

# The Spring Festival Effect on air quality in the megacities of China: View from space

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## Key Points:

- The satellite NO<sub>2</sub> columns observation catch the Spring Festival Effect, which is lowest during the vacation in the megacities in China.
- The tropospheric NO<sub>2</sub> columns density during the vacation around BTH region decreases about 40% than the period before the festival.
- The satellite-based NO<sub>2</sub> columns density during the Spring Festival decrease by 31.8% ~ 44.5% in Beijing, Tianjin, Shanghai and Chongqing.

## Abstract

The Spring Festival is the most important holiday in China, human activity and population mobility may contribute greatly to air quality, especially in the megacities. According to the satellite-based tropospheric nitrogen dioxide (NO<sub>2</sub>) column and ground-based observational concentration of NO<sub>2</sub> in the megacities from 2013 to 2018 around the Spring Festival, we found that NO<sub>2</sub> concentration decreases obviously during the Spring Festival in China, particularly in the megacities. The tropospheric NO<sub>2</sub> columns density around Beijing-Tianjin-Hebei region decreases about 40% than the period before the festival, that in Beijing and Tianjin decreases by 41.6% and 44.5% respectively. Similarly, the other two municipal cities Shanghai and Chongqing decrease by 43.2% and 31.8% in the urban. Consistently, the ground-based NO<sub>2</sub> concentration in the four megacities decrease by 18.9% ~ 38.8%.

Besides, the ground-based NO<sub>2</sub> concentration also decreases by above 20% during the Spring Festival vacation in the other eight megacities in China.

## 1 Introduction

The rapid development in China has contributed to heavy air pollution (Liu et al., 2018). Air pollution has huge impact on human health and the global ecological environment (Ghorani-Azam et al., 2016) even the climate (D'Amato et al., 2014). With the close relationship between air quality and human activity, many researches focus on this point. For example, Demircigil et al. (2014) have done the research of carrying out in buccal epithelial cells from children in an urban city of Turkey to analyze the genotoxic effect of air pollution, and found that seasonal variation in genotoxicity may be related to the time spent at outdoors in summer.

NO<sub>2</sub> is one of the most important air pollutant species, owing to its relatively short lifetime, it would not be transported far from its sources (Richter et al., 2005). Thus, satellite and remote sensing images provide a direct indication of where NO<sub>2</sub> sources are located. Boersma et al. (2011) explored the sensitive factors of tropospheric NO<sub>2</sub> column retrieval algorithm for the Ozone Monitoring Instrument (OMI) to improve the data quality. Bechle et al. (2013) found that the correlation coefficient between annual OMI NO<sub>2</sub> column density and ground data in southern California is up to 0.93. Tropospheric NO<sub>2</sub> columns density of satellite observations products with high quality have also been successfully used in many researches to infer NO<sub>x</sub> emissions and trends (Boersma et al., 2007; Silvern et al., 2019). According to Goldberg et al. (2017), OMI NO<sub>2</sub> has been used to estimate NO<sub>x</sub> emissions in various region around the globe such as the US, southern California, Europe and East Asia (Chang et al., 2017; Duncan et al., 2016). Some researchers have found the trend of NO<sub>x</sub> emission trend in China over this decade, it increased in 2005-2010 (Verstraeten et al., 2015) and tended to be stable in 2011-2012 (Souri et al., 2017) while since 2012 it has decreased. The research by Foy et al. (2016) is on the trend of NO<sub>2</sub> emission from 2005 to 2015 by OMI and the satellite results were in great agreement with the annual trends of the NO<sub>x</sub> emissions inventory until 2014, which showed that the reliability of evaluating trends of NO<sub>x</sub> emission by OMI.

There also has been an emphasis on the effect of the special period on air quality. The Spring Festival is the most important festival in China that most Chinese would move from cities to their hometown for family reunion before the Spring Festival and back from their hometown to cities for work after the Spring Festival, and most of the industries, companies and production sites in the cities are closed during the vacation around the Spring Festival (Tang et al., 2016; Wang et al., 2017; Yao et al., 2019). Wang et al. (2017) chose urban Shenzhen over three consecutive winters of 2014-2016 to research the air pollution based on the effect of the Spring Festival, and the results indicate that the air pollutants decrease heavily during the

Spring Festival relative to non-Spring Festival periods, regardless of the meteorological conditions. Gong et al. (2014) chose the long-term observation from 2001 to 2012 in eastern China and found the feature of the air pollutants around the Spring Festival is the significant reduction of concentration. This study collected the satellite-based tropospheric NO<sub>2</sub> column and the ground-based NO<sub>2</sub> concentration to quantify the Spring Festival effect on air quality in the megacities in China.

## 2 Observation and Methods

Considering the particularity of the Spring Festival, for human activity and population mobility may contribute a lot to air quality in the megacities especially, the typical episode of up to 3 x 18 days is chosen to compare the concentration of NO<sub>2</sub> in different periods of the special time. The satellite-based tropospheric column observed by OMI and ground-based observational concentration of NO<sub>2</sub> in the megacities from the China National Environmental Monitoring Centre (CNEMC) before, during and after the Spring Festival from 2013 to 2018 have been collected for this study.

The data of tropospheric NO<sub>2</sub> columns density are accessed from OMI of EU FP7 project Quality Assurance for Essential Climate Variable (QA4ECV). The main product is the tropospheric NO<sub>2</sub> column density, which is defined as the vertically integrated number of NO<sub>2</sub> molecules between the Earth's ground and the tropopause. It has an overpass time about 13:30~13:50 at local time (Goldberg et al., 2017; Lamsal et al., 2008). The regional tropospheric NO<sub>2</sub> columns of individual days on the globe from 2013 to 2018 are accessed from OMI and then extract and obtain the region data of China.

The daily observation of NO<sub>2</sub> is obtained from the CNEMC, covering 12 megacities in China, including Shanghai(SH), Beijing(BJ), Shenzhen(SZ), Guangzhou(GZ), Chengdu(CD), Hangzhou(HZ), Chongqing(CQ), Wuhan(WH), Xi'an(XA), Tianjin(TJ), Nanjing(NJ) and Zhengzhou(ZZ) from 2013 to 2018(Figure 1). Based on the ranking of China cities business attractiveness 2018 by China Business Network (CBN), we chose those 12 megacities with great financial conditions, public transport facilities, social environment, and large population (China Business Network, 2018).

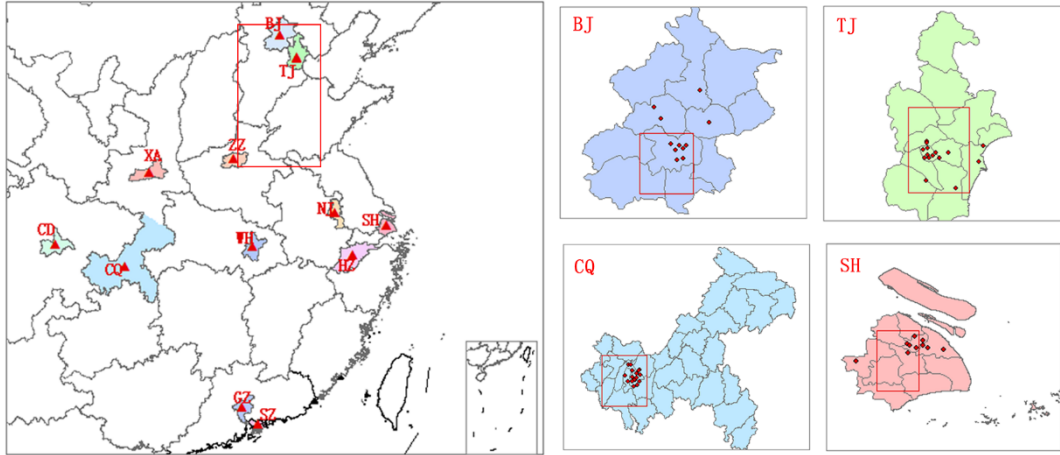


Figure 1. The location of researched cities and observational sites. The red triangle in the left subfigure were the megacities mentioned in this study, while the red box cover Beijing-Tianjin-Hebei (BTH) region. The observational sites in BJ, TJ, CQ and SH was shown as the red points in the right subfigures, and the compared urban area of OMI in each municipality is enclosed by the red box.

The Spring Festival vacation in China traditionally is from Spring Festival, the 1<sup>st</sup> day of the Chinese lunar year, to Lantern Festival, the 15<sup>th</sup> day of the Chinese lunar year. Due to the super vast area of China, the Chinese New Year has been regarded as the largest migration in the world (Huang et al., 2012). Moreover, as the importance of Chinese New Year and long-time for transportation and preparing for it, people need to get home back in advance, so that we defined the complete vacation as during the Spring Festival (DF) from three days before the 1<sup>st</sup> day to the 15<sup>th</sup> day of Chinese lunar year. Then, we compared the different periods to explore the relationship between human activity and air quality, that before the Spring Festival (BF) and after the Spring Festival (AF) has the same 18 days. The specific dates of BF, DF and AF periods are shown in Figure 2.

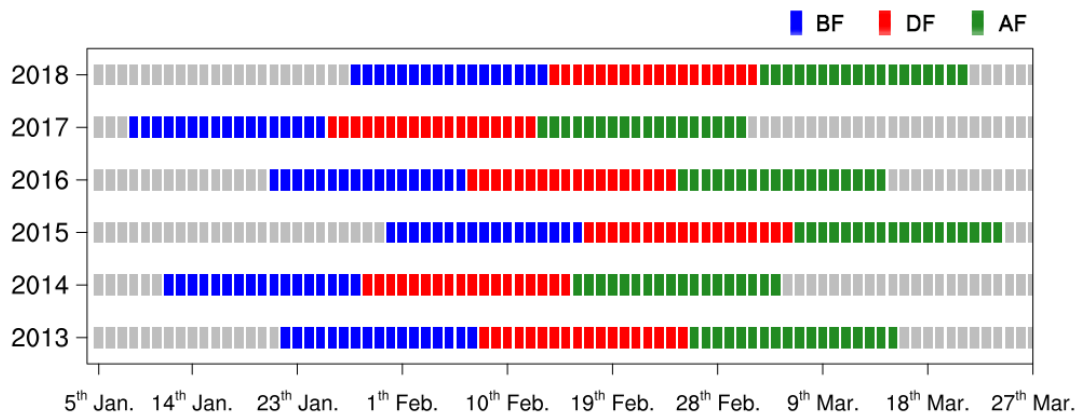


Figure 2. The dates of each period around the Spring Festival from the year 2013 to 2018. The panes with blue, red and green represent the periods before the festival (BF), during the

festival (DF) and after the festival (AF) respectively. The AF period in 2016 including 29<sup>th</sup> Feb.

Considering the pixel size of OMI is  $13 \text{ km} \times 24 \text{ km}$  at nadir or the center of the swath, we collected the satellite data with grid resolution of  $0.25^\circ \times 0.25^\circ$ . As OMI  $\text{NO}_2$  algorithm is highly sensitive to clouds, the OMI  $\text{NO}_2$  has been filtered by removing the data with cloud radiance fractions  $\geq 0.5$  (Bucsela et al., 2008; Hains et al., 2010; Silvern et al., 2019) and solar zenith angles  $\geq 80^\circ$ , and the benefit of this approach is that only valid pixel is used (Goldberg et al., 2017). Accordingly, we applied the ground-based  $\text{NO}_2$  concentration observed by CNEMC in these periods from year 2013 to 2018 to decrease the influence of meteorology, and analyzed the daily concentration change and site distribution of each megacities to explore the special phenomenon of air quality on  $\text{NO}_2$ .

### 3 Spatial and temporal distribution around the Spring Festival

This section describes the change of  $\text{NO}_2$  concentration around the Spring Festival according to the satellite-based tropospheric  $\text{NO}_2$  columns and ground-based observational  $\text{NO}_2$  concentration. The spatial distribution of tropospheric  $\text{NO}_2$  columns is present in Figure 3. It shows the tropospheric  $\text{NO}_2$  columns density in DF is less than BF and AF around Beijing-Tianjin-Hebei (BTH) region and megacities including SH, BJ, SZ, GZ, CD, HZ, CQ, WH, XA, TJ, NJ and ZZ obviously.

The mean tropospheric  $\text{NO}_2$  columns density of BTH region is  $32.53 \text{ molec/cm}^2$  before the Spring Festival, and decreases to  $19.59 \text{ molec/cm}^2$  during the Spring Festival, by about 40%.

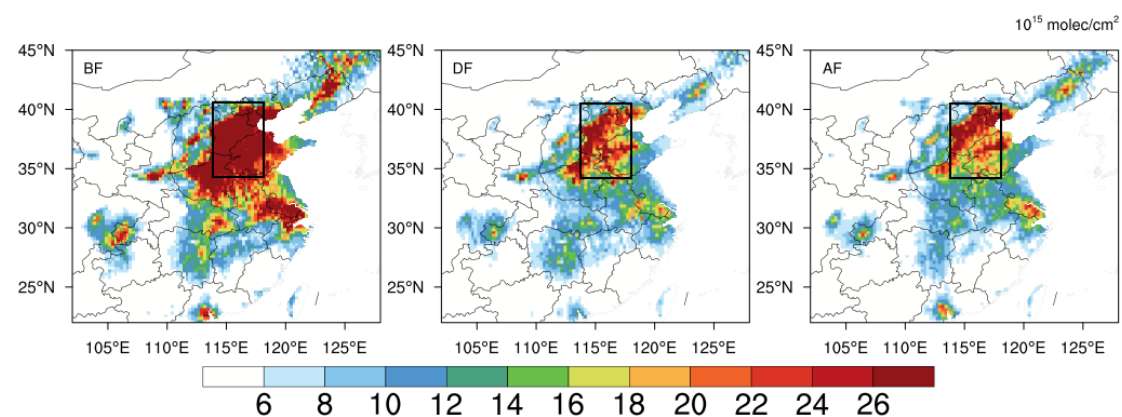


Figure 3. The change of tropospheric  $\text{NO}_2$  columns in most region of China (units:  $10^{15} \text{ molec/cm}^2$ ). The data less than  $6 \times 10^{15} \text{ molec/cm}^2$  has been set to white. The area with black box is around Beijing-Tianjin-Hebei region.

The highest data is over  $50 \text{ molec/cm}^2$  in Beijing before the Spring Festival. The average tropospheric  $\text{NO}_2$  columns density of Beijing is  $41.86 \text{ molec/cm}^2$  in BF and reduces to  $24.46 \text{ molec/cm}^2$  in DF while it rebounds to  $29.92 \text{ molec/cm}^2$  after the

Spring Festival ,which is shown in Figure 4a. As well as the trend in Tianjin (Figure 4b), Shanghai (Figure 4c) and Chongqing (Figure 4d) is the same with Beijing. Similarly to the most region of China, these municipalities all show the tropospheric NO<sub>2</sub> columns density is featured by a notable reduction during the Spring Festival and a relatively weak rebound after the Spring Festival. The average tropospheric NO<sub>2</sub> columns density of Tianjin is 40.98 molec/cm<sup>2</sup> in BF, and it reduces to 22.75 molec/cm<sup>2</sup> in DF while rebounds to 28.58 molec/cm<sup>2</sup> after the Spring Festival. The tropospheric NO<sub>2</sub> columns density of Shanghai decreases from 27.64 molec/cm<sup>2</sup> to 15.69 molec/cm<sup>2</sup> and backs to 18.39 molec/cm<sup>2</sup> over time. As for Chongqing, the data changes from 21.43 molec/cm<sup>2</sup> to 14.62 molec/cm<sup>2</sup> and rebounds to 15.51 molec/cm<sup>2</sup>. Based on the above observation analysis we may conclude that the tropospheric NO<sub>2</sub> columns density over the most well-developed area of China is featured by a notable reduction during the Spring Festival and a relatively weak rebound after the Spring Festival.

The most surprising aspect of the result is that the former decrease dramatically reaches over 40% in Beijing, Tianjin and Shanghai, and the NO<sub>x</sub> emission in these megacities decrease about 40% during the Spring Festival, which mainly emit from vehicle, industry and so on. While the latter rebound in Beijing, Tianjin, and Shanghai only reach about 20%. The tropospheric NO<sub>2</sub> columns density in Tianjin fluctuates more heavily than other cities with 44.5% and 25.6% during the twice period-change respectively. In contrast, Chongqing is the least one regarding the change of tropospheric NO<sub>2</sub> columns density as decreasing 31.8% during the Spring Festival while rebounding 6.2% after the Spring Festival.

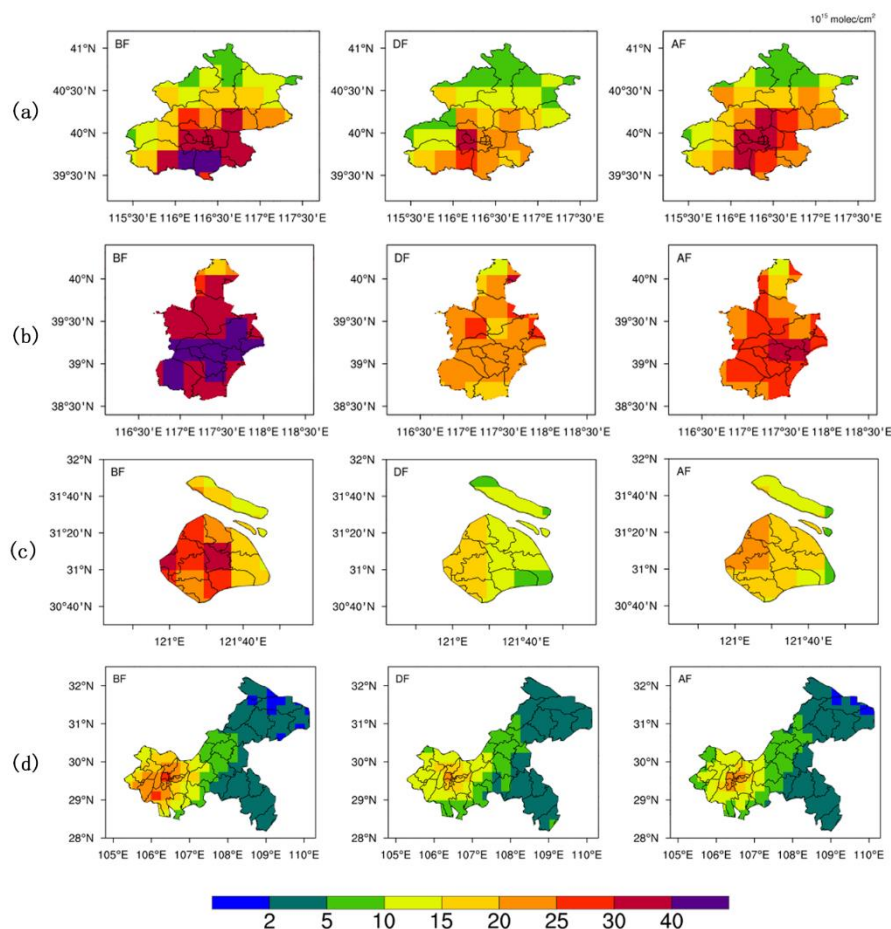
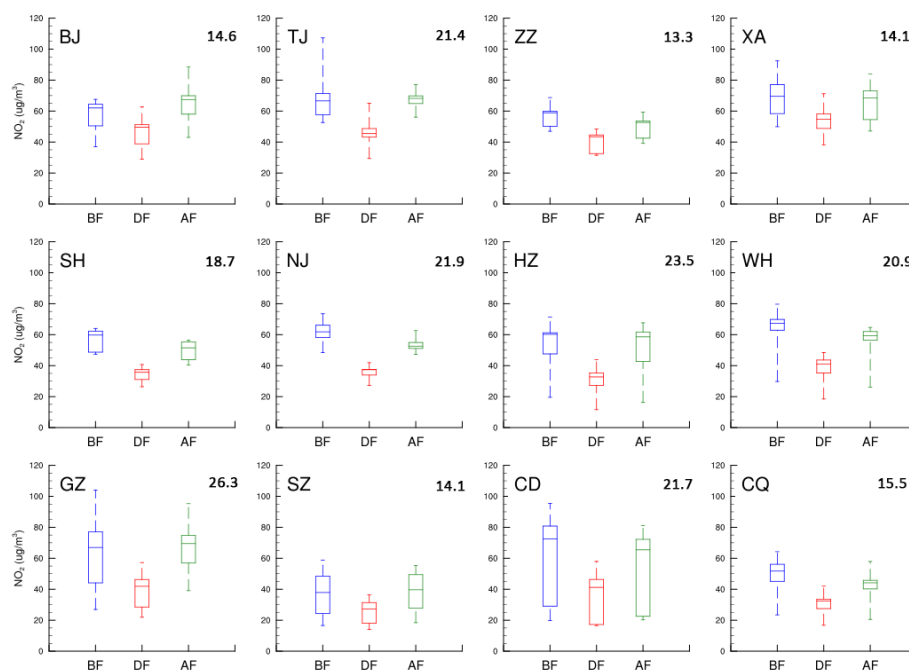


Figure 4. The change of tropospheric  $\text{NO}_2$  columns in Beijing(a), Tianjin(b), Shanghai(c) and Chongqing(d). Units: $10^{15} \text{ molec/cm}^2$ .

The trend of ground-based  $\text{NO}_2$  concentration from CNEMC is coincident well. Figure 5 provides the daily  $\text{NO}_2$  concentration on 18-day averaged of each period in the megacities. As shown in Figure 5, the ground-based  $\text{NO}_2$  concentration during the Spring Festival is much lower than periods before and after the Spring Festival. The mean absolute difference between DF period and the other two periods in Guangzhou, Hangzhou, Nanjing, Chengdu, and Tianjin is up to  $20 \mu\text{g/m}^3$ . In contrast, Zhengzhou is  $13.3 \mu\text{g/m}^3$ , while Xi'an, Shenzhen, and Beijing are lower than  $15 \mu\text{g/m}^3$ .



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187 Figure 5. The box plot of NO<sub>2</sub> concentration observed by CNEMC in the megacities. The box  
 188 in blue, red and green represents BF, DF and AF. The dash in each color box means the  
 189 median NO<sub>2</sub> concentration in the city, and the stick at each end means the maximum and  
 190 minimum. The number at the top right means the average absolute difference between DF  
 191 period and the other two periods.

192 Comparing the results in deep with the satellite-based and ground-based NO<sub>2</sub>  
 193 concentration in Beijing, Tianjin, Shanghai, and Chongqing, they both show  
 194 significantly reduction during the Spring Festival. As shown in Figure 6, the NO<sub>2</sub>  
 195 concentration during the Spring Festival is the lowest one whether the tropospheric  
 196 NO<sub>2</sub> columns density or ground concentration. There are some differences in the two  
 197 types of data, such as the change proportion and concentration temporal distribution  
 198 in the same city. For example, Beijing has the highest concentration in BF by  
 199 tropospheric NO<sub>2</sub> columns density but in AF by ground concentration. The ground  
 200 concentration after festival has recovered evidently while it is not too obviously by  
 201 tropospheric NO<sub>2</sub> columns density in Chongqing. Moreover, the tropospheric NO<sub>2</sub>  
 202 columns density in DF of Beijing has reduced by 41.6% while it only reduced by 18.9%  
 203 on ground-based NO<sub>2</sub> concentration. Tianjin is the same with Beijing for the former is  
 204 44.5% and the latter only up to 29.81%. The change trend of two types data in other  
 205 two municipal cities Shanghai and Chongqing have little difference. It decreases by  
 206 43.2% and 31.8% in the urban of Shanghai and Chongqing according to the  
 207 tropospheric NO<sub>2</sub> columns while it is 38.8%, 38.1%, respectively. This result  
 208 indicates the overall trend is that NO<sub>2</sub> concentration during the Spring Festival  
 209 decrease and is the lowest period, but the specific trend and proportion in different  
 210 data show different results.



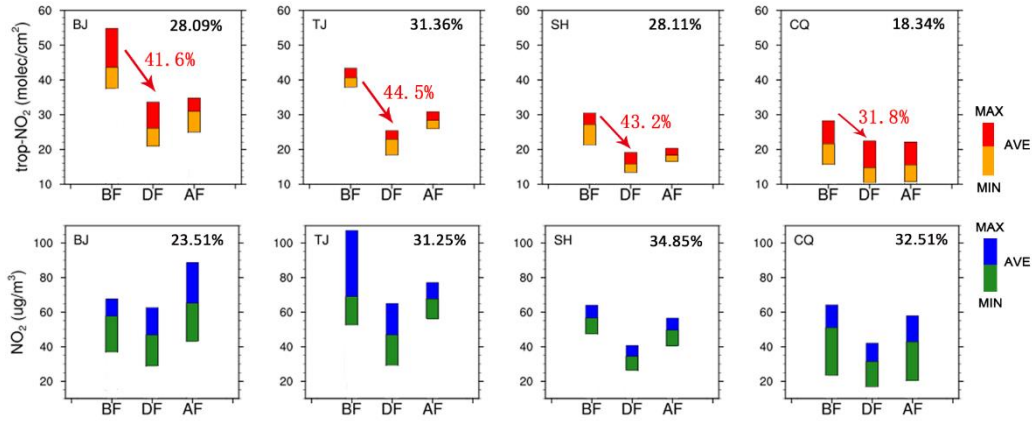


Figure 6. The concentration change of NO<sub>2</sub> in the megacities on average of sites. The number at the top right of each city picture means the average of change proportion between DF period and the other two periods. The red number near the red arrow in tropospheric NO<sub>2</sub> columns subfigure means the reduction proportion from BF period to DF period on average.

The ranking of 12 researched cities in 2018 by China Business Network is Shanghai, Beijing, Shenzhen, Guangzhou, Chengdu, Hangzhou, Chongqing, Wuhan, Xi'an, Tianjin, Nanjing, Zhengzhou in order. The average ground-based observational NO<sub>2</sub> concentration of the 12 megacities during the Spring Festival changes from 59.48 µg/m<sup>3</sup> that before the Spring Festival to 38.94 µg/m<sup>3</sup>, a reduction of 20.54 µg/m<sup>3</sup>. And it rebounds 16.88 µg/m<sup>3</sup>, up to 55.82 µg/m<sup>3</sup> after the Spring Festival. The highest concentration of NO<sub>2</sub> during the Spring Festival is 54.83 µg/m<sup>3</sup>, which is appeared in Xi'an. The top three of the highest DF concentration are in the order of Xi'an, Tianjin and Beijing. And Zhengzhou, Chengdu, Guangzhou, Wuhan, Nanjing, Shanghai, Chongqing and Hangzhou are over 30 µg/m<sup>3</sup> while Shenzhen is the least one with only 25.25 µg/m<sup>3</sup> in ground-based NO<sub>2</sub> concentration. Actually, the average concentration on the whole Spring Festival researched period of Shenzhen is also the least one with only 34.56 µg/m<sup>3</sup>, while Xi'an is 64.22 µg/m<sup>3</sup>, which is also the highest one. The results further verify that air quality is closely related to human activity including the financial conditions, public transport facilities, industry, population and so on.

#### 4 Conclusions

This study researched the satellite-based tropospheric NO<sub>2</sub> columns density by Ozone Monitoring Instrument and ground-based NO<sub>2</sub> concentration published by the China National Environmental Monitoring Centre around the Chinese Spring Festival and found that the concentration of NO<sub>2</sub> during the festival in megacities decreased a lot, which is highly related to human activity and called the "Spring Festival Effect". That means we can use the satellite-based tropospheric NO<sub>2</sub> columns density or ground-based NO<sub>2</sub> concentration as one index to describe the characterization and intensity of human activity.

Air quality during the Spring Festival in some well-developed areas and megacities such as Beijing-Tianjin-Hebei region and most provincial cities is better than other periods before and after the Spring Festival. Around Beijing-Tianjin-Hebei region, the tropospheric NO<sub>2</sub> columns density during the Spring Festival decreases about 40% than the period before the festival. Moreover, the tropospheric NO<sub>2</sub> columns density in Beijing, Tianjin, Shanghai, and Chongqing during the Chinese Spring Festival decrease by 31.8% ~ 44.5% compared with that in period of before the festival.

Meanwhile, the ground-based NO<sub>2</sub> concentration is also less during the Spring Festival than the periods before and after the vacation in China, particularly in the megacities similarly. Compared with the satellite-based tropospheric NO<sub>2</sub> columns density in the researched megacities, the NO<sub>2</sub> concentration during the Chinese Spring Festival decrease by 18.9% ~ 38.8% than period before the festival. Besides, the ground-based NO<sub>2</sub> concentration also decreases by above 20% during the Spring Festival vacation in the other eight megacities in China.

#### **Data availability.**

The data of the satellite-based tropospheric NO<sub>2</sub> columns by OMI is available online via QA4ECV ([http://www.temis.nl/airpollution/no2col/no2regioomi\\_qa.php](http://www.temis.nl/airpollution/no2col/no2regioomi_qa.php)) and the ground-based NO<sub>2</sub> concentration by CNEMC (<http://www.cnemc.cn/>) is available online via [http://www.pm25.in/api\\_doc](http://www.pm25.in/api_doc). The other data are available online via ZENODO (<https://zenodo.org/deposit/3630222>; Li et al., 2020).

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#### **References**

- Bechle, M. J., Millet, D. B., & Marshall, J. D. (2013). Remote sensing of exposure to NO<sub>2</sub>: Satellite versus ground-based measurement in a large urban area. *Atmospheric Environment*, 69, 345–353. <https://doi.org/10.1016/j.atmosenv.2012.11.046>
- Boersma, K. F. [K. F.], Eskes, H. J., Dirksen, R. J., van der A, R. J., Veefkind, J. P. [J. P.], Stammes, P., et al. (2011). An improved tropospheric NO<sub>2</sub> column retrieval algorithm for the Ozone Monitoring Instrument. *Atmospheric Measurement Techniques*, 4(9), 1905–1928. <https://doi.org/10.5194/amt-4-1905-2011>

275 Boersma, K. F. [K. F.], Eskes, H. J., Veefkind, J. P. [J. P.], Brinksma, E. J., van der  
 276 A, R. J., Sneep, M., et al. (2007). Near-real time retrieval of tropospheric NO<sub>2</sub> from  
 277 OMI. *Atmospheric Chemistry and Physics*, 7(8), 2103–2118.  
 278 <https://doi.org/10.5194/acp-7-2103-2007>

279 Bucsela, E. J., Perring, A. E. [A. E.], Cohen, R. C. [R. C.], Boersma, K. F. [K. F.],  
 280 Celarier, E. A., Gleason, J. F. [J. F.], et al. (2008). Comparison of tropospheric NO<sub>2</sub>  
 281 from in situ aircraft measurements with near-real-time and standard product data  
 282 from OMI. *Journal of Geophysical Research*, 113(D16), D05204.  
 283 <https://doi.org/10.1029/2007JD008838>

284 Chang, K.-L., Petropavlovskikh, I., Copper, O. R., Schultz, M. G., & Wang, T. (2017).  
 285 Regional trend analysis of surface ozone observations from monitoring networks in  
 286 eastern North America, Europe and East Asia. *Elem Sci Anth*, 5(0), 50.  
 287 <https://doi.org/10.1525/elementa.243>

288 China Business Network (2018). Ranking of China Cities' Business Attractiveness  
 289 2018. Retrieved from <https://max.book118.com/html/2018/0429/163757892.shtm>

290 D'Amato, G., Bergmann, K. C., Cecchi, L., Annesi-Maesano, I., Sanduzzi, A.,  
 291 Liccardi, G., et al. (2014). Climate change and air pollution: Effects on pollen  
 292 allergy and other allergic respiratory diseases. *Allergo Journal International*, 23(1),  
 293 17–23. <https://doi.org/10.1007/s40629-014-0003-7>

294 Demircigil, G. Ç., Erdem, O., Gaga, E. O., Altuğ, H., Demirel, G., Özden, Ö., et al.  
 295 (2014). Cytogenetic biomonitoring of primary school children exposed to air  
 296 pollutants: Micronuclei analysis of buccal epithelial cells. *Environmental Science*  
 297 *and Pollution Research International*, 21(2), 1197–1207.  
 298 <https://doi.org/10.1007/s11356-013-2001-6>

299 Duncan, B. N., Lamsal, L. N. [Lok N.], Thompson, A. M., Yoshida, Y., Lu, Z.,  
 300 Streets, D. G., et al. (2016). A space-based, high-resolution view of notable changes  
 301 in urban NO<sub>x</sub> pollution around the world (2005-2014). *Journal of Geophysical*  
 302 *Research: Atmospheres*, 121(2), 976–996. <https://doi.org/10.1002/2015JD024121>

303 Foy, B. de, Lu, Z., & Streets, D. G. (2016). Satellite NO<sub>2</sub> retrievals suggest China has  
 304 exceeded its NO<sub>x</sub> reduction goals from the twelfth Five-Year Plan. *Scientific*  
 305 *Reports*, 6, 35912. <https://doi.org/10.1038/srep35912>

306 Ghorani-Azam, A., Riahi-Zanjani, B., & Balali-Mood, M. (2016). Effects of air  
 307 pollution on human health and practical measures for prevention in Iran. *Journal of*  
 308 *Research in Medical Sciences: the Official Journal of Isfahan University of Medical*  
 309 *Sciences*, 21, 65. <https://doi.org/10.4103/1735-1995.189646>

310 Goldberg, D. L., Lamsal, L. N. [Lok N.], Loughner, C. P., Swartz, W. H., Lu, Z., &  
 311 Streets, D. G. (2017). A high-resolution and observationally constrained OMI NO<sub>2</sub>

312 satellite retrieval. *Atmospheric Chemistry and Physics*, 17(18), 11403–11421.  
313 <https://doi.org/10.5194/acp-17-11403-2017>

314 Gong, D.-Y., Wang, W., Qian, Y., Bai, W., Guo, Y., & Mao, R. (2014). Observed  
315 holiday aerosol reduction and temperature cooling over East Asia. *Journal of*  
316 *Geophysical Research: Atmospheres*, 119(11), 6306–6324.  
317 <https://doi.org/10.1002/2014JD021464>

318 Hains, J. C., Boersma, K. F. [K. Folkert], Kroon, M., Dirksen, R. J. [Ruud J.],  
319 Cohen, R. C. [Ronald C.], Perring, A. E. [Anne E.], et al. (2010). Testing and  
320 improving OMI DOMINO tropospheric NO<sub>2</sub> using observations from the  
321 DANDELIONS and INTEX-B validation campaigns. *Journal of Geophysical*  
322 *Research*, 115(D5), D05204. <https://doi.org/10.1029/2009JD012399>

323 Huang, K., Zhuang, G., Lin, Y., Wang, Q., Fu, J. S., Zhang, R., et al. (2012). Impact of  
324 anthropogenic emission on air quality over a megacity – revealed from an intensive  
325 atmospheric campaign during the Chinese Spring Festival. *Atmospheric Chemistry*  
326 *and Physics*, 12(23), 11631–11645. <https://doi.org/10.5194/acp-12-11631-2012>

327 Lamsal, L. N. [L. N.], Martin, R. V., van Donkelaar, A., Steinbacher, M.,  
328 Celarier, E. A., Bucsela, E., et al. (2008). Ground-level nitrogen dioxide  
329 concentrations inferred from the satellite-borne Ozone Monitoring Instrument.  
330 *Journal of Geophysical Research*, 113(D16), D05204.  
331 <https://doi.org/10.1029/2007JD009235>

332 Liu, W., Xu, Z., & Yang, T. (2018). Health Effects of Air Pollution in China.  
333 *International Journal of Environmental Research and Public Health*, 15(7).  
334 <https://doi.org/10.3390/ijerph15071471>

335 Richter, A. [Andreas], Burrows, J. P., Nüss, H., Granier, C., & Niemeier, U. (2005).  
336 Increase in tropospheric nitrogen dioxide over China observed from space. *Nature*,  
337 437(7055), 129–132. <https://doi.org/10.1038/nature04092>

338 Silvern, R. F., Jacob, D. J., Mickley, L. J., Sulprizio, M. P., Travis, K. R.,  
339 Marais, E. A., et al. (2019). Using satellite observations of tropospheric NO<sub>2</sub>  
340 columns to infer long-term trends in US NO<sub>x</sub> emissions: the importance of  
341 accounting for the free tropospheric NO<sub>2</sub> background. *Atmospheric Chemistry and*  
342 *Physics Discussions*, 1–26. <https://doi.org/10.5194/acp-2019-168>

343 Souri, A. H., Choi, Y., Jeon, W., Woo, J.-H., Zhang, Q., & Kurokawa, J.-i. (2017).  
344 Remote sensing evidence of decadal changes in major tropospheric ozone  
345 precursors over East Asia. *Journal of Geophysical Research: Atmospheres*, 122(4),  
346 2474–2492. <https://doi.org/10.1002/2016JD025663>

347 Tang, M., Ji, D.-S., Gao, W.-K., Yu, Z.-W., Chen, K., & Cao, W. (2016).  
348 Characteristics of air quality in Tianjin during the Spring Festival period of 2015.

349 *Atmospheric and Oceanic Science Letters*, 9(1), 15–21.

350 <https://doi.org/10.1080/16742834.2015.1131948>

351 Verstraeten, W. W., Neu, J. L., Williams, J. E., Bowman, K. W., Worden, J. R., &

352 Boersma, K. F. [K. Folkert] (2015). Rapid increases in tropospheric ozone

353 production and export from China. *Nature Geoscience*, 8(9), 690–695.

354 <https://doi.org/10.1038/ngeo2493>

355 Wang, C., Huang, X.-F., Zhu, Q., Cao, L.-M., Zhang, B., & He, L.-Y. (2017).

356 Differentiating local and regional sources of Chinese urban air pollution based on

357 the effect of the Spring Festival. *Atmospheric Chemistry and Physics*, 17(14), 9103–

358 9114. <https://doi.org/10.5194/acp-17-9103-2017>

359 Yao, L., Wang, D., Fu, Q. [Qingyan], Qiao, L., Wang, H., Li, L., et al. (2019). The

360 effects of firework regulation on air quality and public health during the Chinese

361 Spring Festival from 2013 to 2017 in a Chinese megacity. *Environment*

362 *International*, 126, 96–106. <https://doi.org/10.1016/j.envint.2019.01.037>

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