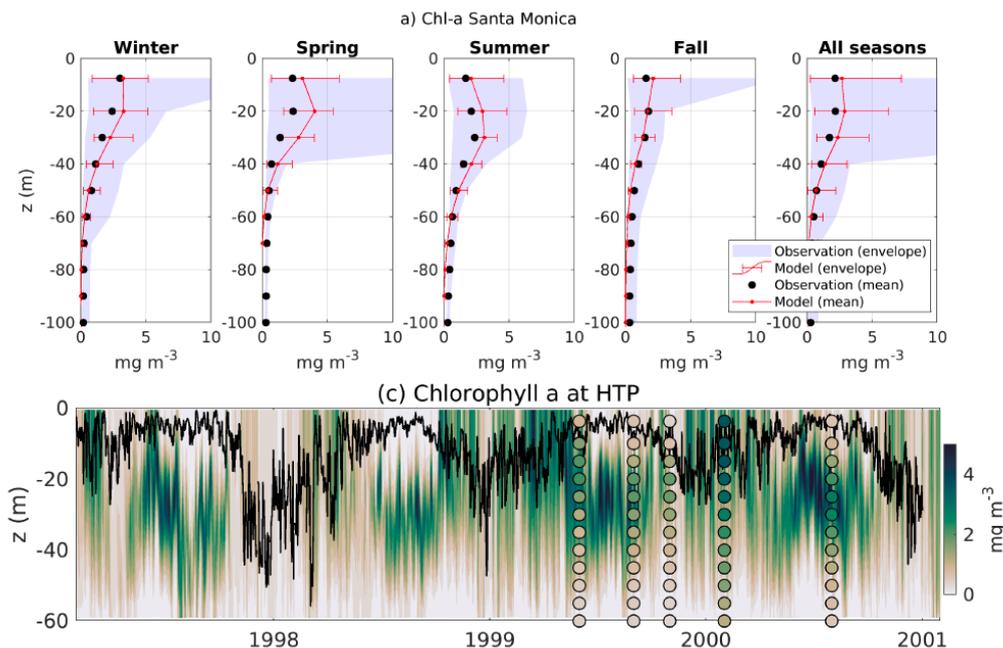


Figure 13: As for Fig. 5, but for chlorophyll concentration. Vertical profiles show a good agreement between simulated and *in situ* data, and display the formation of a subsurface chlorophyll maximum in summer, and a surface maximum in winter and spring. Concentrations in winter vary up to $+5 \text{ mg Chl m}^{-3}$. Note the very low concentrations during 1998 El Niño in the entire water column.

630 The goodness-of-fit statistical metrics (correlation coefficient and cost function) for
 631 chlorophyll are generally *excellent* or *good* for most seasons for all sub-regions (Table 4).
 632 We were most concerned with performance for these metrics because the remaining statistics
 633 may be affected by the aforementioned uncertainties due to the fluorometry calibration. The
 634 observational measurements should be internally consistent (if not accurate), so the shapes
 635 of profiles should be “correct” even if the magnitude is off due to poor calibration, and the
 636 model was able to replicate these shapes accurately. Despite calibration issues, the model
 637 reproduced chlorophyll reasonably well for the northern Bight sub-regions of Santa Monica
 638 Bay (Fig. 13) and Ventura/Oxnard (SI Fig. S11). Similar figures for other regions are shown
 639 in the Supporting Information (Fig. S10-Fig. S13). All show that the model is reproducing
 640 the magnitude and general shape of observed profiles. However, the model did not capture
 641 the variability for most regions (except for Palos Verdes), generally scoring *reasonable* or
 642 *poor* in the ratio of standard deviations for most seasons, particularly spring. This is likely
 643 a result of the spatial and temporal averaging. Chlorophyll is highly variable in space
 644 and time and under-sampling in either of these dimensions will adversely affect variability
 645 estimates for a region and season. Therefore, reasonable performance for these metrics
 646 was not unexpected. This suggests that the model may provide a conservative estimate
 647 of phytoplankton biomass in the southern Bight, while reproducing accurate spatial and
 648 temporal patterns in that biomass.

649 In addition to transporting nutrients from depth, upwelling ‘seeds’ surface waters with
 650 subsurface water masses dominated by selected phytoplankton species, stimulating surface
 651 blooms near the coast (Seegers et al., 2015). The model successfully reproduces this process,

652 wherein the subsurface chlorophyll maximum shoals and intensifies in spring, forced by the
 653 vertical movement of the thermocline driven by upwelling. This seasonal dynamics occurs
 654 across the domain in the model.

655 Offshore, in the more oligotrophic portion of the SCB, the model predicts that more
 656 than 60% of the maximum concentration of phytoplankton biomass remains below the sur-
 657 face year-round, constantly fed by subsurface nutrients injections. This is consistent with
 658 observations of a deep chlorophyll maximum throughout the region (Nezlin et al., 2018;
 659 Mantyla et al., 2008; Seegers et al., 2015), and with observations at the San Pedro Oceanic
 660 Time-Series (SPOT) located between the Palos Verdes Peninsula and Catalina Island (Fig.
 661 3, lower panel). At SPOT, a region weakly influenced by anthropogenic nutrients inputs
 662 at the surface, the model realistically simulates the seasonal cycle of chlorophyll. While
 663 ammonium does not exceed typical "natural" values of $\sim 1 \text{ mmol m}^{-3}$ below the surface,
 664 chlorophyll concentrations regularly reach more than 2 mg m^{-3} between 20 and 40 m in
 665 summer, in agreement with *in situ* measurement (Teel et al., 2018; Beman et al., 2011).
 666 (Additional figures to support the analysis are reported in the Supporting Information, Fig.
 667 S23.)

668 However, in regions more heavily influenced by anthropogenic nutrients, such as the
 669 Santa Monica Bay, the chlorophyll maximum progressively deepens from the surface in
 670 winter to about 25 to 30 m depth in spring and summer, with chlorophyll concentrations
 671 exceeding 5 mg Chl m^{-3} (Fig. 13a). This subsurface chlorophyll maximum is maintained
 672 for four to five months (Fig. 13b) before the stratification is weakened by winter mixing.

673 **4.4.3 Primary production**

674 Validation of the rates of primary production, phytoplankton growth and zooplankton
 675 grazing (Table 5) provides an independent check on mechanisms responsible for chloro-
 676 phyll as a state variable. The spatial and temporal frequency of these data, garnered from
 677 CalCOFI observations and literature values, is low. The most data as well as the most
 678 standardized methodologies are available for primary production. However, many of the
 679 primary production measurements used in this validation do not temporally coincide with
 680 the model period. Despite these uncertainties, the model generally reproduces expected
 681 large-scale patterns and seasonal variability in primary production.

682 This large scale variability was also mentioned in Deutsch et al. (2020). Model and data
 683 both show lower productivity in winter (Fig. 14a,c) and higher in spring (Fig. 14b,d), when
 684 the primary production is high along the coastal band, in the northern Bight around the
 685 Channel Islands (Fig. 14d), consistent with observations (Fig. 14b). This is also consistent
 686 with the so-called "green ribbon" of high-chlorophyll observed along the coast throughout
 687 the SCB (Lucas et al., 2011). The model reasonably reproduces the seasonal cycle of primary
 688 production in each of the subregions.

689 Phytoplankton are generally limited by a combination of nutrients and light, the latter
 690 of which is only limiting at depth in the SCB (Deutsch et al., 2020).

691 In winter, nitrogen is high at the surface in the northern SCB, and thus is not limiting.
 692 In the southern SCB, light and nitrogen are co-limiting due to stronger stratification, leading
 693 to oligotrophic conditions. In spring and through the summer, nitrogen is limiting nearly
 694 everywhere except in the Santa Barbara Channel and near the Channel Islands, where
 695 upwelling and submesoscale eddies maintain high nutrients at the surface.

696 The scatter plots in Fig. 14e-f show a comparison of the simulated primary production
 697 between the *in situ* CalCOFI data and that derived from remote sensing (empirically ad-
 698 justing the Behrenfeld-Falkowski Vertically Generalized Production Model [VGPM]). The
 699 model shows a correlation coefficient of about 0.6 with CalCOFI, similarly to that reported
 700 by Kahru et al. (2009) when comparing the VGPM product with CalCOFI. The model

Santa Monica: chlorophyll

	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.99 E	9E-06 E	0.48 E	0.09 E	0.91 E	0.94 E	714
Spring	0 E	0.93 E	9E-05 E	0.90 E	-0.42 P	0.52 P	-0.49 P	716
Summer	0 E	0.99 E	1E-08 E	0.58 E	-0.07 E	0.60 R	0.47 R	712
Fall	0 E	0.99 E	8E-08 E	0.48 E	0.16 G	0.75 R	0.76 E	718
All Seasons	0 E	0.99 E	4E-08 E	0.50 E	-0.01 E	0.73 R	0.80 E	2860

Table 4: Statistical comparison between *in situ* data and model outputs for chlorophyll profile in Santa Monica Bay.

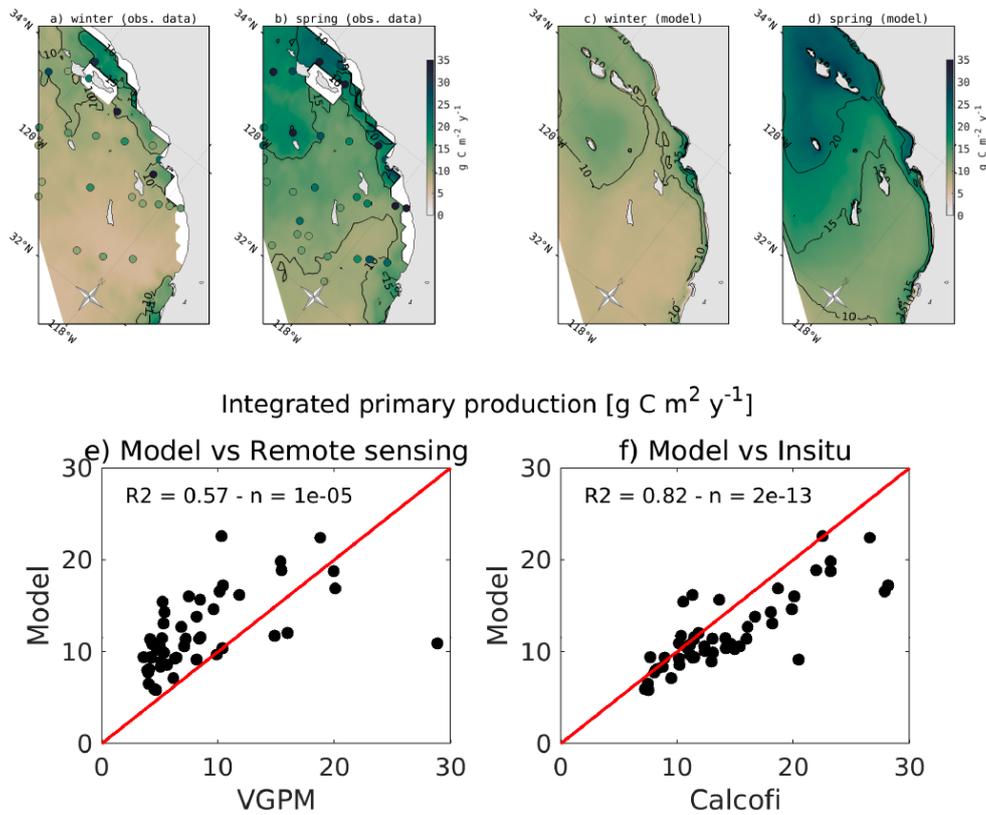


Figure 14: (a)-(b) Maps of vertically integrated Vertically Generalized Production Model (VGPM) net primary production and CalCOFI *in situ* measurements plotted as dots for (a) winter (January and February) and (b) spring (April to June). (c)-(d) Maps of vertically integrated primary production from the model, in (c) winter and (d) spring. Note the higher values for CalCOFI *in situ* measurements as compared to the satellite estimate, in better agreement with the model.

701 shows a stronger correlation with VGPM data, with a correlation coefficient of the order of
702 0.8.

703 Finally, while slightly outside our model domain and simulation period, the modeled
704 phytoplankton growth and zooplankton grazing rates were within the same order of mag-
705 nitude as the measured rates from the California Current Long Term Ecological Research
706 project (CC-LTER, see Landry et al. (2009), and Table 5) in the northern portion of the
707 Bight.

	Bight 13	Literature	Model
Primary production ($\text{g C m}^{-2} \text{ y}^{-1}$)	47.4, 1037.4		250, 1660
Nitrification ($\text{mmol m}^{-3} \text{ d}^{-1}$)	0, 0.225	0.02, 0.08	0.001, 0.27
NO_3^- Uptake Rate ($\text{mmol N mg Chl}^{-1} \text{ d}^{-1}$)	0.005, 2.16		0.03, 0.15
NH_4^+ Uptake Rate ($\text{mmol N mg Chl}^{-1} \text{ d}^{-1}$)	0.10, 8.30		0.08, 0.15
Total Phytoplankton Growth μ (d^{-1})		0.05, 0.8	0.3, 0.4
Grazing (d^{-1})		0.02, 0.5	0.3, 1.5

Table 5: Comparison of biogeochemical rates between published literature and model. Values are minimum and maximum. Literature values come from Landry et al. (2009); Li et al. (2011). Bight 13 is extracted from (McLaughlin et al., 2021) study.

708 4.5 Carbonate system and oxygen parameters

709 The model predicts changes in dissolved oxygen and carbon-system parameters related
710 to photosynthesis and respiration, as well as horizontal transport and vertical mixing. As
711 described in section 4.4.1, the coasts of Los Angeles and Santa Barbara are hot-spots of
712 intensified plankton activity, and both systems are impacted by high variability and small-
713 scale eddy circulation. In the upper layers, photosynthesis increases both dissolved oxygen
714 and pH (Figs. 16 and 18), consistent with observations in these regions. The Santa Monica
715 Bay shows the highest oxygen production rates ($60 \text{ mmol m}^{-2} \text{ d}^{-1}$), followed by the Santa
716 Barbara coast ($57 \text{ mmol m}^{-2} \text{ d}^{-1}$), while rates in the Orange County and San Diego coasts
717 are nearly two times lower. Oxygen and carbon are further replenished at the surface by
718 air-sea gas exchange with the atmosphere. Export of newly-fixed organic carbon leads in
719 both regions to high remineralization rates that consume oxygen and release carbon dioxide
720 at depth. We simulate similar high organic matter export (around $30 \text{ mmol m}^{-2} \text{ d}^{-1}$) in
721 the Santa Barbara and Los Angeles coasts (see Supporting Information: Fig. S24).

722 The reliability of these predictions can be tested by validation of dissolved oxygen and
723 carbonate system parameters. There are several sources of uncertainty in the dissolved
724 oxygen, pH, and aragonite saturation state observational records, which affect data-model
725 comparisons. For dissolved oxygen, sensors are relatively accurate and precise and have a
726 rapid response time ($< 1\text{s}$) when generating vertically resolved profiles. Repeated field
727 measurement accuracy for CTD dissolved oxygen sensors was reported to be approximately 8
728 mmol m^{-3} (Coppola et al., 2013). The pH observational record is particularly fraught with
729 uncertainty. An evaluation of pH sensor data in the SCB indicated that, while sensor pH
730 measurements were well correlated with discrete bottle samples collected at the same depth,
731 there was a clear bias in pH, with sensor measurements under-predicting bottle measure-
732 ments and high variability in the differences between paired bottle and sensor measurements
733 (ΔpH ranging from ± 0.5) (McLaughlin, Dickson, et al., 2017). The aragonite saturation
734 state is estimated using an algorithm developed for the region (Juraneck et al., 2011) for both
735 *in situ* observations and model output, because complete measurements of carbon-system
736 parameters required to calculate Ω_{Ar} are missing. For all three variables, spatial and tem-

737 poral under-sampling, particularly during seasons with high variability, adds uncertainty to
 738 the data-model comparison.

739 **4.5.1 Vertical gradients and seasonal variability of dissolved oxygen**

740 The model reproduces observed seasonal and spatial patterns in dissolved oxygen con-
 741 centration (Fig. 15), accurately simulating magnitude, vertical and horizontal gradients,
 742 and variability. Quantitative statistical analysis (see Table 6) indicated that the model per-
 743 formance was ‘*excellent*’ or ‘*good*’ for nearly all metrics for all regions and seasons. The
 744 lowest performance of the model was characterized as ‘*poor*’ for two sub-regions for the
 745 Nash-Sutcliff Model Efficiency during Spring, and ‘*reasonable*’ for some metrics in some
 746 sub-regions, which may be related to under-sampling during seasons with high variability,
 747 as described above. Similar to temperature, we tested whether the variability in spring
 748 may be impacting the performance statistics by extracting random profiles for the region
 749 (not shown, expressed with large error-bars in the spring season plots in Fig. 16), which
 750 show how dissolved oxygen on a single arbitrary day can more closely align with the ob-
 751 servations. This supports the hypothesis that observational uncertainty is behind the lack
 752 of observational agreement with the model. Model performance was lowest in the Orange
 753 County and San Diego subregions, where model predictions tended to overestimate dissolved
 754 oxygen, consistent with the chlorophyll under-prediction, a likely consequence of the lack of
 755 cross-border inputs from Mexican waters.

756 The model also reproduces the seasonality in dissolved oxygen in all subregions (Fig.
 757 16), characterized by large meridional and vertical variability. Near the Channel Islands,
 758 dissolved oxygen varies at 50 m by up to 140 mmol O₂ m⁻³ between the highest winter
 759 values and the lowest summer values, reflecting the dynamics of upwelling, productivity, and
 760 air-sea gas-exchange. Offshore the coasts of Santa Monica and San Diego, the variability
 761 between winter and summer is of the order of 80-90 mmol O₂ m⁻³. Surface concentrations
 762 are everywhere above 240 mmol O₂ m⁻³ year-round, consistent with observations. The
 763 highest summer concentrations are observed at the depth of the deep chlorophyll maximum,
 764 reflecting photosynthesis, while decreasing at depth to below 150 mmol O₂ m⁻³. These
 765 patterns are generally consistent with observations in the same regions.

766 During the 1998 El Niño event, the model shows a net decrease of dissolved oxygen near
 767 the surface, and a net increase below it. During this period, the entire upper layer (0-80
 768 m) is characterized by a homogeneous oxygen concentration of about 240 mmol O₂ m⁻³
 769 over almost the entire SCB (not shown). Only the San Pedro and Santa Monica Bays show
 770 higher concentrations, which we attribute to the local anthropogenic nutrient enrichment
 771 and subsequent blooms (see Fig. 19). This is consistent with observations of the 1998 El
 772 Niño event in California coastal waters (Chavez et al., 2002; Booth et al., 2014).

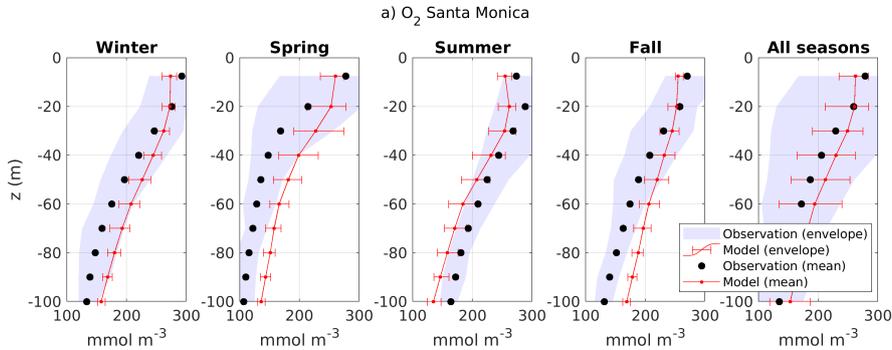


Figure 15: As for Fig. 5, but for oxygen concentration.

Santa Monica								
	H	Correlation Coefficient	p-value	Cost Function	Oxygen Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.97 E	9E-07 E	0.14 E	-0.09 E	1.20 G	0.77 E	716
Spring	0 E	0.91 E	3E-04 E	0.26 E	-0.23 R	1.03 E	0.37 R	702
Summer	0 E	0.99 E	2E-10 E	0.07 E	0.07 E	0.99 E	0.86 E	712
Fall	0 E	0.97 E	2E-06 E	0.19 E	-0.14 G	1.49 P	0.42 R	718
All Seasons	0 E	0.97 E	3E-06 E	0.14 E	-0.11 G	1.18 G	0.69 E	2848

pH								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.99 E	2E-08 E	0.01 E	0.01 E	0.59 P	0.57 G	632
Spring	0 E	0.97 E	2E-06 E	0.02 E	-0.02 E	1.45 P	0.15 P	702
Summer	0 E	0.96 E	9E-06 E	0.01 E	0.01 E	1.01 E	0.85 E	712
Fall	0 E	0.97 E	3E-06 E	0.01 E	0.01 E	1.49 P	0.78 E	715
All Seasons	0 E	0.97 E	5E-06 E	0.01 E	-0.01 E	1.12 G	0.84 E	2761

Table 6: Statistical comparison between *in situ* data and model outputs for dissolved oxygen and pH profile in Santa Monica Bay.

773 **4.5.2 Vertical gradients and seasonal variability of carbon-system parameters**
774

775 Together with pH, the saturation state of aragonite (Ω_{Ar}) is often used as a metric
776 to identify the potential impact of Ocean Acidification on marine calcifiers, because it is a
777 measure of the availability of carbonate ions for calcium carbonate precipitation (Bednarsek
778 et al., 2019). Ω_{Ar} shows similar vertical variability as dissolved oxygen (Juraneck et al., 2009;
779 Alin et al., 2012). Similar to oxygen loss, reduction in pH and Ω_{Ar} in the upper layers is
780 generally caused by coastal upwelling or by local physical processes (Feely et al., 2018). We
781 utilize sensor pH data sets to evaluate vertical profiles in the carbonate system. Because of
782 the known uncertainty in pH measurements, we are most concerned with how well the model
783 reproduced the shape of the profiles (i.e., goodness of fit estimates, as with chlorophyll).
784 Sensor-derived pH profile measurements should be internally consistent within a data set
785 (if the sensor is working properly and if pressure issues are minimal), providing some value
786 to goodness of fit assessments. Given these constraints, the data-model comparisons for
787 pH sensor data were generally ‘*excellent*’ or ‘*good*’ for all sub-regions and all seasons. Un-
788 surprisingly, the model performance reproducing observational means and variability was
789 generally ‘*reasonable*’ or ‘*poor*’ for most sub-regions and seasons, with some, if not most, of
790 this disagreement due to difficulties in conducting a validation of the model with large un-
791 certainties in sensor-derived pH profiles. Recently, the CalCOFI program has incorporated
792 Ω_{Ar} into its sampling design. Although the data do not line up with the model period, they
793 are useful for evaluating seasonal variability in the model. Generally, the model reproduces
794 seasonal and vertical variability in Ω_{Ar} , with higher saturation states in the summer and
795 fall, when waters are generally more stratified, and lower values in winter and spring, when
796 upwelling brings undersaturated waters closer to the surface. Ω_{Ar} is also much lower and
797 more highly variable at depth. These patterns are consistent with observations throughout
798 the SCB (McLaughlin et al., 2018).

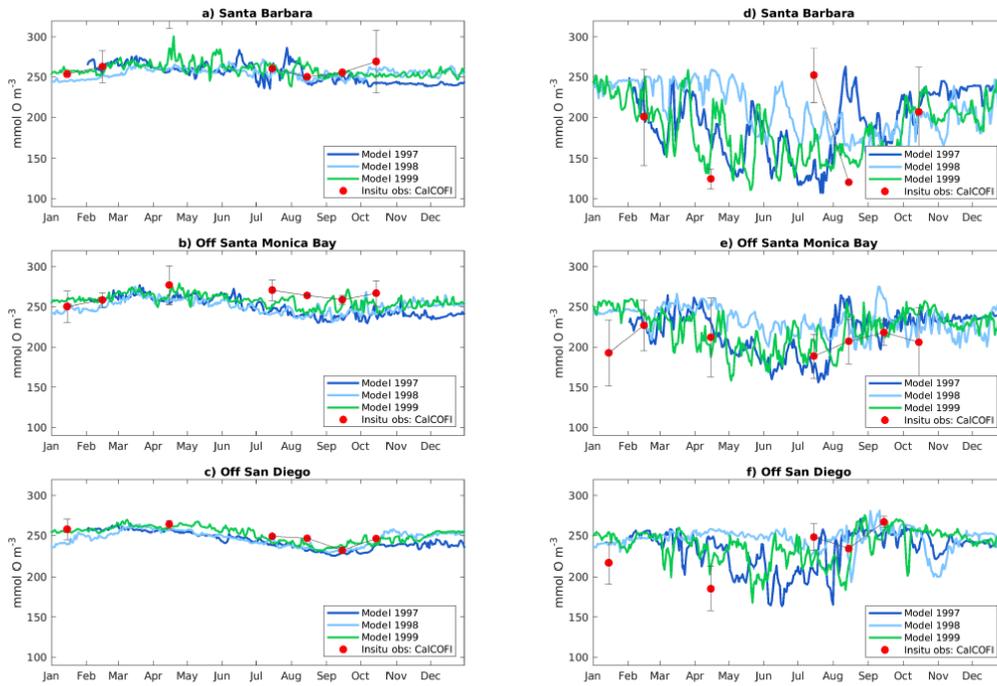


Figure 16: Comparison of dissolved oxygen concentration between different years of model output, and a climatology from CalCOFI *in situ* data. SB is near the center of Santa Barbara Channel, SM is offshore the Santa Monica Bay, and SD is offshore San Diego. Left panels show surface concentrations, right panels concentrations at 50 m depth.

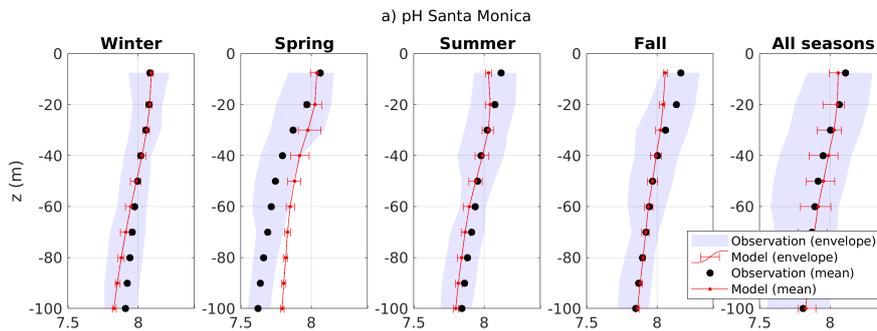


Figure 17: As for Fig. 5 but for dissolved pH.

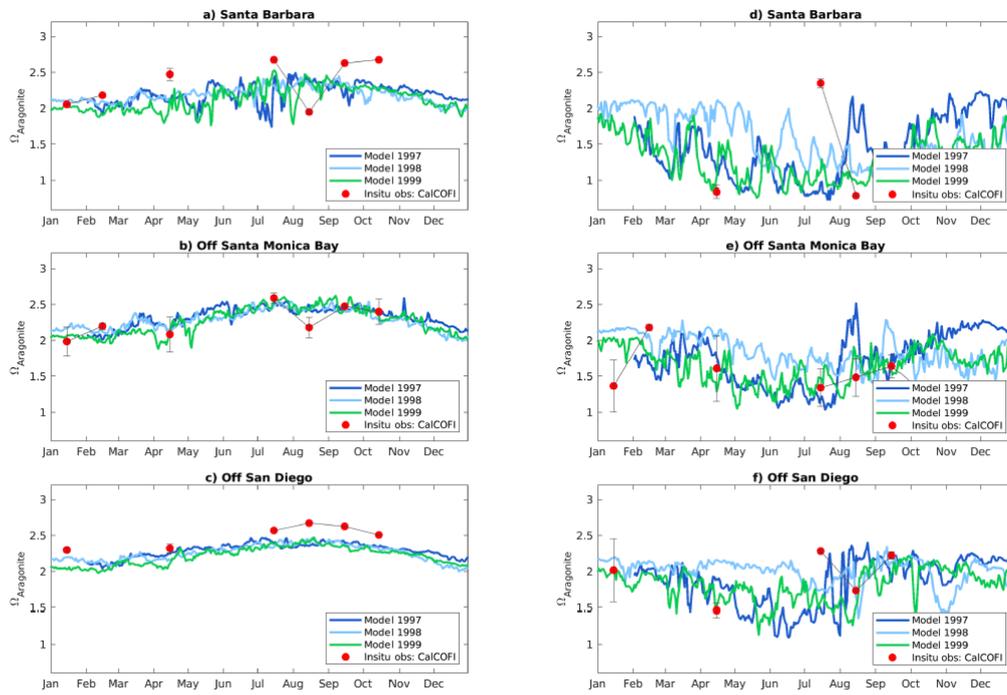


Figure 18: Comparison of the saturation state of aragonite between different years of model output, and a climatology from CalCOFI *in situ* data. SB is near the center of Santa Barbara Channel, SM is offshore the Santa Monica Bay, and SD is offshore San Diego. Left panels show surface values, right panels values at 50 m depth.

799 **5 Summary**

800 In this study, we demonstrated the readiness of a high-resolution, dynamically down-
801 scaled, physical-biogeochemical model to mechanistically investigate links between a com-
802 prehensive reconstruction of terrestrial and atmospheric nutrient inputs, coastal eutrophica-
803 tion, and biogeochemical change in the SCB coastal waters. This modeling platform is
804 an important achievement because it strikes a balance of capturing the forcing of coast-
805 wide basin mesoscale phenomena, while capturing the combined effects of bathymetry and
806 submesoscale eddies that intensify transport of nutrients and biological material. Moreover,
807 this model allows to run hindcast simulations of primary production, ocean acidification and
808 oxygen loss at timescales that can approach the multi-annual frequencies of intrinsic ocean
809 variability, making the grand challenge of disentangling natural variability, climate change,
810 and local anthropogenic forcing a tractable task in the near-term.

811 ROMS has a long history of validation and management acceptance through various
812 applications in the CCS (e.g. Marchesiello et al. (2003); Capet et al. (2004); Capet, Colas, et
813 al. (2008); Capet, Campos, and Paiva (2008); Capet, McWilliams, et al. (2008); Shchepetkin
814 and McWilliams (2011); Renault, Molemaker, Gula, et al. (2016)). In contrast, experience
815 with BEC within the SCB is more limited. Our validation study of coastal eutrophication
816 gradients in the SCB nearshore complements the U.S. West Coast-wide study of (Deutsch
817 et al., 2020) and strengthens confidence that the basic CCS BEC model formulation, forcing
818 and parameterization is appropriate not only for coastwide analyses, but also for detailed
819 local studies of coastal eutrophication in the highly urbanized SCB. The representation of
820 physical processes such as vertical mixing and horizontal circulation was consistent across
821 the model and measurements. The model reproduces the main structure of the climato-
822 logical upwelling front and cross-shore isopycnal slopes, and the mean current patterns and
823 associated temperature gradients. We also demonstrate good agreement between model
824 simulations and the mean distributions and variability of key ecosystem metrics, including
825 surface nutrients and productivity, and subsurface O₂ and carbonate saturation. The spatial
826 patterns of primary production, phytoplankton growth rates, and zooplankton grazing are
827 broadly consistent with measured rates. The distribution of primary production is governed
828 by the trade-off between nutrient and light limitation, a balance that reproduces and explains
829 the observed spatial variations in the depth of the deep chlorophyll maximum. Statistical
830 measures of model agreement on biogeochemical state variables was excellent to good and
831 the range of predicted biogeochemical rates on par with observations. Under the realistic
832 flow fields produced by ROMS, the conformity of model predictions with a rich observational
833 dataset is a strong demonstration of model validity for coastal eutrophication applications.
834 We also demonstrated that the model responds with confidence to the variability caused by
835 El Niño, modifying the vertical distribution of the physical and biogeochemical properties
836 across the upper ocean of the entire Bight, as illustrated by the three-dimensional change
837 in key ocean variables shown in Fig. 19.

838 While the broad agreement between the model and observations for a range of variables
839 is encouraging, there remain aspects of the model that require further study and improve-
840 ment. For example, phytoplankton diversity is limited in the model, preventing it from
841 properly simulating events such as dinoflagellate-driven red tides, which occur over short
842 periods on limited coastal scales in the spring. Despite the good performance of the model
843 in reproducing total primary production and grazing rates, the model does not include mul-
844 tiple zooplankton functional groups, thus providing little information on the dynamics and
845 transfer of energy of higher trophic levels. From a hydrodynamics point of view, with a hor-
846 izontal resolution of 300m, the model does not directly resolve physical processes occurring
847 at scales of tens of meters (Dauhajre et al., 2019), for example the dilution and entrainment
848 of buoyant wastewater plumes, which is now parameterized, or the vertical and horizontal
849 transport of tracers in the very nearshore surf zone.

850 Quantitative and qualitative results of confidence assessments are essential for informing
851 management decisions, evaluating management strategies, and providing a basis for risk

852 analyses. The most successful management approaches are those that explicitly incorporate
 853 uncertainty (e.g. Taylor et al. (2000)). An assessment of model validation must consider
 854 the complex combination of model and observational uncertainties (Allen et al., 2007),
 855 including: 1) uncertainty/error in the model, with the inclusion of intrinsic variability; 2)
 856 uncertainty/error in measured data; 3) uncertainty from the difference in spatial scales of the
 857 model output relative to the measured data used in the comparison (specifically, comparing
 858 a 0.3 km grid cell to a discrete sampling station); and 4) uncertainty from the difference in
 859 temporal averaging of the model output relative to the measured data. For parameters in
 860 which we have high confidence in the observational record, i.e., temperature and dissolved
 861 oxygen, model performance statistics show excellent agreement for mean profiles, vertical
 862 and horizontal gradients, as well as seasonal variability. The model reproduces chlorophyll
 863 reasonably well, albeit with some biases, which can be in part attributed to a simplified
 864 representation of plankton diversity, measurement uncertainty, sparseness of *in situ* data,
 865 cloud cover and algorithm biases in satellite products. Variables such as pH and ammonium
 866 show lower agreement, likely due to measurement uncertainty and sampling bias, but general
 867 spatial and temporal patterns are correctly reproduced in the model.

868 Greater clarity is needed in the requirements for model performance and uncertainty to
 869 support decisions on management of SCB coastal water quality and eutrophication (Boesch,
 870 2019). These requirements are likely to be driven largely by the approach that will be used
 871 to interpret a "significant impact" (e.g. existing water quality pH and dissolved oxygen
 872 criteria, or biologically relevant thresholds; (Weisberg et al., 2016)), as these have signifi-
 873 cant implications for required model precision and accuracy on different spatial and
 874 temporal scales. Future efforts to constrain uncertainty could include sensitivity analyses
 875 and model ensemble comparisons of BEC with other biogeochemical models that feature
 876 increasingly complex representations of planktonic functional groups, benthic communities,
 877 and sediment-pelagic interactions. Finally, long-term investments are needed in coupled
 878 chemical-biological observations of phytoplankton and zooplankton diversity and community
 879 structure. These observations are critical to provide understanding of the evolution of lower
 880 trophic ecosystem structure with climate change, and their relationship with biogeochemical
 881 cycles linked to ocean acidification and oxygen loss (Sailley et al., 2013). Ultimately, the
 882 need to constrain uncertainty will likely scale with the economic import of management
 883 decisions under consideration, which could range from increased monitoring requirements
 884 to multi-billion dollar non-point source controls and wastewater treatment plant upgrades.

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 892 et al., 2020). In situ observation data to generate the figures and statistics are available in
 893 (Kessouri, McLaughlin, et al., 2020). Local land-based and atmospheric data can be found
 894 in (Sutula et al., 2021a).

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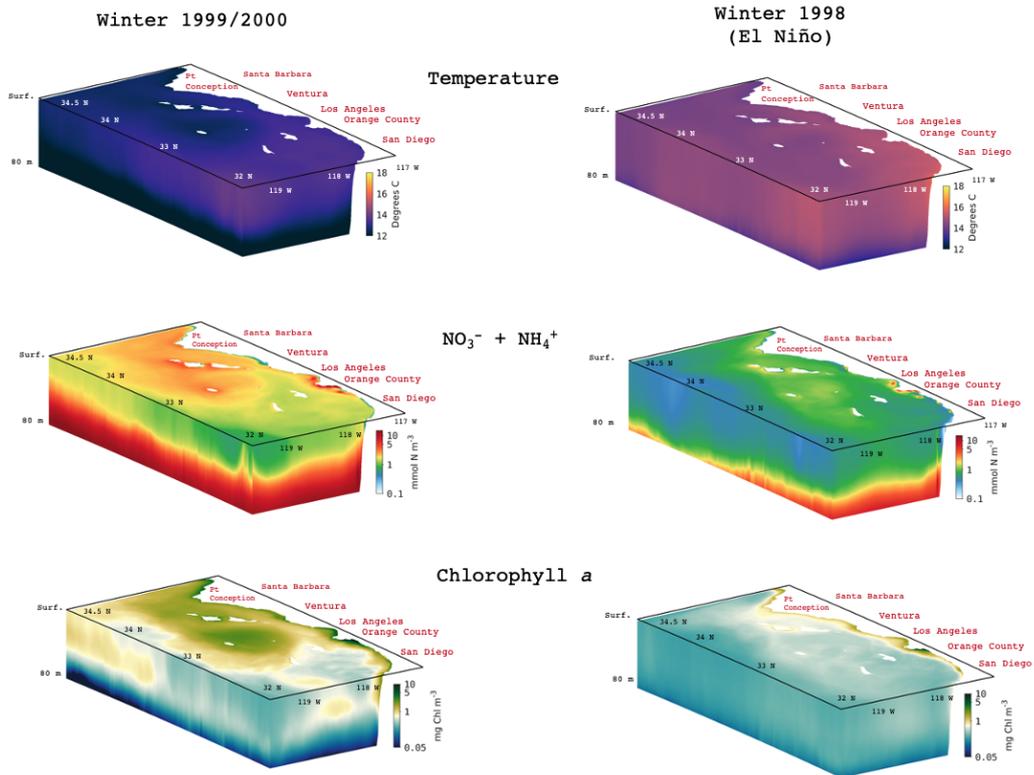


Figure 19: Three-dimensional illustration of temperature, DIN ($\text{NO}_3^- + \text{NH}_4^+$) and chlorophyll in the Southern California Bight. Panels show winter 1999 and 2000 (left panels), winter 1998 during El Niño period (right panels).

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Supporting Information for “Configuration and validation of an oceanic physical and biogeochemical model to investigate coastal eutrophication in the Southern California Bight”

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- 1296 coasts.

Ventura/Oxnard								
Temperature								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.95 E	2E-05 E	0.06 E	-0.04 E	1.11 G	0.77 E	469
Spring	0 E	0.98 E	7E-07 E	0.09 E	-0.09 E	0.80 G	0.59 G	468
Summer	0 E	0.98 E	1E-06 E	0.04 E	-0.02 E	1.08 E	0.94 E	468
Fall	0 E	0.89 G	5E-04 E	0.09 E	-0.08 E	0.97 E	0.50 R	469
All Seasons	0 E	0.95 E	3E-05 E	0.08 E	-0.06 E	1.03 E	0.74 E	1874
Oxygen								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.98 E	1E-06 E	0.14 E	-0.09 E	1.20 R	0.77 E	469
Spring	0 E	0.92 E	1E-04 E	0.25 E	-0.21 R	1.06 E	0.47 R	454
Summer	0 E	0.99 E	1E-09 E	0.07 E	0.08 E	1.03 E	0.84 E	468
Fall	0 E	0.98 E	1E-06 E	0.19 E	-0.14 G	1.52 P	0.44 R	469
All Seasons	0 E	0.98 E	1E-06 E	0.14 E	-0.010 G	1.20 R	0.73 E	1860
Chlorophyll-a								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.99 E	1E-08 E	0.43 E	-0.06 E	0.90 E	0.96 E	469
Spring	0 E	0.97 E	3E-06 E	0.86 E	-0.42 P	0.51 P	-0.47 P	468
Summer	0 E	0.96 E	1E-05 E	0.59 E	-0.04 E	0.64 R	0.51 G	468
Fall	0 E	0.94 E	5E-05 E	0.53 E	0.20 R	0.71 R	0.50 R	469
All Seasons	0 E	0.99 E	9E-11 E	0.50 E	-0.01 E	0.71 R	0.79 E	1874
pH								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.99 E	2E-08 E	0.01 E	0.01 E	0.62 R	0.57 G	455
Spring	0 E	0.97 E	2E-07 E	0.01 E	-0.01 E	1.45 P	0.25 R	454
Summer	0 E	0.97 E	2E-06 E	0.01 E	0.01 E	1.06 E	0.84 E	468
Fall	0 E	0.98 E	6E-07 E	0.01 E	0.01 E	1.43 P	0.75 E	467
All Seasons	0 E	0.98 E	1E-06 E	0.01 E	-0.01 E	1.13 G	0.89 E	1844
Ammonia								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.86 G	0.34 P	0.53 E	0.24 R	1.56 P	0.58 G	11
Spring	0 E	0.99 E	0.02 E	1.40 G	-1.87 P	0.36 P	-10.52 P	12
Summer	0 E	0.92 E	0.25 P	2.59 R	0.28 R	1.62 P	0.67 E	12
Fall	0 E	0.92 E	0.26 P	4.42 P	-2.77 P	0.35 P	-10.48 P	12
All Seasons	0 E	0.89 G	0.04 E	0.77 E	-0.49 P	0.91 E	0.01 P	47

Table S1: Statistical comparison of vertical profiles of temperature, dissolved oxygen, chlorophyll *a*, pH, and ammonium concentration at Ventura/Oxnard (City of Oxnard stations) monitoring region. Letters next to numbers indicate model performance: E = Excellent, G = very good, R = reasonable, P = Poor.

Palos Verdes								
Temperature								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.95 E	2E-05 E	0.04 E	-0.01 E	1.13 G	0.87 E	469
Spring	0 E	0.94 E	6E-05 E	0.24 E	-0.11 G	0.75 R	0.19 P	466
Summer	0 E	0.98 E	7E-07 E	0.03 E	0.03 E	1.19 G	0.91 E	466
Fall	0 E	0.88 G	7E-04 E	0.11 E	-0.11 G	0.75 R	-0.27 P	468
All Seasons	0 E	0.94 E	5E-05 E	0.07 E	-0.06 E	0.98 E	0.74 E	1869
Oxygen								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.99 E	9E-09 E	0.03 E	0.02 E	1.05 E	0.94 E	469
Spring	0 E	0.83 G	3E-03 E	0.24 E	-0.22 R	0.87 G	0.01 P	466
Summer	0 E	0.97 E	2E-06 E	0.08 E	0.08 E	0.87 G	0.69 E	466
Fall	0 E	0.97 E	7E-04 E	0.08 E	-0.06 E	1.14 G	0.76 E	467
All Seasons	0 E	0.98 E	1E-06 E	0.06 E	-0.05 E	0.97 E	0.88 E	1868
Chlorophyll-a								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.75 R	0.01 E	0.70 E	0.25 R	0.83 G	0.19 P	469
Spring	0 E	0.83 G	2E-03 E	0.72 E	0.06 E	1.01 E	0.65 E	466
Summer	0 E	0.67 R	0.04 E	0.73 E	0.38 R	1.28 R	0.17 P	466
Fall	0 E	0.99 E	1E-10 E	0.48 E	0.25 R	1.01 E	0.91 E	468
All Seasons	0 E	0.95 E	2E-05 E	0.55 E	0.30 R	1.00 E	0.70 E	1869
pH								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	1 P	0.99 E	2E-07 E	0.03 E	-0.03 E	1.36 R	-2.96 P	469
Spring	0 E	0.84 G	2E-03 E	0.01 E	0.01 E	1.44 P	0.67 E	466
Summer	1 P	0.99 E	2E-09 E	0.03 E	0.03 E	1.62 P	-2.80 P	466
Fall	1 P	0.96 E	9E-06 E	0.02 E	0.02 E	1.54 P	-1.29 P	468
All Seasons	0 E	0.96 E	1E-05 E	0.01 E	0.01 E	1.50 P	0.73 E	1869
Ammonia								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0
Spring	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0
Summer	0 E	0.82 G	0.18 P	0.31 E	0.15 G	1.05 E	0.53 G	18
Fall	0 E	0.79 R	0.21 P	0.30 E	0.32 R	1.38 R	-0.15 P	18
All Seasons	0 E	0.84 G	0.16 P	0.30 E	0.32 R	1.34 R	-0.01 P	36

Table S2: Same as Table S1 for Palos Verdes (LACSD stations) monitoring region.

Orange County								
Temperature								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.96 E	1E-05 E	0.05 E	-0.01 E	1.27 R	0.84 E	160
Spring	0 E	0.95 E	3E-05 E	0.11 E	-0.11 G	0.81 G	0.44 R	533
Summer	0 E	0.99 E	2E-08 E	0.02 E	-0.01 E	0.96 E	0.98 E	533
Fall	0 E	0.92 E	2E-04 E	0.08 E	-0.07 E	1.06 E	0.66 E	536
All Seasons	0 E	0.95 E	3E-05 E	0.07 E	-0.05 E	1.07 E	0.79 E	1762
Oxygen								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.98 E	1E-06 E	0.14 E	-0.12 G	1.03 E	0.67 E	150
Spring	0 E	0.90 E	4E-04 E	0.19 E	-0.16 G	0.96 E	0.51 G	533
Summer	0 E	0.99 E	5E-08 E	0.07 E	0.07 E	0.91 E	0.80 E	534
Fall	0 E	0.92 E	4E-07 E	0.09 E	-0.02 E	1.06 E	0.80 E	536
All Seasons	0 E	0.95 E	9E-06 E	0.09 E	-0.06 E	1.07 E	0.81 E	1753
Chlorophyll-a								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.98 E	8E-07 E	0.97 E	0.53 P	2.25 P	0.49 R	160
Spring	0 E	0.74 R	0.02 E	2.24 R	0.29 R	2.12 P	0.46 R	533
Summer	0 E	0.94 E	7E-05 E	0.57 E	-0.16 G	0.83 G	0.76 E	535
Fall	0 E	0.92 E	1E-04 E	0.55 E	0.41 P	1.32 R	0.62 G	536
All Seasons	0 E	0.91 E	3E-04 E	0.47 E	0.33 R	1.80 P	0.63 G	1764
pH								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.98 E	1E-06 E	0.01 E	0 E	1.26 R	0.88 E	160
Spring	0 E	0.79 R	0.01 E	0.01 E	-0.01 E	1.32 R	0.28 R	533
Summer	1 P	0.96 E	2E-05 E	0.02 E	0.02 E	1.32 R	-1.13 P	534
Fall	1 P	0.98 E	1E-06 E	0.03 E	-0.03 E	2.02 P	-2.62 P	536
All Seasons	0 E	0.93 E	9E-05 E	0.01 E	-0.01 E	1.43 P	0.51 G	1763
Ammonia								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0
Spring	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0
Summer	1 P	0.97 E	0.14 P	0.86 E	0.83 P	3.89 P	-2.38 P	48
Fall	0 E	0.71 R	0.50 P	0.43 E	0.37 R	0.44 P	-8.71 P	48
All Seasons	1 P	0.76 R	0.14 P	0.62 E	0.63 P	2.36 P	-1.67 P	96

Table S3: Same as Table S1 for Orange County (OCSD stations) monitoring region.

San Diego								
Temperature								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.98 E	3E-08 E	0.02 E	-0.01 E	0.92 E	0.95 E	875
Spring	0 E	0.93 E	1E-05 E	0.09 E	-0.09 E	0.77 R	0.30 R	870
Summer	0 E	0.98 E	5E-08 E	0.04 E	-0.01 E	0.99 E	0.94 E	872
Fall	0 E	0.92 E	3E-05 E	0.08 E	-0.08 E	0.79 R	0.29 R	752
All Seasons	0 E	0.98 E	4E-07 E	0.05 E	-0.05 E	0.83 G	0.70 E	3369
Oxygen								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.96 E	1E-06 E	0.11 E	-0.09 E	1.05 E	0.74 E	875
Spring	1 P	0.87 G	3E-04 E	0.33 E	-0.32 R	0.87 G	-0.45 P	870
Summer	0 E	0.99 E	5E-12 E	0.27 E	-0.20 R	1.47 P	0.51 G	872
Fall	0 E	0.92 E	3E-05 E	0.22 E	-0.16 G	1.59 P	0.37 R	752
All Seasons	0 E	0.97 E	4E-07 E	0.18 E	-0.16 G	1.18 G	0.55 G	3369
Chlorophyll-a								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.99 E	6E-09 E	0.51 E	0.60 P	2.59 P	0.39 R	868
Spring	0 E	0.94 E	4E-06 E	0.84 E	0.36 R	1.59 P	0.73 E	866
Summer	0 E	0.99 E	6E-09 E	0.27 E	0.18 G	1.28 R	0.90 E	870
Fall	0 E	0.89 G	2E-04 E	0.35 E	0.43 P	2.05 P	0.50 R	728
All Seasons	0 E	0.98 E	9E-09 E	0.57 E	0.39 R	1.66 P	0.70 E	3332
pH								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	0 E	0.99 E	5E-09 E	0.01 E	-0.01 E	1.07 E	0.49 R	875
Spring	1 P	0.91 E	5E-05 E	0.02 E	-0.02 E	1.41 P	-0.96 P	872
Summer	1 P	0.99 E	1E-09 E	0.02 E	-0.02 E	2.32 P	-0.07 P	844
Fall	1 P	0.98 E	2E-04 E	0.01 E	-0.01 E	1.77 P	-0.14 P	752
All Seasons	1 P	0.98 E	6E-08 E	0.02 E	-0.02 E	1.59 P	-0.18 P	3343
Ammonia								
	H	Correlation Coefficient	p-value	Cost Function	Percentage Bias	Ratio of Standard Deviations	Nash-Sutcliffe Model Efficiency	Number of observations
Winter	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0
Spring	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0
Summer	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0
Fall	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0
All Seasons	N/A	N/A	N/A	N/A	N/A	N/A	N/A	0

Table S4: Same as Table S1 for San Diego (City of San Diego stations) monitoring region.

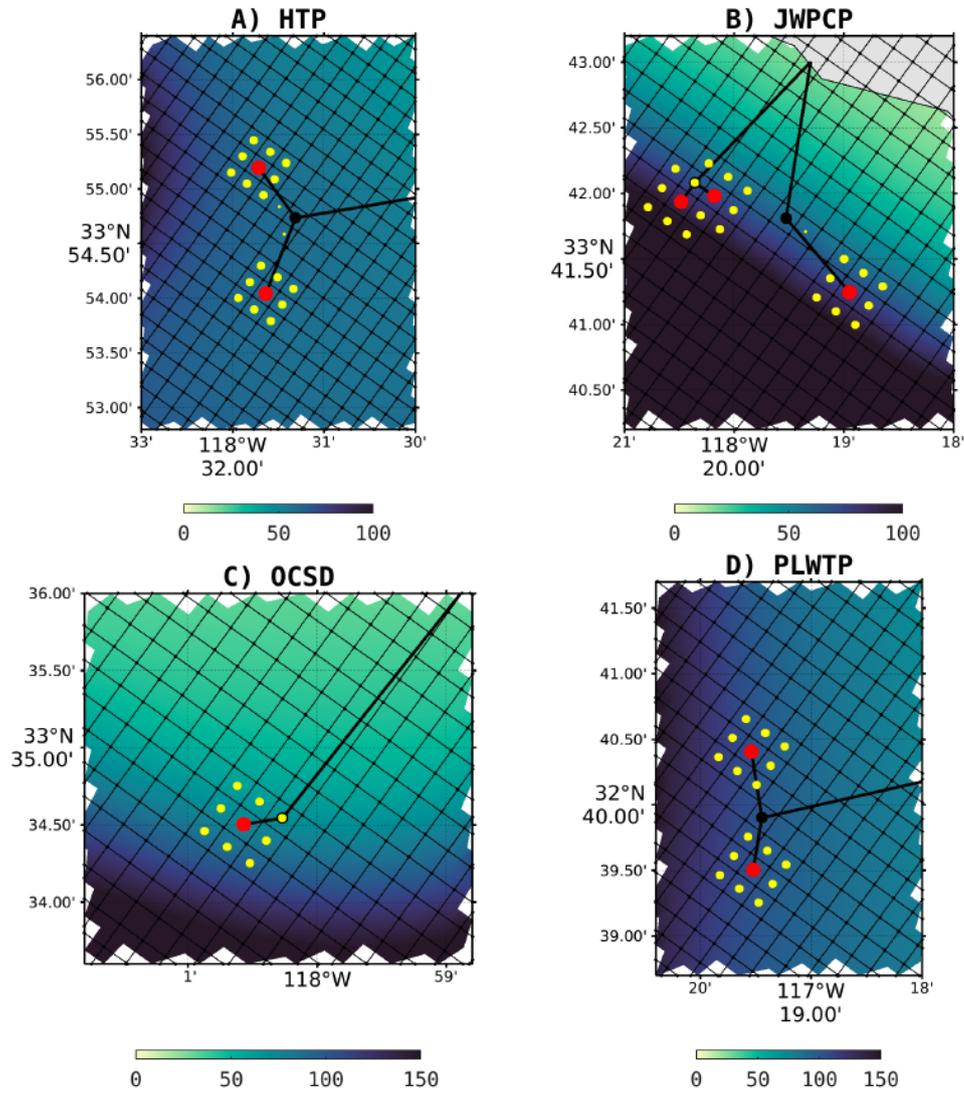


Figure S1: Spatial distribution of the point sources to simulate and to dilute the freshwater, nutrients and organic matter fluxes for the 4 majors POTW underwater outfalls locations. Color scale is bathymetry. Vertically integrated, the grid cells with the red dots discharge $4/9$ of the respective flow at each diffuser, the grid cells with yellow dots north, south, east and west of the red dots discharge $1/9$ of the discharge, and the yellow dots in the corners discharge $1/36$ of the volume flux.

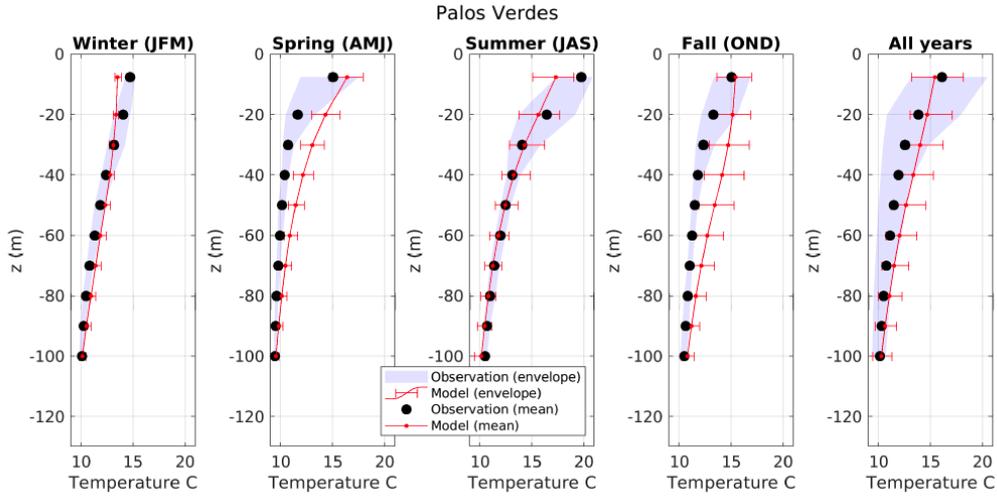


Figure S2: Seasonal profiles of average temperature off of Palos Verdes. The red line and red bars are the spatial and temporal means and the variability from the model. The black dots and the gray shade are the spatial and temporal mean and the variability from *in situ* data (LACSD stations). These profiles are showing agreement on intensity, seasonality and shape of the vertical profile with exceptionally high concentrations at mid-depth.

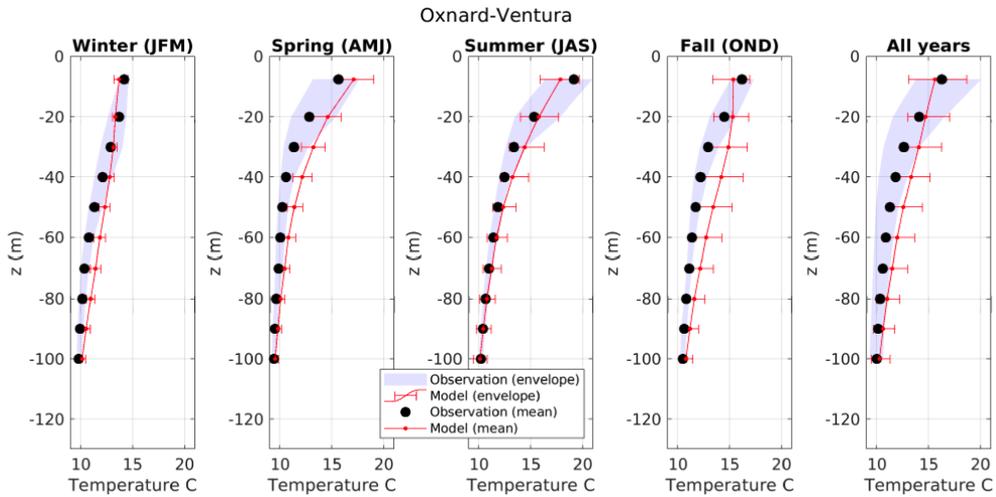


Figure S3: Same as Fig S2 for Oxnard/Ventura using City of Oxnard stations.

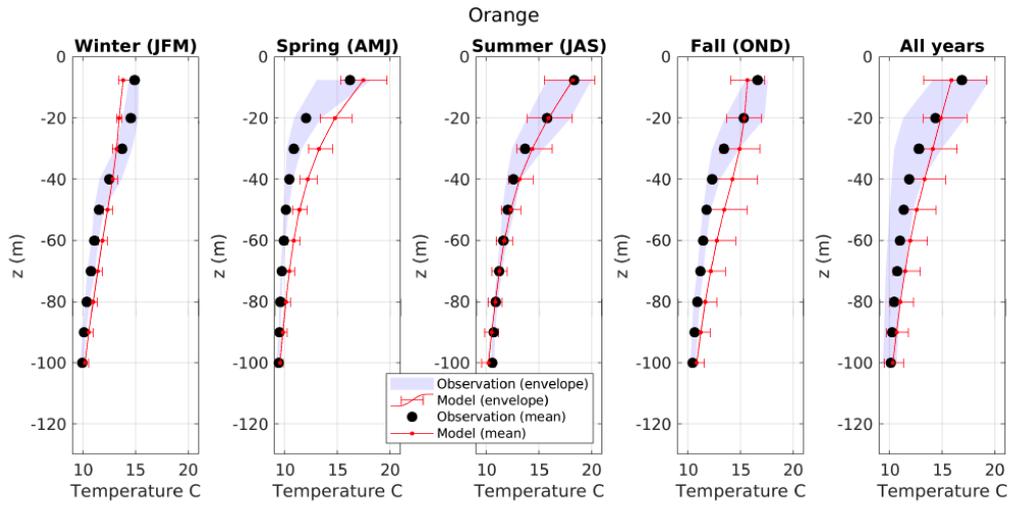


Figure S4: Same as Fig S2 for Orange County using OCSF stations.

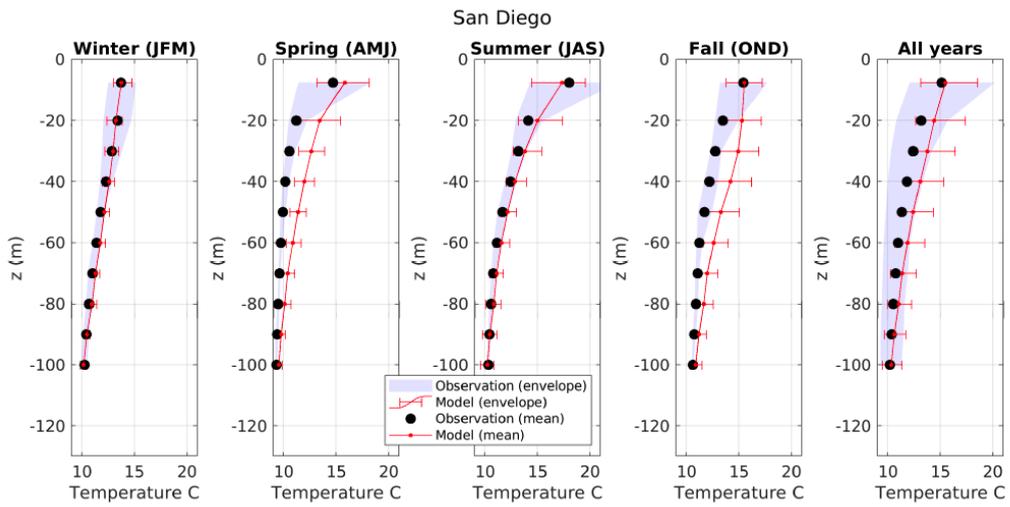


Figure S5: Same as Fig S2 for San Diego using City of San Diego stations.

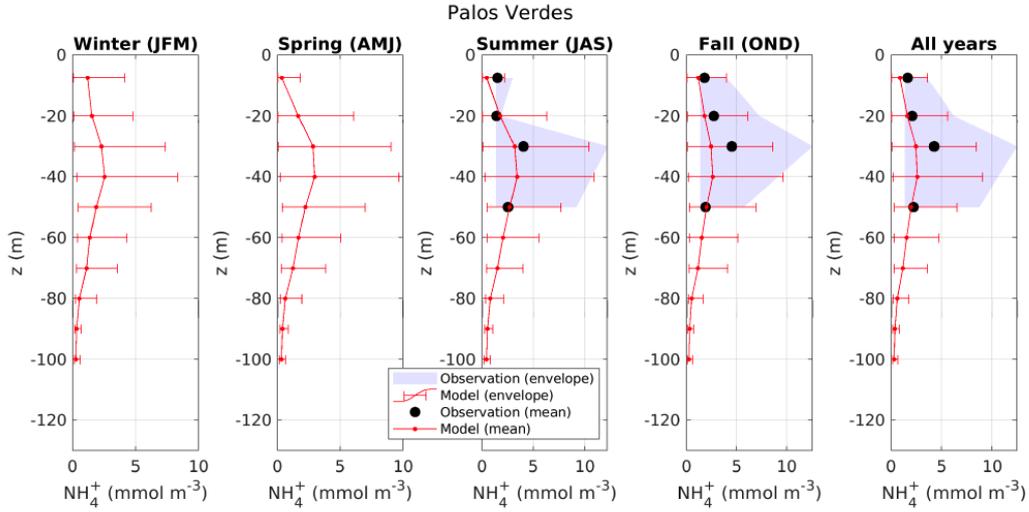


Figure S6: Seasonal profiles of average ammonium concentration off of Palos Verdes. The red line and red bars are the spatial and temporal means and the variability from the model. The black dots and the gray shade are the spatial and temporal mean and the variability from *in situ* data (LACSD stations). These profiles are showing agreement on intensity, seasonality and shape of the vertical profile with exceptionally high concentrations at mid-depth.

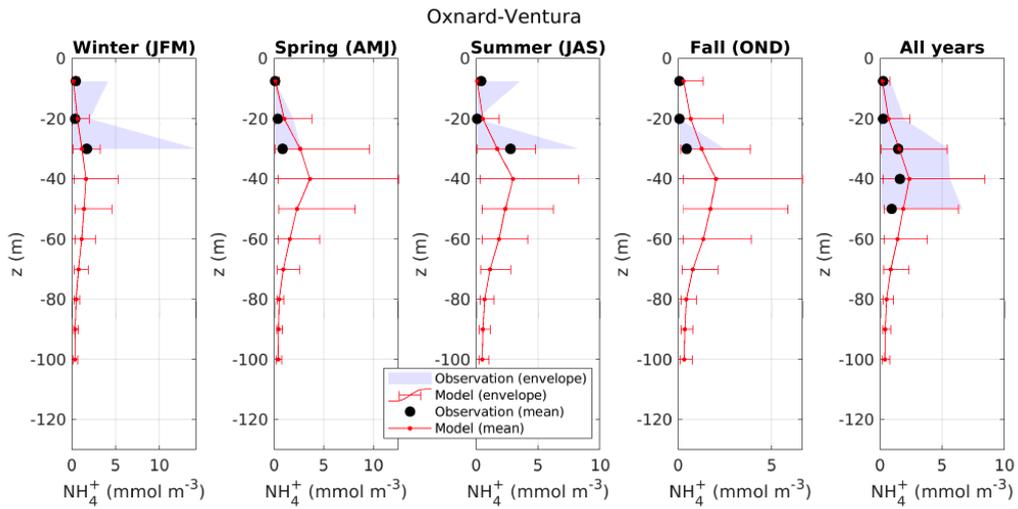


Figure S7: Same as Fig S6 for Oxnard/Ventura using City of Oxnard stations

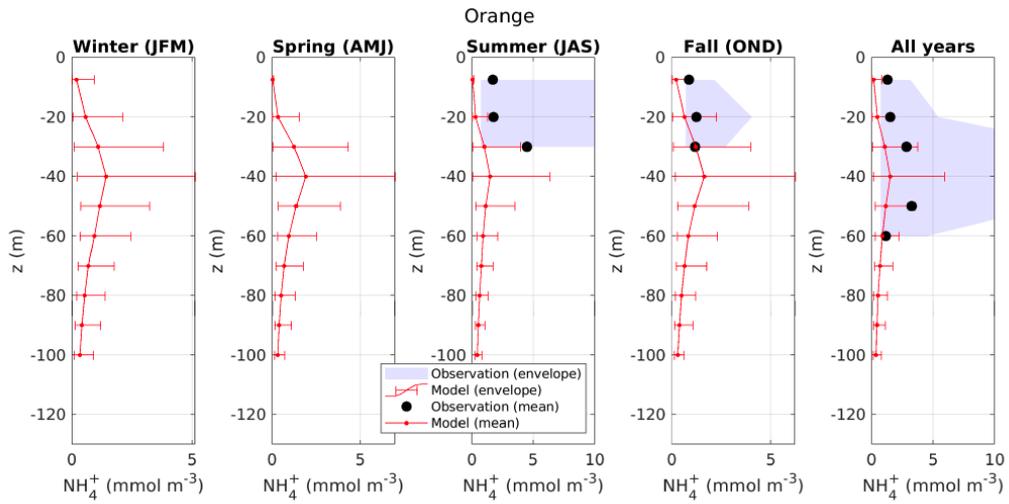


Figure S8: Same as Fig S6 for Orange County using OCSD stations.

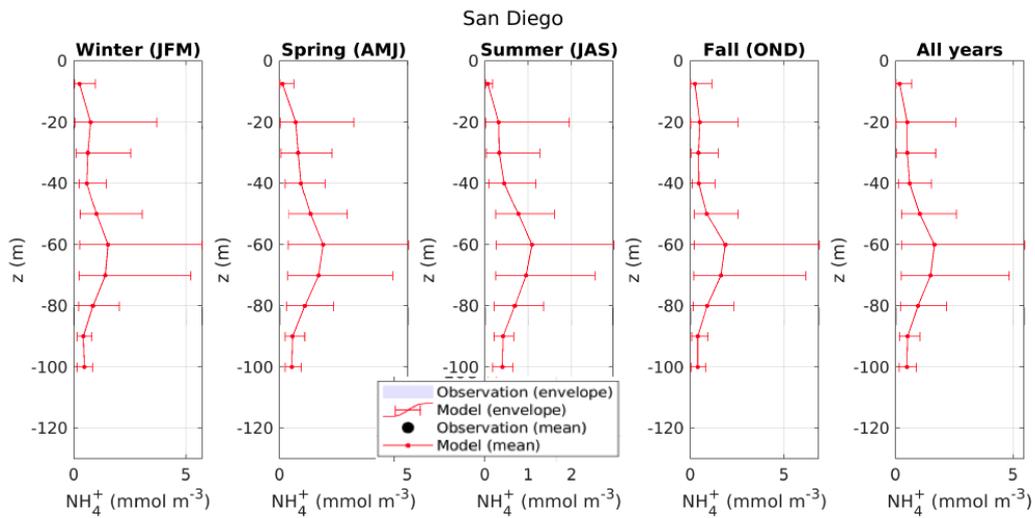


Figure S9: Same as Fig S6 for San Diego using City of San Diego stations. *In situ* data are missing but we wanted to report out the depth of maximum anthropogenic plume, in contrary to other subregion, in San Diego area, the plume rarely reaches 20 m.

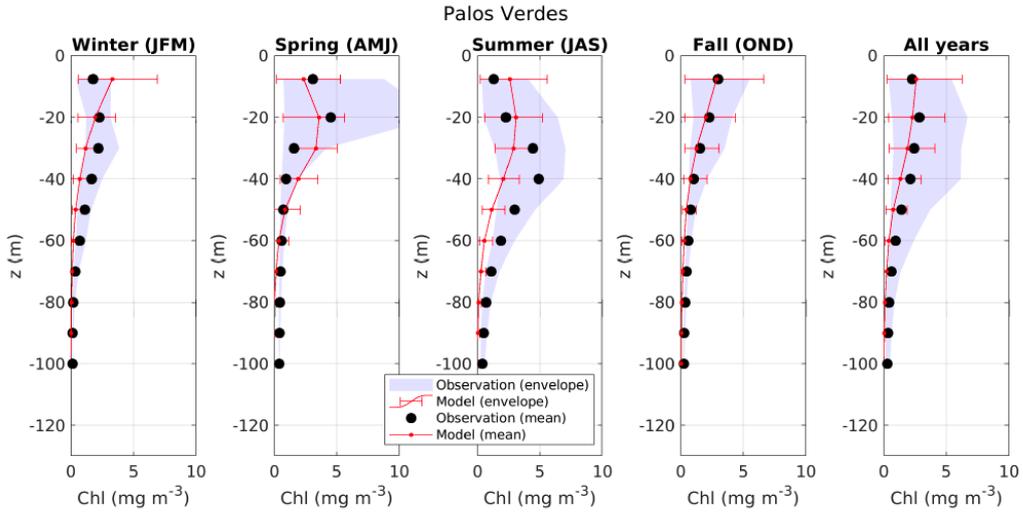


Figure S10: Seasonal profiles of average chlorophyll *a* concentration off of Palos Verdes. The red line and red bars are the spatial and temporal means and the variability from the model. The black dots and the gray shade are the spatial and temporal mean and the variability from *in situ* data (LACSD stations). These profiles are showing agreement on intensity, seasonality and shape of the vertical profile with exceptionally high concentrations at mid-depth.

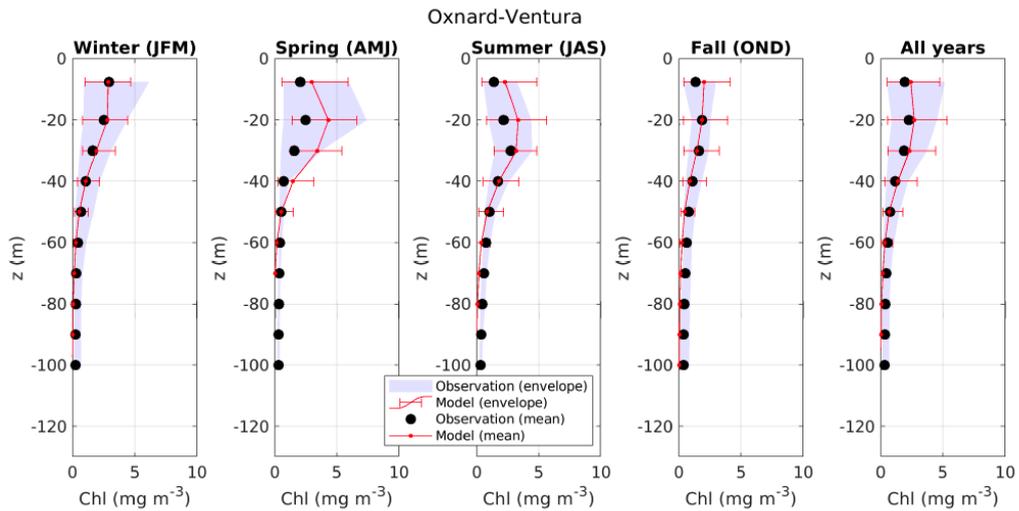


Figure S11: Same as Fig S10 for Oxnard/Ventura using City of Oxnard stations.

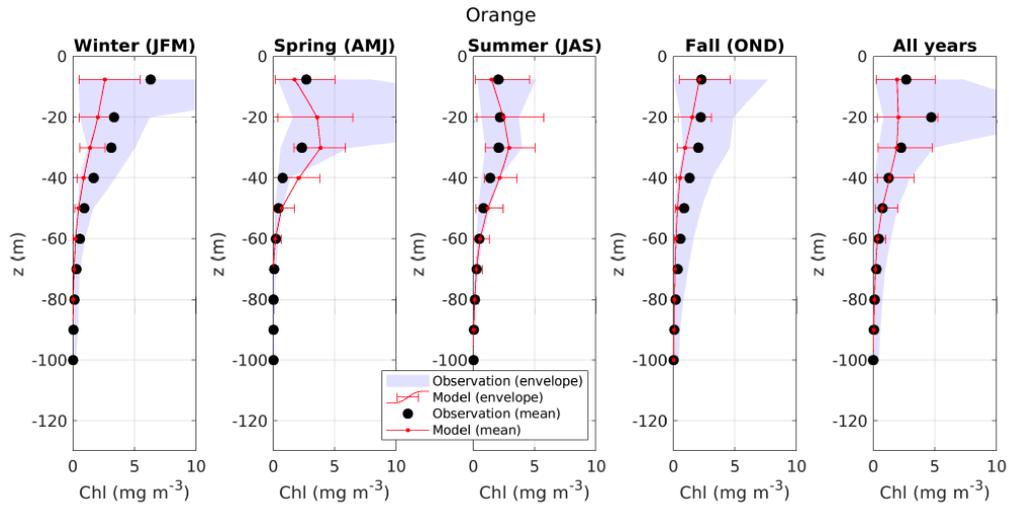


Figure S12: Same as Fig S10 for Orange County using OCSD stations.

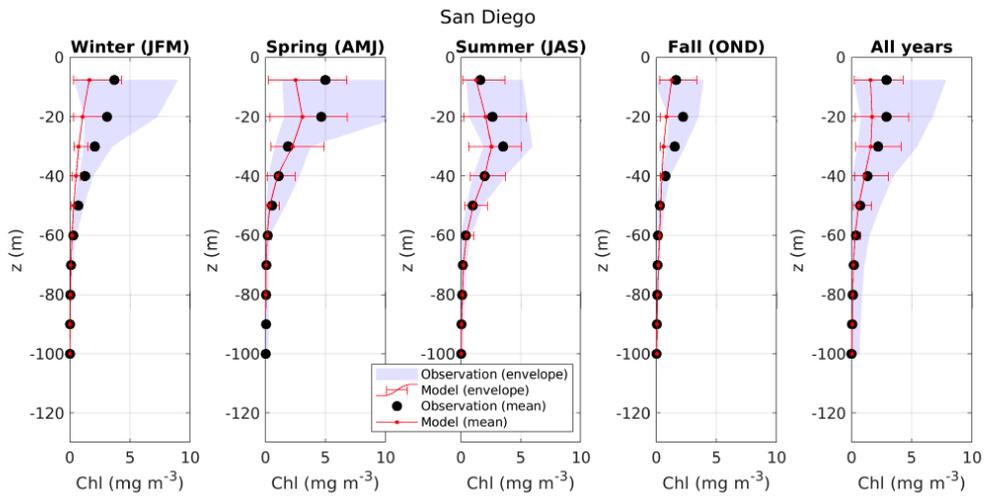


Figure S13: Same as Fig S10 for San Diego using City of San Diego stations.

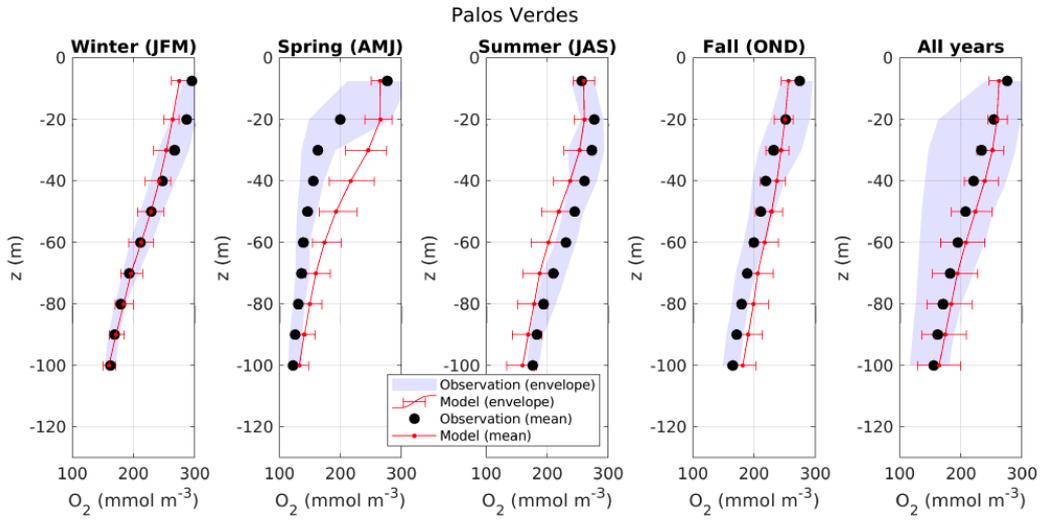


Figure S14: Seasonal profiles of average dissolved oxygen concentration off of Palos Verdes. The red line and red bars are the spatial and temporal means and the variability from the model. The black dots and the gray shade are the spatial and temporal mean and the variability from *in situ* data (LACSD stations). These profiles are showing agreement on intensity, seasonality and shape of the vertical profile with exceptionally high concentrations at mid-depth.

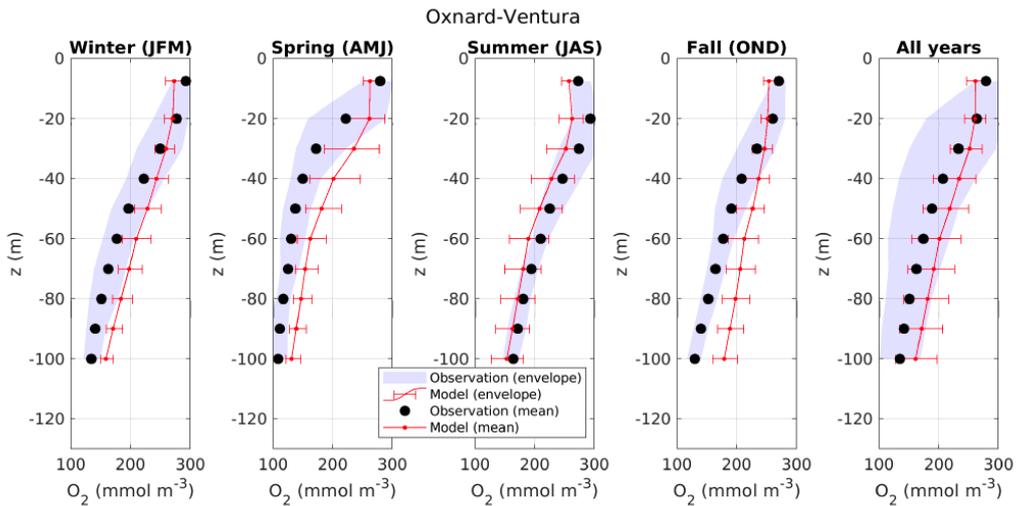


Figure S15: Same as Fig S14 for Oxnard/Ventura using City of Oxnard stations.

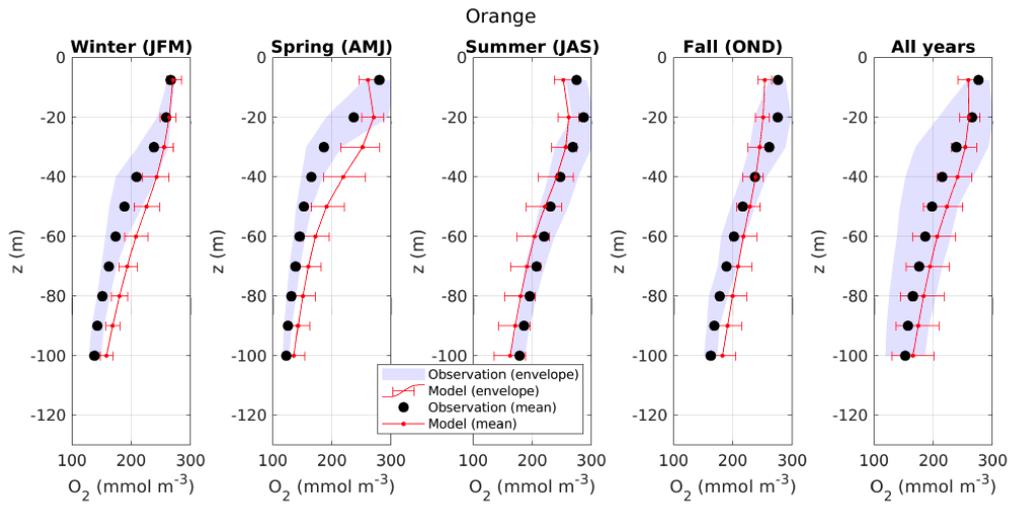


Figure S16: Same as Fig S14 for Orange County using OCSD stations.

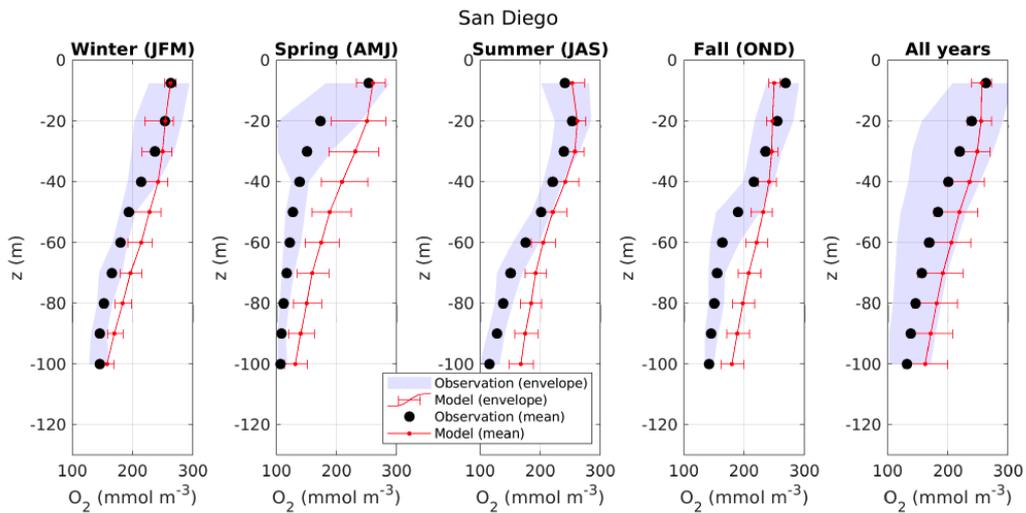


Figure S17: Same as Fig S14 for San Diego using City of San Diego stations.

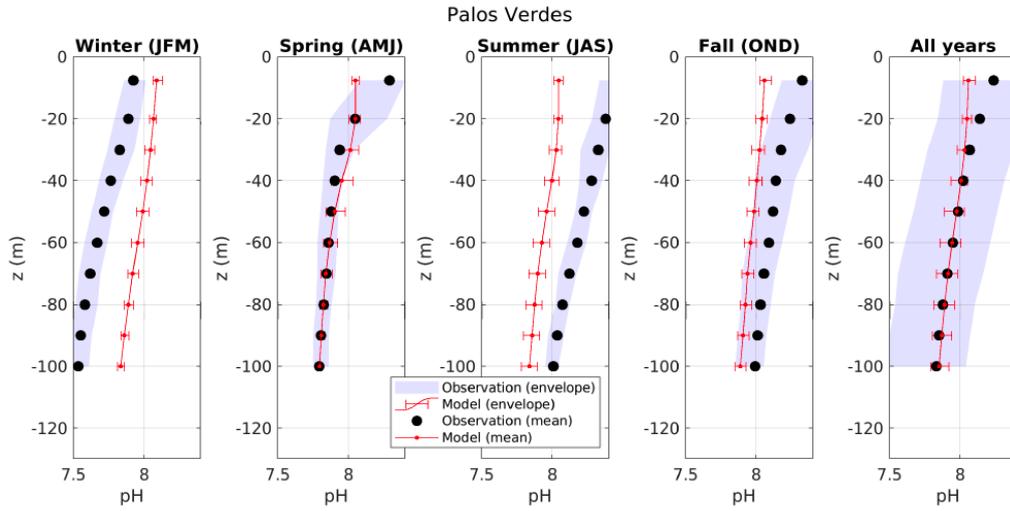


Figure S18: Seasonal profiles of average pH (seawater scale) off of Palos Verdes. The red line and red bars are the spatial and temporal means and the variability from the model. The black dots and the gray shade are the spatial and temporal mean and the variability from *in situ* data (LACSD stations). These profiles are showing agreement on intensity, seasonality and shape of the vertical profile with exceptionally high concentrations at mid-depth.

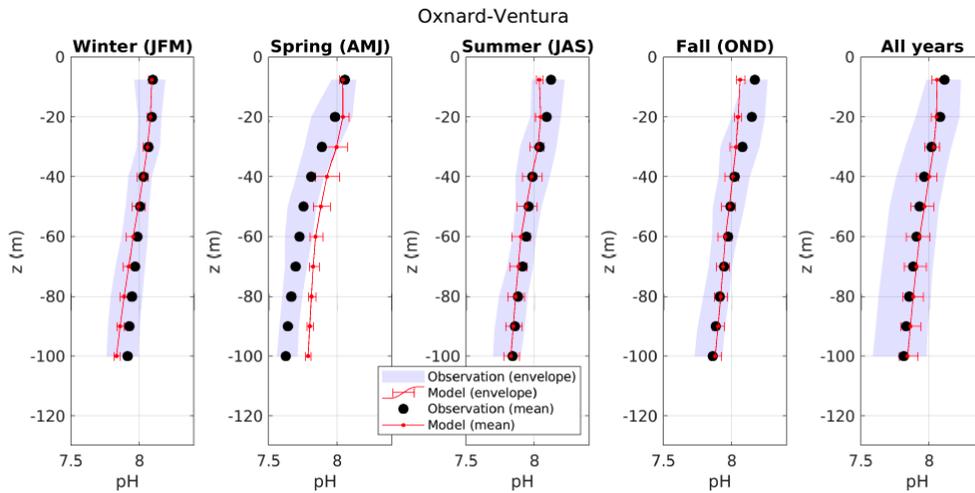


Figure S19: Same as Fig S18 for Oxnard/Ventura using City of Oxnard stations.

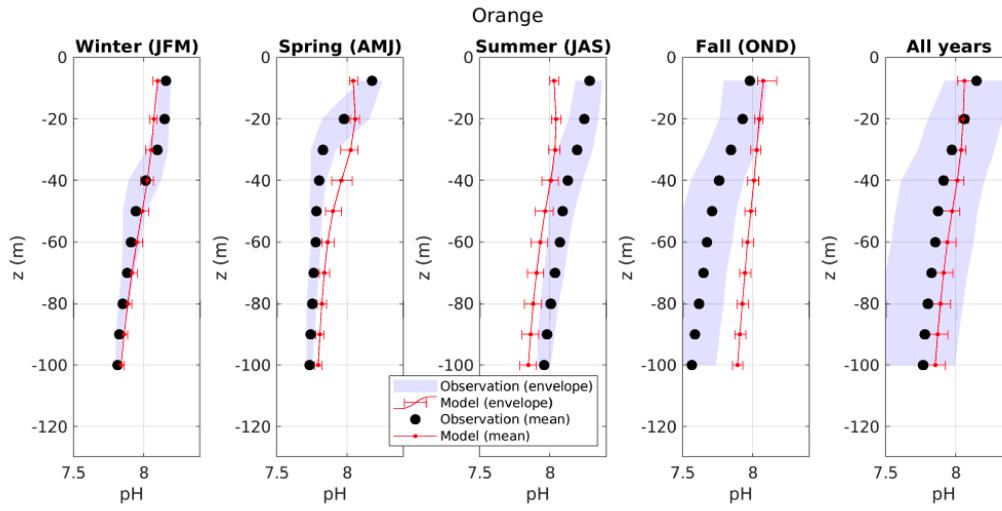


Figure S20: Same as Fig S18 for Orange County using OCSD stations.

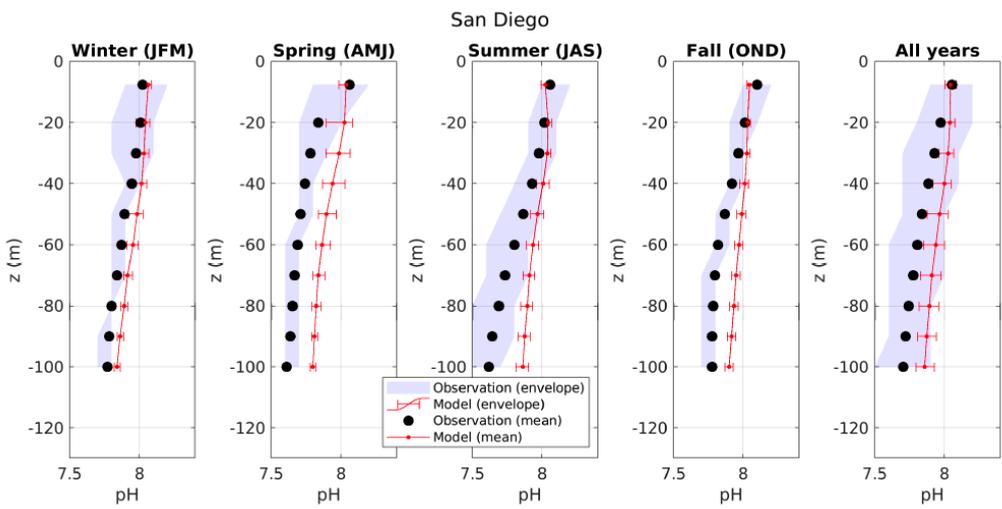


Figure S21: Same as Fig S18 for San Diego using City of San Diego stations.

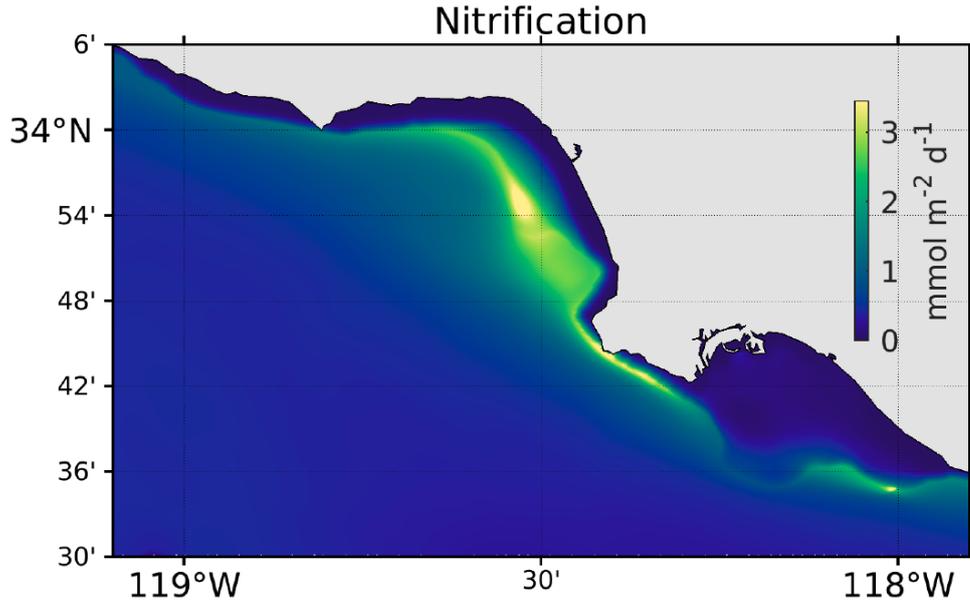


Figure S22: Average nitrification rate in Santa Monica and San Pedro bays. This figure shows the high rates around the locations of the outfalls that results from the release of high concentrations of ammonium below the thermocline.

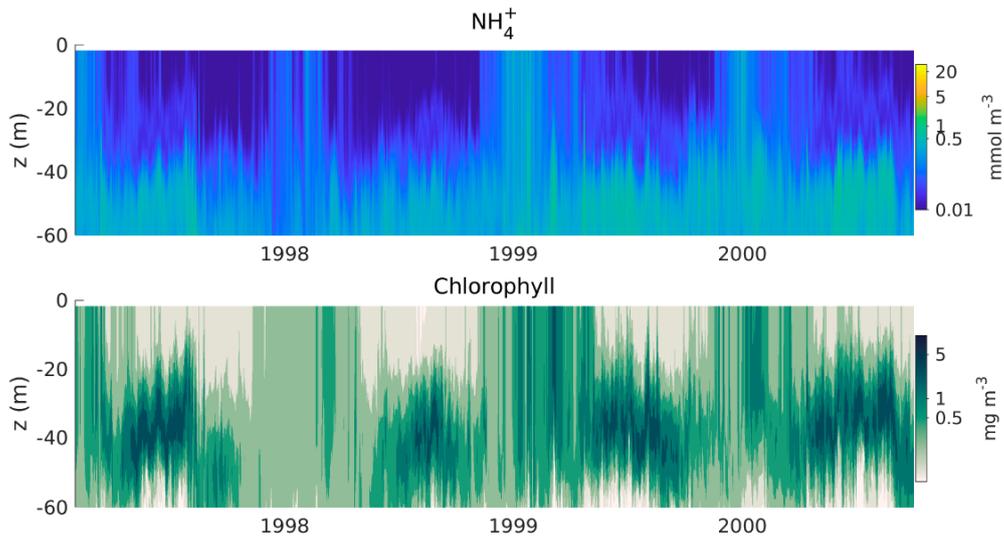


Figure S23: (upper panel) Hovmöller of ammonium concentration at San Pedro Oceanic Time-series (SPOT) located mid-distance between Los Angeles coast and Catalina Island. (lower panel) idem as (b) for chlorophyll *a* concentration. The Hovmöllers show 1) ammonium concentration off Los Angeles coast are not affected by anthropogenic loads. 2) Deep chlorophyll *a* maximum is trapped below at subsurface for 70% of the time and reach concentration of about 2 mmol Chl m^{-3} . Depth of the subsurface chlorophyll *a* maximum shows a seasonal cycle where it varies between 20 and 40m.

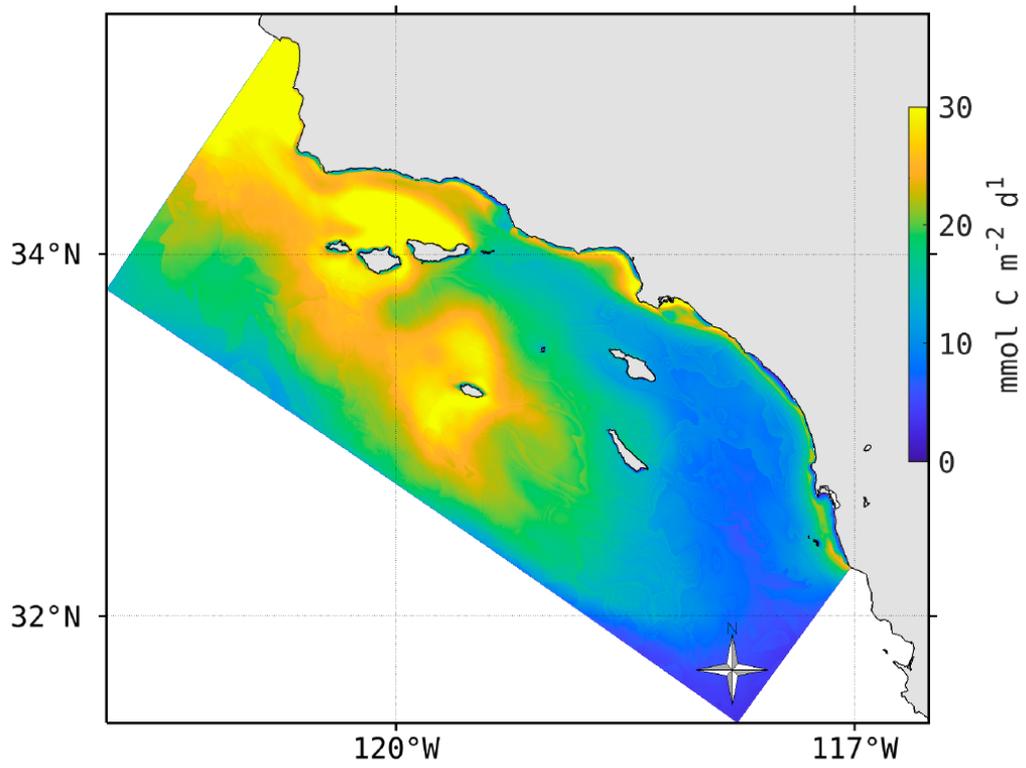


Figure S24: Summer time 1997-2000 average carbon export at 40 m in the SCB. The map shows hot-spots of intense carbon export in Santa Barbara and Los Angeles coasts.