



**Abstract**

Reflecting recent advances in our understanding of soil organic carbon (SOC) turnover and persistence, a new generation of models increasingly makes the distinction between the more labile soil particulate organic matter (POM) and the more persistent mineral-associated organic matter (MAOM). Unlike the typically poorly defined conceptual pools of traditional SOC models, the POM and MAOM pools can be directly measured for their carbon content and isotopic composition, allowing for pool-specific data assimilation. However, the new-generation models' predictions of POM and MAOM dynamics have not yet been validated with pool-specific carbon and  $^{14}\text{C}$  observations. In this study, we evaluate 5 influential and actively developed new-generation models (CORPSE, Millennium, MEND, MIMICS, SOMic) with pool-specific and bulk soil  $^{14}\text{C}$  measurements of 77 mineral topsoil profiles in the International Soil Radiocarbon Database (ISRaD). We find that all 5 models consistently overestimate the  $^{14}\text{C}$  content ( $\Delta^{14}\text{C}$ ) of POM by 67‰ on average, and 3 out of the 5 models also strongly overestimate the  $\Delta^{14}\text{C}$  of MAOM by 74‰ on average, indicating that the models generally overestimate the turnover rates of SOC and do not adequately represent the long-term stabilization of carbon in soils. These results call for more widespread usage of pool-specific carbon and  $^{14}\text{C}$  measurements for parameter calibration, and may even suggest that some new-generation models might need to restructure their simulated pools (e.g. by adding inert pools to POM and MAOM) in order to accurately reproduce SOC dynamics.

**1 Introduction**

The terrestrial carbon reservoir sequesters an estimated 29% of anthropogenic  $\text{CO}_2$  emissions each year (Friedlingstein et al., 2022), significantly reducing the accumulation rate of  $\text{CO}_2$  in the atmosphere and thus slowing down climate change. However, the future role of the terrestrial carbon reservoir as a net  $\text{CO}_2$  sink is uncertain, as Earth System Models (ESMs) produce a wide range of projections for the net land-atmosphere carbon flux over the course of the 21st century, partly due to high uncertainties in the carbon-climate feedback (Friedlingstein et al., 2014; Arora et al., 2020). Moreover, a study by He et al. (2016) using the radiocarbon ( $^{14}\text{C}$ ) isotope suggests that some of the most widely used CMIP5 (Coupled Model Intercomparison Project Phase 5) ESMs may be systematically overestimating the future land carbon sink, further casting doubt on the reliability of future land sink predictions. All five ESMs tested in their study strongly underestimated the  $^{14}\text{C}$  age of soil organic carbon, which indicates an overestimation of the simulated carbon cycling rates, particularly in the most stable soil carbon pools. After He et al. (2016) adjusted the soil carbon cycling rates to fit the observed  $^{14}\text{C}$  data, the ESMs ended up predicting  $40\pm 27\%$  lower carbon sequestration by the terrestrial sink in the 21st century than with their default parameters. This result puts into question the ability of current ESMs to accurately model soil carbon dynamics, and highlights the importance of validating model predictions with  $^{14}\text{C}$  data.

Almost all ESMs rely on soil organic carbon (SOC) modules that are ultimately based either on the Century model (Parton et al., 1987) (e.g., CESM2, Danabasoglu et al., 2020) or the RothC model (Coleman & Jenkinson, 1996) (e.g., JULES, Clark et al., 2011). Even though Century and RothC have been used for many decades to predict SOC dynamics in various landscapes with moderate success (Leifeld et al., 2008; Leifeld, 2008; Leifeld et al., 2009; Abramoff et al., 2022; H. Zhang et al., 2020), both modeling frameworks were developed in the 1980s, and thus reflect the comparatively limited understanding of soil carbon cycling of that time. Indeed, the model design of RothC is inspired by the now obsolete humification theory (Lehmann & Kleber, 2015; Schmidt et al., 2011), and neither RothC nor Century explicitly simulate specific processes of SOC cycling, such as physico-chemical protection of SOC or adsorption and desorption of dissolved organic carbon, because their mechanisms were previously not understood well enough.

65 According to our current understanding, the most important control on SOC sta-  
66 bility is not so much the molecular composition or “quality” of organic matter, but rather  
67 its protection from microbial and abiotic decomposition through occlusion in aggregates  
68 and mineral association (Kleber et al., 2011; Dungait et al., 2012; Lehmann & Kleber,  
69 2015; Lavallee et al., 2020). When SOC gets enclosed into aggregates or stabilized onto  
70 soil mineral surfaces through the action of pedogenic oxides, in particular iron, aluminum  
71 and calcium associated with clay particles (Rasmussen, Heckman, et al., 2018; Rowley  
72 et al., 2018; Vogel et al., 2014), it becomes less accessible to decomposers and thus sig-  
73 nificantly increases its residence time in soils (Basile-Doelsch et al., 2020; Schrumppf et  
74 al., 2013; Doetterl et al., 2015). A new generation of SOC models is now being devel-  
75 oped to incorporate the theory of SOC protection through occlusion and interactions with  
76 soil minerals into our carbon cycle predictions. A common feature of new-generation soil  
77 models is their distinction between particulate organic matter (POM) and mineral-associated  
78 organic matter (MAOM). The POM pool largely consists of partially decomposed lit-  
79 ter fragments smaller than 2 mm (Lavallee et al., 2020; Basile-Doelsch et al., 2020), which  
80 are usually covered with a thin mineral coating (Wagai et al., 2009). On the other hand,  
81 the MAOM pool contains organic matter chemically adsorbed onto reactive mineral sur-  
82 faces, as well as strongly bound micro-aggregates formed around sand, silt, or clay par-  
83 ticles (Basile-Doelsch et al., 2020; Lavallee et al., 2020). Unlike the carbon pools of RothC  
84 and Century, the POM and MAOM pools of the new-generation models can be opera-  
85 tionally defined with experimental protocols by which they can be separated from soil  
86 samples and then analyzed individually for their elemental and isotopic composition (von  
87 Lützow et al., 2007). This allows for a closer look into the processes governing soil car-  
88 bon stabilization and for potentially much larger datasets for model calibration and val-  
89 idation. However, the use of pool-specific measurements to validate models is still lim-  
90 ited, even for new-generation models (Y. Zhang et al., 2021, Table S1).

91 The theory that protection and accessibility are the most important controls on  
92 SOC stability is strongly supported by  $^{14}\text{C}$  studies (Gaudinski et al., 2000; Schrumppf et  
93 al., 2013, 2021), which could indicate that new-generation SOC models might perform  
94 better with  $^{14}\text{C}$  than the traditional SOC models integrated into ESMs.  $^{14}\text{C}$  is an effec-  
95 tive carbon cycle tracer because it is chemically indistinguishable from the other carbon  
96 isotopes and therefore participates in the same carbon exchange mechanisms as the more  
97 abundant  $^{12}\text{C}$  and  $^{13}\text{C}$  isotopes. Over the past century, the atmospheric  $^{14}\text{C}$  levels have  
98 undergone dramatic changes, most notably as a result of thermonuclear weapons tests  
99 in the 1950s and '60s, which have almost doubled the amount of atmospheric  $^{14}\text{CO}_2$  in  
100 the Northern Hemisphere (see Figure 2). As this bomb-derived  $^{14}\text{CO}_2$  spreads into the  
101 terrestrial carbon reservoirs through photosynthesis and into oceans through air-sea gas  
102 exchanges (Graven et al., 2020), the level of enrichment in bomb-derived  $^{14}\text{C}$  across dif-  
103 ferent terrestrial and oceanic carbon reservoirs helps to evaluate the speed and magni-  
104 tude of carbon exchanges with the atmosphere on annual and decadal scales. Meanwhile  
105 for slower-cycling reservoirs such as deep soils or permafrost, the level of  $^{14}\text{C}$  depletion  
106 due to radioactive decay (half-life of  $5700\pm 30$  years (Roberts & Southon, 2007)) helps  
107 to estimate the time scales of carbon stabilization in those reservoirs on the order of cen-  
108 turies and millennia.  $^{14}\text{C}$  is therefore a powerful tool to study the exchanges and stor-  
109 age of carbon from decadal to millennial time scales. However, new-generation models  
110 do not generally implement  $^{14}\text{C}$  simulations, and only a handful have systematically as-  
111 similated observed  $^{14}\text{C}$  data (e.g., Tipping & Rowe, 2019; Braakhekke et al., 2014; Ahrens  
112 et al., 2020).

113 In this study, we use  $^{14}\text{C}$  measurements of the organic carbon in the mineral top-  
114 soil to evaluate the performance of five new-generation SOC models: CORPSE (Sulman  
115 et al., 2014), MEND-new (G. Wang et al., 2022), Millennial v2 (Abramoff et al., 2022),  
116 MIMICS-CN v1.0 (Kyker-Snowman et al., 2020), and SOMic 1.0 (Woolf & Lehmann,  
117 2019). These models were chosen because they are open source, actively developed, and  
118 influential in the soil modeling community. Leveraging the measurability of their pools,

119 we compare these models' predictions to  $^{14}\text{C}$  measurements of POM and MAOM, in ad-  
 120 dition to the total soil  $^{14}\text{C}$ . This provides a detailed picture of the modeled SOC dynam-  
 121 ics and enables us to carry out an in-depth analysis of the models' performances.

## 122 2 Methods

123 Throughout this paper, we report the  $^{14}\text{C}$  content of a given carbon sample as  $\Delta^{14}\text{C}$ ,  
 124 which is the deviation of the sample's  $^{14}\text{C}/^{12}\text{C}$  ratio from the "modern" standard, cor-  
 125 responding to the pre-industrial atmospheric  $^{14}\text{CO}_2/^{12}\text{CO}_2$  ratio (Trumbore et al., 2016).

### 126 2.1 Pool-specific carbon and radiocarbon measurements

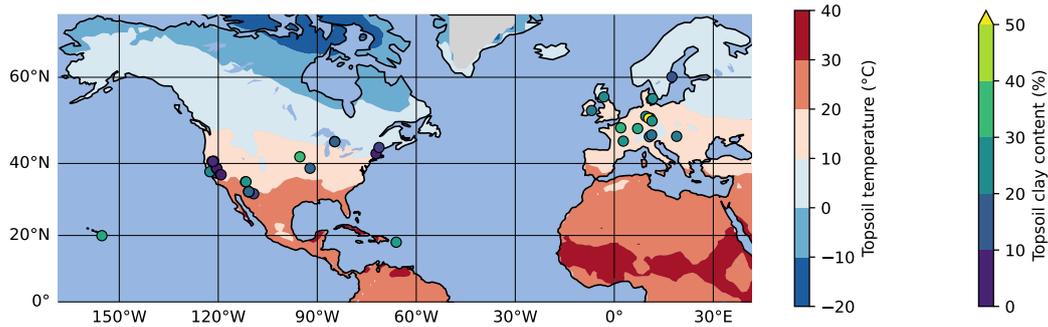
127 We compare model predictions to three types of measured data for the topsoil: (1)  
 128 the total SOC stocks in the topsoil, (2) the relative mass contributions of POM and MAOM  
 129 to the total SOC stocks, and (3) the  $\Delta^{14}\text{C}$  of POM, MAOM, and bulk SOC.

130 For this study, we will use the International Soil Radiocarbon Database (ISRaD)  
 131 (Lawrence et al., 2020) for carbon and  $^{14}\text{C}$  measurements of POM and MAOM obtained  
 132 from soil samples using a combination of density fractionation and ultra-sonication. Den-  
 133 sity fractionation with ultra-sonication is currently one of the most effective and com-  
 134 monly employed methods for separating POM and MAOM (Golchin et al., 1994; Griepen-  
 135 trog et al., 2015, 2014; Cerli et al., 2012; von Lützow et al., 2007; Poeplau et al., 2018).  
 136 This method separates the soil into three "density fractions" – the free light fraction, oc-  
 137 cluded light fraction, and heavy fraction – in a three step process: (1) obtain the free  
 138 light fraction from the soil sample by density fractionation; (2) in the remaining sam-  
 139 ple, destroy loosely-bound aggregates with ultra-sonication, thus releasing the occluded  
 140 fraction; (3) isolate the occluded light fraction from the relatively denser heavy fraction  
 141 by density fractionation. The resulting free and occluded light fractions correspond ap-  
 142 proximately to the POM pool, while the heavy fraction is a good proxy for the MAOM  
 143 pool (Mikutta et al., 2019; Lavalley et al., 2020). We will from now on refer to the soil  
 144 density fractions (light and heavy) by the names of the corresponding pools (POM and  
 145 MAOM, respectively).

146 ISRaD provides carbon and  $^{14}\text{C}$  data for the bulk soil, and the free light, occluded  
 147 light, and heavy fractions. We derive the relative carbon contributions and  $\Delta^{14}\text{C}$  of POM  
 148 with a weighted average of the free and occluded light fractions, and we directly asso-  
 149 ciate MAOM with the heavy fraction in ISRaD. When the  $\Delta^{14}\text{C}$  of the bulk soil is not  
 150 measured or reported in ISRaD, we calculate it with a weighted average of POM  $\Delta^{14}\text{C}$   
 151 and MAOM  $\Delta^{14}\text{C}$ . Since most of the available  $^{14}\text{C}$  data is for the topsoil, we will eval-  
 152 uate models only for the top 5 cm or top 10 cm of the mineral soil. The current version  
 153 of ISRaD (v 2.5.5.2023-09-20) contains complete  $^{14}\text{C}$  datasets of the POM and MAOM  
 154 density fractions in the topsoil of 77 soil profiles spread across 39 sampling sites, cover-  
 155 ing forests, shrubland, cultivated landscapes, and rangeland and grassland. Almost all  
 156 of the sampling sites are in North America and Europe, and the remaining sites are lo-  
 157 cated in Hawaii and Puerto Rico (see map in Figure 1). The dataset does not contain  
 158 any permafrost, thermokarst, peatland, or wetland soils, and 75 of the 77 samples are  
 159 from 1997-2015, with only one sample from 1949 and one sample from 1978. As shown  
 160 in Figure 2, most datapoints bear a positive  $\Delta^{14}\text{C}$  value, demonstrating an enrichment  
 161 in bomb-derived  $^{14}\text{C}$  in the topsoil.

### 162 2.2 Selection of new-generation models

163 We reviewed the literature to find new-generation models whose pools are fully com-  
 164 patible with the observed POM and MAOM density fractions, and that have already been  
 165 tested with a range of soil types and environments. Table 1 gives an overview of the fea-  
 166 tures and capabilities of such new-generation models, almost all of which have been de-



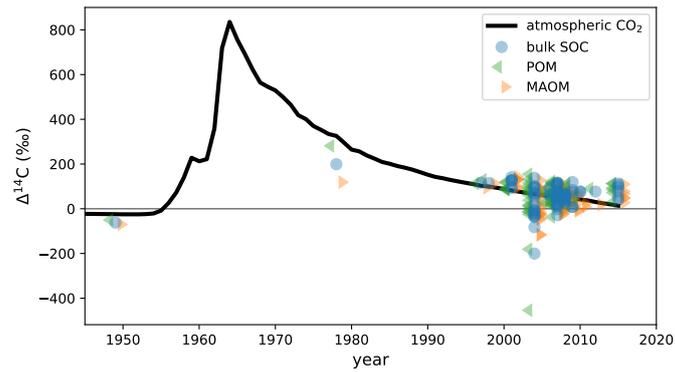
**Figure 1.** Map of selected topsoil sampling sites from ISRaD (Lawrence et al., 2020). 37 of the 39 sites are located in North America and Europe, and the two remaining sites are in Hawaii and Puerto Rico. All sites have a complete  $^{14}\text{C}$  dataset for bulk soil and all density fractions for the top 5 or 10 cm of the mineral soil. The map also shows two of the most important environmental controls on soil carbon persistence: soil temperature (at 4 cm depth, averaged over 1970-2010 period, 1 degree horizontal resolution) from the CESM2 Large Ensemble product (Rodgers et al., 2021) on the map background, and clay content in the topsoil from ISRaD or SoilGrids (Poggio et al., 2021) for each sampling site.

167 developed starting in the 2010s. Many new-generation SOC models also explicitly repre-  
 168 sent the microbial biomass as a separate carbon pool, since microbes are the main drivers  
 169 of SOC turnover (Crowther et al., 2019; Basile-Doelsch et al., 2020; Schimel, 2023). The  
 170 newest version of the MEND model simulates a variety of microbial exo-enzyme pools  
 171 in addition to its microbial biomass pools (G. Wang et al., 2022). About half of the mod-  
 172 els listed in Table 1 have already been implemented with  $^{14}\text{C}$ . However, none of them  
 173 have systematically assimilated fraction-specific  $^{14}\text{C}$  data, instead relying on  $^{14}\text{C}$  data  
 174 of bulk SOC or  $^{14}\text{CO}_2$  data from soil respiration.

175 For this  $^{14}\text{C}$  study, we chose to evaluate the following models, as their code is open-  
 176 source and they have produced successful SOC predictions for a variety of ecosystems:

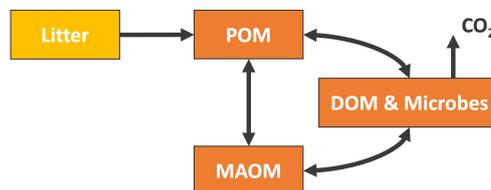
- 177 • Millennial v2 (with Michaelis-Menten kinetics), Abramoff et al. (2022),
- 178 • SOMic 1.0, Woolf and Lehmann (2019),
- 179 • MEND-new (with default equations), G. Wang et al. (2022),
- 180 • CORPSE (version from GitHub repository bsulman/CORPSE-fire-response), first  
 181 described in Sulman et al. (2014),
- 182 • MIMICS-CN v1.0, Kyker-Snowman et al. (2020).

183 Figure 3 shows the general structure of the above models. All the selected mod-  
 184 els have pools which we can associate to the POM and MAOM fractions (see section S2  
 185 in the Supporting Information for details on how we associate the pools to each fraction),  
 186 and they all have at least one microbial biomass pool. We generally chose to evaluate  
 187 the most recent version of each model. However, we found an error in the  $^{14}\text{C}$  implemen-  
 188 tation of the most recent version of MIMICS (Y. Wang et al., 2021) (see section S4.2 in  
 189 the Supporting Information), so we chose to use the coupled carbon-nitrogen version MIMICS-  
 190 CN published one year prior in Kyker-Snowman et al. (2020). See section S1 and Fig-  
 191 ures S1-S5 in the Supporting Information for more details on the exact versions and im-  
 192 plementations of each model.



**Figure 2.** Measured  $\Delta^{14}\text{C}$  data of the POM and MAOM density fractions and total soil organic carbon (SOC) at the selected topsoil profiles from ISRaD (Lawrence et al., 2020), overlaid on the atmospheric  $\Delta^{14}\text{CO}_2$  curve of the Northern Hemisphere (Graven et al., 2017). All POM and MAOM fractions shown here were produced using the method of density fractionation with ultra-sonication. These ISRaD data were originally published in Baisden et al. (2002); Berhe et al. (2012); Harden et al. (2002); Heckman (2010); Heckman et al. (2018); Lybrand et al. (2017); Marín-Spiotta et al. (2008); McFarlane et al. (2013); Meyer et al. (2012); Rasmussen, Throckmorton, et al. (2018); Schrumpf et al. (2013).

193 Note that the MIND model (Fan et al., 2021) would have been a great candidate  
 194 for evaluation, too, but only a subset of the modeled pools was run globally, so some of  
 195 its parameters (e.g.  $V_{\max,P}$  and  $K_{M,P}$ ) do not have fitted values outside of 4 experimen-  
 196 tal test cases.



**Figure 3.** General structure of the new-generation models which we chose for this study. The MIMICS and CORPSE models additionally feature a  $\text{CO}_2$  flux leaving MAOM and POM, which depends on the carbon use efficiency of the microbes. The SOMic and CORPSE models do not allow any flux from the DOM, Microbe, or MAOM pools back into the POM pool. More detailed diagrams for the MEND, Millennial, SOMic, CORPSE, and MIMICS models can be found in the Supporting Information (Figures S1-S5). Abbreviations: POM = particulate organic matter ; MAOM = mineral-associated organic matter ; DOM = dissolved organic matter.

### 197 2.3 Model input data

198 For each measurement site, the models are run with local environmental forcing  
 199 data from 1850 to 2014. The initial conditions in 1850 are found by spinning up the mod-  
 200 els, looping over a “pre-industrial” year, where the forcing data is averaged over the 1850-  
 201 1879 period, until the system reaches equilibrium, i.e. does not experience any signifi-

**Table 1.** Summary of features and capabilities of new-generation models. All of the listed models are compatible with the distinction between POM and MAOM and have been used to produce predictions for a variety of soil profiles. The models selected for evaluation with  $^{14}\text{C}$  in this study are indicated with an asterisk (\*). The first two columns are the year of the first publication and, if applicable, the year of the latest published revision of each model at the time of writing. The “Open-source,” “Implements  $^{14}\text{C}$ ,” and “Explicitly models” columns are checkmarked if at least one version of the model has open-source code, implements  $^{14}\text{C}$  simulations, or explicitly models a specified pool or feature, respectively. In the “Vertical mixing” subcolumn, models with a downward arrow ( $\downarrow$ ) simulate any kind of downward transport or leaching for at least one of their pools, often in dissolved form, and sometimes using an advection equation. Models featuring an up-down arrow ( $\updownarrow$ ) additionally implement vertical mixing for at least one of their pools with a diffusion equation.

Model name	First publication	Latest revision	Open-source	Implements $^{14}\text{C}$	Explicitly models				Notes
					DOM	Microbes	Enzymes	Vertical mixing	
* Millennial <sup>1</sup>	2018	2022	✓		✓	✓		$\downarrow$	
* SOMic <sup>2</sup>	2019		✓	✓	✓	✓		$\downarrow$	
* MEND <sup>3</sup>	2013	2022	✓	✓	✓	✓	✓		$^{14}\text{C}$ only in 2015
* CORPSE <sup>4</sup>	2014	2020	✓			✓			
* MIMICS <sup>5</sup>	2014	2021	✓	✓		✓		$\updownarrow$	$^{14}\text{C}$ and $\updownarrow$ only in 2021
MIND <sup>6</sup>	2021		✓			✓			
AggModel <sup>7</sup>	2013		✓						incubation model
JSM <sup>8</sup>	2020		(✓)	✓	✓	✓		$\updownarrow$	source code accessible upon request
COMISSION <sup>9</sup>	2015	2020		✓	✓	✓		$\updownarrow$	$^{14}\text{C}$ introduced in v2.0
Tipping & Rowe <sup>10</sup>	2019			✓	✓			$\downarrow$	
MEMS <sup>11</sup>	2019	2021			✓	✓		$\updownarrow$	$\updownarrow$ introduced in v2.0
SOMPROM <sup>12</sup>	2011	2014		✓				$\updownarrow$	$^{14}\text{C}$ introduced in 2014
CAST <sup>13</sup>	2013							$\downarrow$	
Struc-C <sup>14</sup>	2009								
PROCAAS <sup>15</sup>	2020								incubation model

<sup>1</sup>Abramoff et al. (2018, 2022) ; <sup>2</sup>Woolf and Lehmann (2019) ; <sup>3</sup>G. Wang et al. (2013, 2015, 2022) ; <sup>4</sup>Sulman et al. (2014, 2017); Salazar et al. (2018); Hicks Pries et al. (2018); Moore et al. (2020) ; <sup>5</sup>Wieder et al. (2014, 2015); H. Zhang et al. (2020); Kyker-Snowman et al. (2020); Y. Wang et al. (2021) ; <sup>6</sup>Fan et al. (2021) ; <sup>7</sup>Segoli et al. (2013) ; <sup>8</sup>Yu et al. (2020) ; <sup>9</sup>Ahrens et al. (2015, 2020) ; <sup>10</sup>Tipping and Rowe (2019) ; <sup>11</sup>Robertson et al. (2019); Y. Zhang et al. (2021) ; <sup>12</sup>Braakhekke et al. (2011, 2013, 2014) ; <sup>13</sup>Stamati et al. (2013) ; <sup>14</sup>Malamoud et al. (2009) ; <sup>15</sup>Liu et al. (2020)

202 cant inter-annual variability. More details on the spinup methods for each model are given  
203 in section S1 in the Supporting Information.

204 The selected models require a number of constant and time-dependent forcing data  
 205 to be run at each study site. We assume that soil properties such as sand, clay and silt  
 206 content, soil density, and land use are time-invariant since pre-industrial times. Where  
 207 these site-specific soil properties are not reported in ISRaD, they are taken from the Soil-  
 208 Grids database (Poggio et al., 2021). The MIMICS model also requires the lignin con-  
 209 tent of litter inputs, which we set to be a constant value depending only on the land use  
 210 type. We assume that the lignin content is 25% for forest litter and 7% for shrubland  
 211 litter (Rahman et al., 2013, Table 1). For grassland and cultivated landscapes, we as-  
 212 sume a lignin content of 9% based on measurements of grasses at the seeding stage (Armstrong  
 213 et al., 1950, Table 1). Weather-dependent and other dynamic environmental properties,  
 214 such as soil temperature and  $^{14}\text{C}$  influx, are taken from global model predictions with  
 215 monthly time resolution. We use the monthly averaged CESM2 Large Ensemble (CESM2-  
 216 LE) product (Rodgers et al., 2021) for vertically resolved soil temperature and moisture,  
 217 above- and below-ground net primary production (NPP), total gross primary produc-  
 218 tivity (GPP), and the carbon-to-nitrogen ratio and  $\Delta^{14}\text{C}$  of total litter carbon from 1850  
 219 to 2014 with 1 degree spatial resolution. Since the below-ground NPP from the CESM2-  
 220 LE output is not vertically resolved, we derive the topsoil portion of the below-ground  
 221 NPP using the exponential function model from Xiao et al. (2023). For nitrogen depo-  
 222 sition rates, we use monthly data simulated by the NCAR Chemistry-Climate Model Ini-  
 223 tiative (CCMI) on a 0.5 degree grid from 1860 to 2016 (Tian et al., 2018). We extend  
 224 this data back to 1850 by setting the monthly nitrogen deposition rates for the 1850-1860  
 225 period to be equal to the average monthly rates over the 1860-1870 period.

226 Since none of the selected models represent lateral carbon transport or upward ver-  
 227 tical mixing of soil carbon, the simulated topsoil systems receive all of their carbon ex-  
 228 clusively from vegetation inputs. We can therefore estimate the carbon influx into the  
 229 soil with the NPP, and the  $\Delta^{14}\text{C}$  of the influx with the  $\Delta^{14}\text{C}$  of litter from the CESM2-  
 230 LE product. In the case of the MEND model, we use GPP instead of NPP as a model  
 231 input, as prescribed by MEND’s developers.

### 232 3 Results

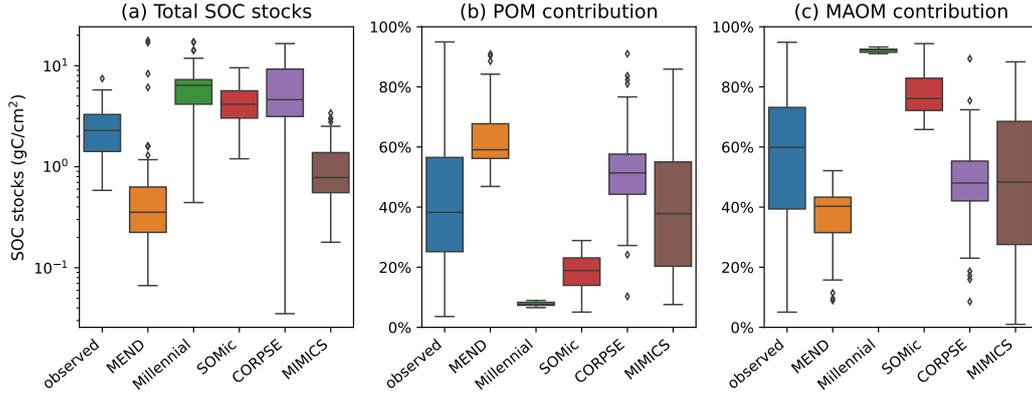
233 We produced carbon and  $^{14}\text{C}$  predictions with the MEND, Millennial, SOMic, CORPSE  
 234 and MIMICS models for the 77 selected soil profiles, and compared them to the observed  
 235 carbon and  $^{14}\text{C}$  data from ISRaD. Our main performance metrics are the root mean squared  
 236 error (RMSE) and mean bias of the predictions with respect to the 6 observational datasets  
 237 described in Section 2.1. Table 2 gives a summary of the model performances, and Fig-  
 238 ures S8-S12 in the Supporting Information show plots of predictions against observations  
 239 for each variable and each model. Note that the MEND model failed to run on 12 of the  
 240 77 selected soil profiles due to some numerical instability, and was unable to produce  $^{14}\text{C}$   
 241 data for 3 other profiles. Note also that the SOC stocks for 17 of the 77 selected profiles  
 242 are not available in ISRaD.

#### 243 3.1 Carbon stocks and partitioning between pools

244 The SOMic, Millennial, and CORPSE models tend to overestimate the topsoil SOC  
 245 stocks of the selected soil profiles, while MEND and MIMICS underestimate the SOC  
 246 stocks, as shown in Figure 4a. In their predictions of SOC partitioning into POM and  
 247 MAOM, the new-generation models generally fail to cover the full range of variability  
 248 in the observations, with the exception of the MIMICS model (see Figure 4b-c). The CORPSE  
 249 and MIMICS models perform the best, and both have a RMSE of around 20 percent-  
 250 age points, and a bias of around 10 points or less in magnitude. Meanwhile, the remain-  
 251 ing models have an average RMSE of 35 points and an average absolute bias of around  
 252 25 points in their predictions of POM and MAOM contributions to total SOC stocks (see  
 253 Table 2).

**Table 2.** Root mean squared error (RMSE) and mean bias for each model and each dataset. In the case of the MEND model, the RMSE and bias were calculated based on results of  $n = 62$  profiles for the  $\Delta^{14}\text{C}$  rows,  $n = 52$  for SOC stocks, and  $n = 65$  for the rows of POM and MAOM contributions. For all other models,  $n = 77$  for all rows, except SOC stocks, where  $n = 60$ .

		MEND	Millennial	SOMic	CORPSE	MIMICS	Average
Bulk SOC $\Delta^{14}\text{C}$ (‰)	RMSE	84	115	101	90	80	94
	Bias	+59	+69	+46	+35	0	+42
POM $\Delta^{14}\text{C}$ (‰)	RMSE	94	120	100	119	129	112
	Bias	+50	+63	+56	+86	+80	+67
MAOM $\Delta^{14}\text{C}$ (‰)	RMSE	103	117	102	83	74	96
	Bias	+83	+82	+57	-3	-39	+36
SOC stocks (kgC/m <sup>2</sup> )	RMSE	4.1	3.8	3.2	6.2	2.3	3.9
	Bias	-1.3	+2.7	+1.9	+4.0	-1.6	+1.1
POM contribution (%)	RMSE	35	40	32	23	17	29
	Bias	+24	-33	-22	+11	-2	-4
MAOM contribution (%)	RMSE	35	41	30	21	21	30
	Bias	-24	+35	+20	-9	-9	+2



**Figure 4.** Observed and modeled total SOC stocks in the topsoil (top 5 or 10 cm of mineral soil) plotted on a log-transformed axis in subplot (a), and contributions of the POM and MAOM pools to the topsoil SOC stocks in subplots (b) and (c), respectively. Black diamonds are outliers. In (a),  $n = 60$  for the boxplot of observed data,  $n = 65$  for MEND, and  $n = 77$  for all other models. In (b) and (c),  $n = 77$  for all boxplots, except for MEND, where  $n = 65$ .

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### 3.2 Performance with $^{14}\text{C}$

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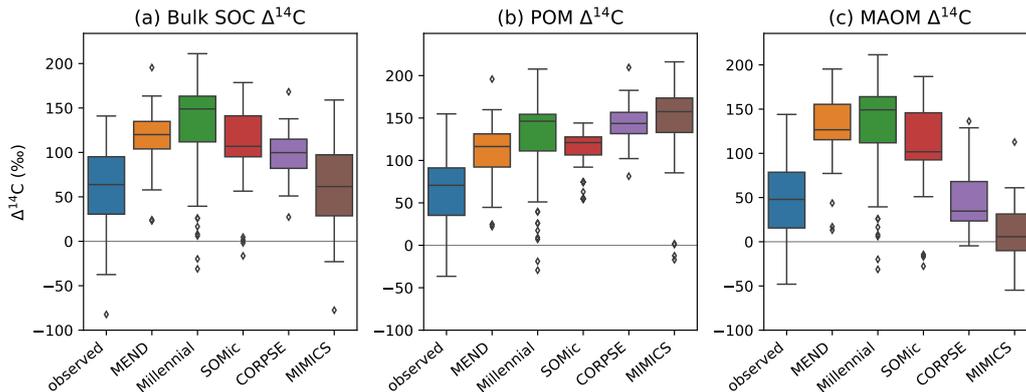
With the notable exception of MIMICS, the new-generation models consistently overestimate the  $\Delta^{14}\text{C}$  of bulk SOC, and their  $^{14}\text{C}$  predictions do not capture the full variability of the observations (see Figure 5a). This is reminiscent of the ESMs'  $^{14}\text{C}$  predictions (He et al., 2016), which also overestimate the  $\Delta^{14}\text{C}$  of SOC and underestimate its variability. Therefore, our results could suggest that the new generation of soil models may be facing similar issues as the traditional SOC models incorporated into ESMs.

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The pool-specific  $^{14}\text{C}$  results, shown in Figure 5b-c, shed a more critical light on the performance of MIMICS with the  $\Delta^{14}\text{C}$  of bulk SOC. MIMICS overestimates the  $\Delta^{14}\text{C}$

263 of POM by 80‰ and underestimates the  $\Delta^{14}\text{C}$  of MAOM by around 40‰ on average,  
 264 and these biases happen to cancel out in such a way that MIMICS produces very good  
 265 predictions for the  $\Delta^{14}\text{C}$  of bulk SOC with a RMSE of just 80‰ and no bias, the best  
 266 performance among the evaluated models (see Table 2). All five models overestimate the  
 267  $\Delta^{14}\text{C}$  of POM, with an average positive bias of 67‰, and SOMic, Millennial, and MEND  
 268 also overestimate MAOM  $\Delta^{14}\text{C}$  by 74‰ on average. CORPSE is good at predicting the  
 269  $\Delta^{14}\text{C}$  of MAOM with effectively no bias, but its POM  $\Delta^{14}\text{C}$  predictions have the largest  
 270 bias (+119‰) among all the models. On average, the evaluated models have a positive  
 271 bias between 36‰ and 67‰, and a RMSE around 100‰ in their  $\Delta^{14}\text{C}$  predictions for  
 272 the POM, MAOM, and bulk SOC (see Table 2 for more details).



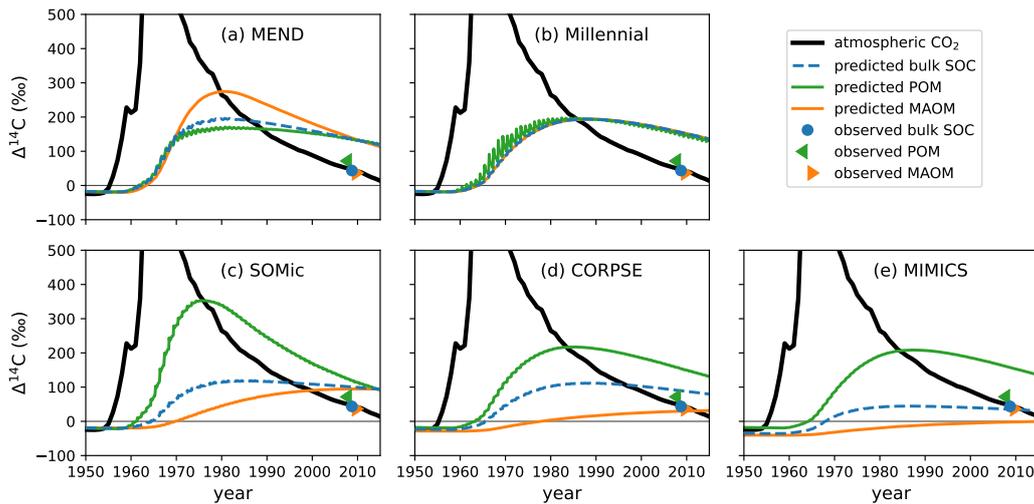
**Figure 5.** Observed and modeled  $\Delta^{14}\text{C}$  of total SOC (a), POM (b), and MAOM (c) in the topsoil (top 5 or 10 cm of mineral soil). Black diamonds are outliers. Note that some extreme outliers are outside of plotting range. To have a uniform and consistent  $^{14}\text{C}$  dataset, we excluded the 1949 and 1978 samples so that we end up with more compact data spanning only 18 years at the tail end of the bomb spike. Therefore,  $n = 75$  for all boxplots, except for MEND's, where  $n = 62$ .

273 The models produce contrasting predictions for the evolution of soil  $^{14}\text{C}$  over the  
 274 second half of the 20th century. In the example of an alpine pasture (Figure 6), we can  
 275 see that the CORPSE, SOMic and MIMICS models predict  $\Delta^{14}\text{C}$  curves for POM which  
 276 are distinct from MAOM, while the MEND and Millennial models produce similar  $\Delta^{14}\text{C}$   
 277 dynamics for POM and MAOM. That is because the Millennial and MEND models have  
 278 faster turnover rates than the other models, and their pools rapidly exchange carbon be-  
 279 tween themselves.

### 280 3.3 Role of environmental parameters

281 We further investigate how simulations depend on soil temperature and clay con-  
 282 tent, as these are considered some of the most important factors controlling SOC turnover  
 283 and persistence (Basile-Doelsch et al., 2020; Leifeld et al., 2009).

284 Higher soil temperatures enhance microbial activity and generally increase the turnover  
 285 rate of carbon in soils (German et al., 2012; Leifeld et al., 2009; Sierra et al., 2015). While  
 286 the observed SOC stocks and POM and MAOM contributions are not correlated with  
 287 temperature (Figure 7a-c), the observed  $\Delta^{14}\text{C}$  of POM, MAOM, and bulk SOC signifi-  
 288 cantly increase with higher temperature (Figure 7d-f), probably due to shorter carbon  
 289 residence times in warmer soils. In contrast, the predicted  $\Delta^{14}\text{C}$  of POM, MAOM, and  
 290 bulk SOC are either uncorrelated or negatively correlated with soil temperature. All of



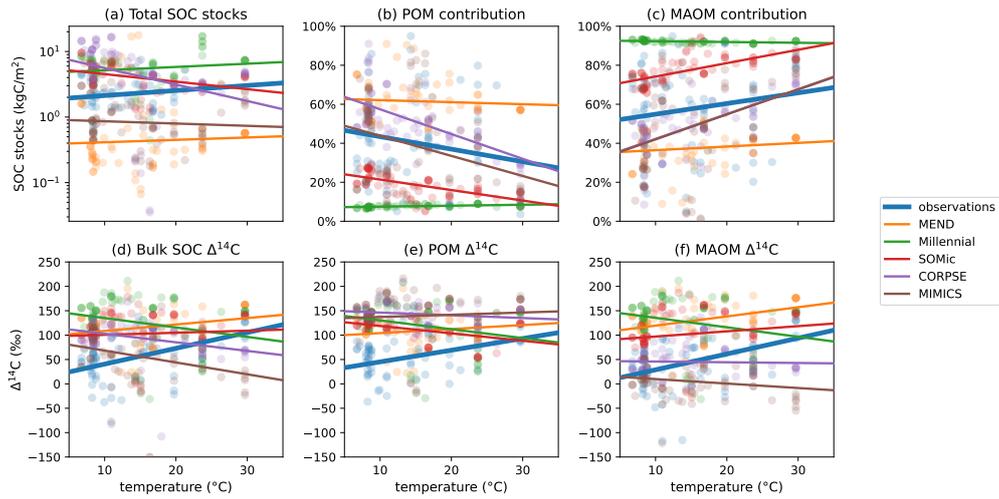
**Figure 6.** Observed and predicted  $\Delta^{14}\text{C}$  of POM, MAOM, and bulk SOC in the top 10 cm of the mineral soil of a pasture in the Matsch valley, Italy. The observed  $^{14}\text{C}$  data from 2008 are published in Meyer et al. (2012). The atmospheric  $\Delta^{14}\text{CO}_2$  of the Northern Hemisphere (Graven et al., 2017) is shown for reference. With the SOMic, CORPSE and MIMICS models, the predicted  $\Delta^{14}\text{C}$  of POM is distinct from the predicted  $\Delta^{14}\text{C}$  of MAOM. On the other hand, the POM and MAOM pools in MEND and Millennial have very similar  $\Delta^{14}\text{C}$  signals throughout the bomb-spike period.

291 the selected models modify carbon decomposition rates with a temperature-dependent  
 292 scaling factor (Abramoff et al., 2022; Woolf & Lehmann, 2019; Kyker-Snowman et al.,  
 293 2020; G. Wang et al., 2022; Sulman et al., 2014), but these results could indicate that  
 294 they may need to increase or change the effect of temperature on carbon turnover rates.

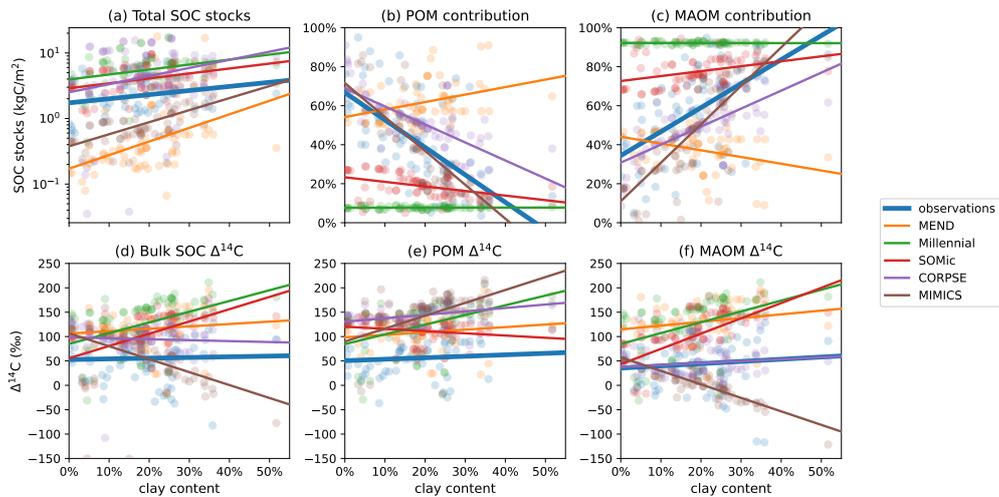
295 In Figure 8c, the clay content of the sampled topsoils seems to be a decisive factor  
 296 controlling the observed contribution of MAOM carbon to the total SOC stocks, with  
 297 higher clay content correlating with higher MAOM contribution. This is also true for  
 298 the MAOM contributions predicted by the MIMICS and CORPSE models, which pro-  
 299 duce the most accurate predictions of MAOM contribution (see Table 2). However, MIM-  
 300 ICS appears to struggle with correctly simulating the effects of increased clay content  
 301 on overall SOC dynamics, as evidenced by the inaccurate relationships of SOC stocks  
 302 and  $\Delta^{14}\text{C}$  with clay (see Figure 8a and Figure 8d-f). It appears that MIMICS correctly  
 303 reproduces the evolution of MAOM contribution with clay content by increasing the resi-  
 304 dence time of carbon in MAOM, which in turn lowers the  $\Delta^{14}\text{C}$  of MAOM and increases  
 305 SOC stocks, contrary to the observations.

## 306 4 Discussion

307 The comparison of new-generation model predictions with  $^{14}\text{C}$  observations reveals  
 308 inaccuracies in the estimations of the time scales of carbon exchanges and stabilization  
 309 in soils. Just like ESMs, most new-generation models overestimate the  $\Delta^{14}\text{C}$  of bulk soil  
 310 organic carbon (SOC) and they, too, may therefore be overestimating the effectiveness  
 311 of soils as a net atmospheric  $\text{CO}_2$  sink in the 21st century (He et al., 2016). The biases  
 312 in the predictions of the repartition of SOC between particulate organic matter (POM)  
 313 and mineral-associated organic matter (MAOM) may also affect the accuracy of future  
 314 projections. POM and MAOM have been shown to have different sensitivities to envi-



**Figure 7.** Relationship of observed and predicted carbon and  $\Delta^{14}\text{C}$  data with respect to mean annual temperature of the topsoil (averaged over the 1970-2010 period). Circles are data-points, and lines are best linear fits through the points. The observed  $\Delta^{14}\text{C}$  of bulk SOC, POM, and MAOM have a strong positive relationship with temperature. Meanwhile, the predicted  $\Delta^{14}\text{C}$  are more weakly and sometimes negatively correlated with temperature. The linear fit line of CORPSE in subplot (c) is completely covered by the linear fit line of MIMICS. Note that we once again excluded the 1949 and 1978 samples for these plots.



**Figure 8.** Relationship of observed and predicted carbon and  $\Delta^{14}\text{C}$  data with respect to clay content in the topsoil. Circles are datapoints, and lines are best linear fits through the points. CORPSE and MIMICS successfully reproduce the positive relationship between topsoil clay content and the observed MAOM contribution to total SOC stocks in subplot (c). However, in subplot (f), MIMICS has a strong negative correlation of MAOM  $\Delta^{14}\text{C}$  with clay content, unlike the observations, which do not show a correlation. The linear fit line of CORPSE in subplot (f) overlaps with that of the observations. Note that we once again excluded the 1949 and 1978 samples for these plots.

ronmental variables such as temperature and are thus expected to react differently to a changing climate (Hicks Pries et al., 2017; Kleber et al., 2011). Therefore, if models do not correctly partition SOC into POM and MAOM and misrepresent their  $^{14}\text{C}$ , they will probably produce inaccurate predictions of SOC dynamics under climate change.

We identify three likely reasons why the new-generation models generally underperform with  $^{14}\text{C}$ , and discuss how these problems could potentially be solved:

1. Insufficient datasets for the calibration of carbon turnover rates,
2. Lack of a “passive” pool with very slow turnover to account for inert SOC components,
3. Modeled pools do not capture the full range of SOC turnover rates.

The last point raises questions on the effectiveness of the new-generation models and the POM-MAOM distinction as a whole. This invites further research on the stability of the different constituents of SOC and a discussion on the most effective way to partition SOC into representative measurable pools.

#### 4.1 Insufficient calibration datasets

Our  $^{14}\text{C}$  results suggest that the new-generation models selected for this study overestimate some carbon turnover rates. The most extreme case is Millennial v2, which gives its micro-aggregate pool and mineral-adsorbed carbon pool turnover times of just a few months (see section S5 of supplement). On the other hand,  $^{14}\text{C}$ -based studies find that the MAOM fraction, which includes micro-aggregates and mineral-adsorbed carbon, typically turns over on time scales of many decades or centuries (Gaudinski et al., 2000; Schrumpp & Kaiser, 2015; van der Voort et al., 2017; Baisden et al., 2002). The overestimation of turnover rates may be due to inadequate or insufficient data for the calibration of the models’ turnover parameters. Even though new-generation models have measurable pools, they do not usually assimilate pool-specific carbon and  $^{14}\text{C}$  data, probably because such data are currently very sparse. The only models in our evaluation to calibrate their parameters with pool-specific carbon data are CORPSE (with data from only 2 soil profiles, according to Y. Zhang et al., 2021, Table S1) and Millennial (as described in Abramoff et al., 2022), and none of them assimilated pool-specific  $^{14}\text{C}$  data. Instead, new-generation models primarily rely on less informative bulk soil data, as well as some soil incubation results, for parameter optimization. However, as the dataset of fraction-specific carbon and  $^{14}\text{C}$  measurements is growing larger, new-generation models should start to take full advantage of the measurability of their pools and assimilate those highly informative data.

#### 4.2 Lack of passive pool

Another explanation for the consistent overestimation of soil  $\Delta^{14}\text{C}$  by new-generation models is the inability of the models to account for the presence of practically inert compounds in the soil, which negatively offset the bulk  $\Delta^{14}\text{C}$ . For example, some soils with a history of wildfires may contain a considerable fraction of pyrogenic carbon, which is composed of highly durable aromatic compounds and can remain in soils over thousands of years (Eckmeier et al., 2009; Hajdas et al., 2007; Leifeld, 2008). Due to its extreme longevity, pyrogenic carbon is depleted in  $^{14}\text{C}$  as a result of radioactive decay, bringing down the overall  $\Delta^{14}\text{C}$  of both POM (van der Voort et al., 2017) and MAOM (Soucémariadin et al., 2019). In deeper soils, the  $\Delta^{14}\text{C}$  of SOC can be even further depleted due to a larger proportion of petrogenic carbon, which is devoid of  $^{14}\text{C}$  (van der Voort et al., 2019). Whereas the two major traditional SOC models explicitly account for such extremely old components with a “passive” pool (1000 year turnover time) in the Century model (Parton et al., 1987) and an “inert organic matter” pool (no turnover at all) in the RothC model (Coleman & Jenkinson, 1996), the new-generation models effectively force virtually in-

ert components to fit into their actively cycling carbon pools. By creating a passive pool to account for inert compounds such as pyrogenic carbon, the new-generation models would be able to lower the overall  $\Delta^{14}\text{C}$  of POM and MAOM, and more accurately reproduce the measured  $^{14}\text{C}$  data.

### 4.3 Search for more representative measurable pools

Finally, the underperformance of the models with respect to  $^{14}\text{C}$  may also be due to a choice of pools which are not truly representative of the full spectrum of turnover rates of the different SOC components. Whereas traditional models simply define the number and turnover rates of their SOC pools such that they can reproduce observed SOC dynamics while minimizing degrees of freedom, new-generation models also need to make sure their pools are at once easily measurable and representative of the various time scales of soil carbon persistence. If a measurable pool contains two or more components with very different turnover rates, the model may not be able to correctly reproduce the  $\Delta^{14}\text{C}$  of that pool because it assumes a single, homogeneous turnover rate for the entire carbon pool. Although some models already split POM into various subpools with contrasting turnover times (e.g., soluble and insoluble litter pools in SOMic, or oxidizable and hydrolysable POM pools in MEND), they miss the most recalcitrant POM pool of pyrogenic carbon, which even in minute amounts can significantly alter the  $\Delta^{14}\text{C}$  and apparent turnover of POM (Leifeld, 2008). Some new-generation models subdivide the MAOM pool into micro-aggregates and mineral-adsorbed carbon (e.g., Millennial), or into an active layer of adsorbed DOC and a more stable MAOM component (e.g., MEND). However, those MAOM subpools might still not be homogeneous enough in their turnover times for effective  $^{14}\text{C}$  simulations. Recent  $^{14}\text{C}$  studies determining the stability of MAOM under the action of peroxide oxidation show that it may be necessary to further split clay-sized MAOM into two measurable subpools which are decomposable or resistant to microbial exo-enzymes (Schrumpf et al., 2021; Jagadamma et al., 2010). Additionally, “continuous” SOC fractionation methods such as ramped pyrolysis oxidation (Stoner et al., 2023) could provide a much higher resolution of the SOC turnover rate spectrum. However, the resulting measurable pools are more difficult to interpret in terms of their role in the soil carbon cycle, and their incorporation into mechanistic SOC models is therefore less straightforward.

### 4.4 Limitations of this study

The accuracy of our model evaluation is affected by multiple factors. Though we took care to accurately match the modeled pools to the measured fractions (see section S2 in Supporting Information), the correspondences are imperfect and further complicated by non-standardized definitions and density cut-offs for the light and heavy fractions published on ISRaD. Nevertheless, this does not change the overall overestimation of soil  $\Delta^{14}\text{C}$  by most models. The use of forcing data from possibly inaccurate CESM2-LE and CCMI outputs with low spatial resolution may also affect the accuracy of our model evaluation. Furthermore, the  $\Delta^{14}\text{C}$  of the carbon inputs from the CESM2-LE product could be inaccurate, especially in soils with a thick organic layer, which pre-ages the carbon before it enters the mineral soil. However, the consistency and magnitude of the models’ overestimation of the topsoil’s  $\Delta^{14}\text{C}$  with respect to observed data indicate that this overestimation is evidently a real pattern among the studied models. Finally, it is also important to note that our study only produces an incomplete picture of model performances on a global scale, since most of the measured datapoints represent North American and European forest ecosystems.

## 5 Summary

Despite their incorporation of the latest advances in soil sciences, new-generation soil organic carbon (SOC) models currently show similar discrepancies with  $^{14}\text{C}$  data as the traditional SOC models. The new-generation models' consistent overestimation of the  $\Delta^{14}\text{C}$  in both particulate organic matter (POM) and mineral-associated organic matter (MAOM) and their inaccurate partitioning of SOC between POM and MAOM suggest that these models underestimate the time scales of carbon storage in soils and might produce unreliable future predictions under climate change. To improve their predictions, new-generation models should take advantage of the measurability of their pools and calibrate their parameters with the rapidly growing dataset of pool-specific carbon and  $^{14}\text{C}$  measurements in addition to incubation and bulk soil data. They may also have to reconsider their model design and simulate measurable pools which better capture the full spectrum of carbon turnover rates present in the soils. In particular, the consideration of highly persistent soil carbon such as pyrogenic carbon could significantly improve  $^{14}\text{C}$  predictions. As more effective measurable pools are being discovered and the dataset of pool-specific  $^{14}\text{C}$  data is expanding, new-generation soil models have the potential to eventually supersede the traditional SOC models employed by ESMs if they take full advantage of the measurability of their pools and assimilate the available data.

## 6 Open Research

The source code to download the input data, run the models, and reproduce all the results presented in this manuscript is available on our GitHub repository <https://github.com/asb219/evaluate-SOC-models>.

Our final implementations of Millennium, CORPSE, MIMICS, and the  $^{14}\text{C}$  component of MEND are available as python modules in our repository. For the carbon and nitrogen components of MEND, the Fortran source code is in <https://github.com/asb219/MEND> (forked from <https://github.com/wanggangsheng/MEND>), which is added as a "git submodule" to our repository. We use the `install_github` function of the `devtools` package in R to compile the C++ code of the SOMic model released as "v1.1-asb219" in <https://github.com/asb219/somic1> (forked from <https://github.com/domwoolf/somic1>) and install it as an R package. We download data from SoilGrids with the `soilgrids` python package (<https://github.com/gantian127/soilgrids>).

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