

# Elemental and biochemical nutrient limitation of zooplankton: a meta-analysis

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## Authorship

PT, MS, HH, and DvdW conceived and designed the meta-analysis. PT and CK extracted data. PT performed the statistical analyses and wrote the draft manuscript. All authors revised the manuscript and approved its final version for submission.

## Data availability statement

The primary data are available from original published studies; the dataset of extracted effect sizes and metadata used for analyses will be uploaded to Dryad upon acceptance; annotated code used for analyses will be uploaded to Zenodo.

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## Abstract

Primary consumers in aquatic ecosystems are frequently limited by the quality of their food, often expressed as phytoplankton elemental and biochemical composition. Effects of these food quality indicators vary across studies, and the relative importance of elemental (nitrogen and phosphorus) versus biochemical (fatty acid and sterol) limitation in aquatic food webs has been debated. Here we present results of a meta-analysis using >100 experimental studies, which confirms that limitation by N, P, essential fatty acids, and sterols all have significant negative effects on zooplankton performance. However, effects varied by grazer response (growth versus reproduction), specific manipulation, and across taxa. P limitation had greater effects on zooplankton growth than fatty acids, but P and fatty acid limitation had equal effects on reproduction. Furthermore, we show that nutrient co-limitation in zooplankton occurs, that indirect effects induced by P limitation exceed direct effects of mineral P limitation, that effects of nutrient amendments using laboratory phytoplankton isolates exceed those using natural field communities, and that algal physiology mediates zooplankton responses to nutrient limitation. Our meta-analysis reconciles contrasting views about the role of various food quality indicators, and their interactions, for zooplankton performance, and provides a mechanistic understanding of how environmental change affects trophic transfer.

## Introduction

The interface between primary producers and their consumers is marked by high variability in the efficiency of energy transfer, which can constrain the functioning of food webs. This is linked to spatial and temporal variation in both the quantity and quality of food resources available to herbivores across the “phytochemical landscape” (Hunter 2016). While classical models in ecology suggest that density-dependent bottom-up forces (i.e., food quantity) determine energy fluxes to higher trophic levels (Lindeman 1942; Hairston *et al.* 1960), more recent evidence suggests that food quality significantly modifies this relationship and that consumer efficiency depends largely on the quality of producers as well (Sterner & Elser 2002; Arts *et al.* 2009). Specifically, in aquatic systems, trophic transfer has been linked to a small subset of nutrients that explain a large proportion of variation in zooplankton production, including nitrogen (N), phosphorus (P), essential fatty acids (EFAs), and sterols. These are all classified as essential nutrients, meaning that if they are in limiting concentrations in the food, zooplankton often face impaired growth and reproduction, even when edible phytoplankton are high in abundance (Sterner *et al.* 1993, Brett and Müller-Navarra 1997, Müller-Navarra *et al.* 2000, Sterner and Elser 2002, Urabe *et al.* 2002).

Despite the known importance of these multiple dimensions of food quality for consumers, there is a long-standing debate regarding whether elemental (i.e., N and P) or biochemical (i.e., EFA and sterol) nutrient limitation is generally more important in shaping zooplankton production, summarized in Gulati and DeMott (1997). Ecological stoichiometry provides a conceptual framework using N and P requirements of producers and consumers to explain trophic interactions, with greater mismatches in algal and consumer C:P ratios in particular causing severe impairment of zooplankton performance (Sterner & Elser 2002). Similarly, variation in the

concentration of individual fatty acid molecules in food was shown to explain zooplankton growth and reproduction in eutrophic lakes (Müller-Navarra *et al.* 2000). There are clear physiological explanations for the importance of elemental N, P, individual EFAs, and sterols. In short, N is needed to build proteins, P-rich RNA is required for growth (Elser *et al.* 1996), EFAs regulate membrane fluidity and act as precursors for signaling molecules involved in reproduction (Parrish 2009), and sterols are vital for membrane structure and hormone synthesis (Martin-Creuzburg & von Elert 2009). Despite these clear mechanisms, effects of manipulating the different indicators of nutritional value vary widely across studies and environmental contexts, resulting in a lack of consensus regarding the overall effects of N, P, EFA, and sterol limitation on zooplankton.

In addition to the individual effects of elemental and biochemical nutrient limitation, there is evidence that these dimensions of food quality can interact with one another. For example, it has been suggested that EFA limitation is only important under replete food P conditions, and vice versa (Gulati & DeMott 1997). This hypothesis implies a degree of co-limitation in the broad sense, in which the effects of adding one nutrient depend on the availability of another. However, there are many possible forms of co-limitation with diverging implications, including simultaneous co-limitation, independent co-limitation, as well as the closely related concept of serial limitation (see Harpole *et al.* (2011) and Sperfeld *et al.* (2016) for a disambiguation of such terms and a guide to the complexity of co-limitation). Moreover, independent co-limitation can be additive, sub-additive, or super-additive, depending upon what effect is induced by co-addition relative to individual addition of nutrients. Whether co-limitation occurs, and in which form it is embodied, however, varies across studies.

In addition to co-limitation, elemental and biochemical nutrient levels in the food often directly co-vary with one another, as well as with other aspects of food quality and the environment. For example, P limitation can simultaneously alter the C:P ratio, fatty acid composition and sterol content of phytoplankton (Müller-Navarra 1995; Ahlgren *et al.* 1998; Klein Breteler *et al.* 2005). Simultaneously, it can also affect their cell size, digestibility, colony size, and toxicity (Lürling & Van Donk 1997; Van Donk *et al.* 1997; Brandenburg *et al.* 2020), thus intrinsically linking these distinct dimensions of food quality. This implies a direct dependence between different components of nutritional quality in the case of, e.g., a causative link between P limitation and EFA content or P limitation and digestibility. Therefore, effects of P manipulation may depend on the extent to which indirect effects (i.e., through biochemical compositional changes) are controlled for in the experimental design. This provides one clear example of how the experimental and environmental context may explain variation in nutrient limitation effects across studies. Other influential factors include the taxonomic identity and diversity of study organisms, the traits they express, whether lab isolates or natural phytoplankton communities were used, and the environmental conditions (e.g., temperature, irradiation) used.

To reconcile variation in observations to date, and the range of factors that can influence the grazer-phytoplankton interface, we conducted a meta-analysis which quantified the general effects of variation in phytoplankton elemental and biochemical nutrient content on zooplankton grazer performance. We searched the literature for studies that manipulated food quality either directly (e.g., via nutrient supplementation) or indirectly (e.g., across species or natural seston type) while also controlling for confounding effects like food quantity. The included studies were comprised of controlled laboratory experiments using assays where the growth, reproduction, or survival of zooplankton were determined. Field observations or mesocosm studies were included only if they had accompanying assays that disentangled effects of changes in food quantity from

food quality. For included studies with direct manipulations and a clear control/treatment structure, we calculated effect sizes as Hedges'  $g$ , whereas for indirect manipulations and gradient designs we calculated Fisher's  $Z$ -transformed correlation coefficients as an effect size. As predictors for both approaches, we extracted: molar N:C and P:C ratios, the concentrations of bulk fatty acids and individual EFAs previously reported to be essential (see Table 1 for descriptions), and total sterol concentrations per unit food carbon. Using this database, we tested the following hypotheses:

H1: Phytoplankton N:C, P:C, EFA content, and sterol content – as measures for nutritional quality – all individually exert a significant influence on zooplankton performance in terms of growth, reproduction, and survival (H1a), but the effect sizes vary by the zooplankton response type (e.g., growth versus reproduction) and across broad nutrient types (H1b).

H2: Zooplankton experience co-limitation by multiple nutrient types, as evidenced by greater effect sizes of co-addition than individual nutrient addition.

H3: Other eco-physiological manifestations of food quality (e.g., digestibility) and variation in the environmental and experimental context across studies significantly influence the magnitude of both elemental and biochemical nutrient limitation.

## Methods

### Data collection and screening

Studies were obtained from a Web of Science ([www.webofscience.com](http://www.webofscience.com)) search (last accessed February 11, 2021) using the search term: *'((stoichiometry or C:N or C:P or "food quality" or "nutritional quality" or "nutritional geometry" or "fatty acid\*" or "biochemical compos\*" or "\*sterol\*") AND (phytoplankton) AND (zooplankton or herbivor\* or consum\* or grazer or*

"filter-feed\*") AND (growth or product\* or defense or grazing or feeding)). We also examined the references of previous reviews involving zooplankton food quality (Brett & Müller-Navarra 1997; Gulati & DeMott 1997; Sterner & Schulz 1998; Frost *et al.* 2005; Martin-Creuzburg & von Elert 2009; Parrish 2009; Hessen *et al.* 2013; Twining *et al.* 2016; Winder *et al.* 2017) to amend the original search. This yielded a total of 1372 articles which we screened, resulting in a total of 122 articles which met all criteria for inclusion. For full documentation of the systematic literature review, see Fig S1 which presents a flow chart of the screening process following PRISMA (preferred reporting items for systematic reviews and meta-analyses) guidelines (O'Dea *et al.* 2021). We included studies that met all the following core criteria: (1) they quantified variation in phytoplankton food quality and/or performed a nutrient supplementation experiment, (2) they quantified zooplankton growth, reproduction, or survival in response to phytoplankton food quality variation, (3) they controlled for confounding effects of food quantity.

We also limited the scope of our synthesis to include only published studies with phytoplankton as a food source for zooplankton. In other words, we did not include studies or portions of studies where only heterotrophic bacteria, benthic algae, terrestrial organic matter, etc. were the sole food supply. Similarly, we also did not include studies with benthic grazers and limited the scope to pelagic zooplankton. Lastly, we limited the scope of food quality parameters to include food N:C, P:C, essential fatty acid content, and sterol content (see Table 1 below). Studies measuring only bulk protein, carbohydrates, or lipids were therefore excluded. Here we refer to stoichiometric ratios in the form of molar "nutrient:C" instead of "C:nutrient" as this is more intuitive (higher nutrient:C ratios reflects higher food quality) and thereby comparable to effects of increasing FA and sterol concentrations per unit food carbon. N:C and P:C ratios were log-transformed for all analyses, as non-transformed ratios can lead to biased results (Isles 2020). While definitions of essential fatty acids vary, we included fatty acids defined as essential by Kainz *et al.* (2004), as



well as several closely related  $\geq 18$ -carbon fatty acids that can be converted to these EFAs, in order to broaden our data beyond only EPA and DHA. Although there is evidence that other nutrients like iron (Lind & Jeyasingh 2018; Jeyasingh & Pulkkinen 2019) and essential amino acids (Fink *et al.* 2011; Koch *et al.* 2013) influence zooplankton production, we did not include these as predictors due to the limited number of experimental case studies.

In addition to the suite of food quality and zooplankton performance variables, we also collected information for an array of categorical and numerical variables to act as covariates, provide further context for each study, and explain heterogeneity across studies, to the extent that this information was available. We included the form of food quality manipulation (direct or indirect, see below), the strength of nutrient manipulation, and the nutrient concentration of phytoplankton both pre- and post-manipulation. We considered taxonomy (from broad to species-level) of algae and zooplankton, and noted whether phytoplankton were isolates or natural seston communities. We further retrieved information on the general environment (freshwater versus marine), grazer life stage, food quantity (as mg C L<sup>-1</sup>), environmental covariates (light, temperature, salinity, pH, pCO<sub>2</sub>, dissolved inorganic N, P, and Si), and experimental parameters (duration, volume, type of experimental units). Particulate Si:C ratios in diatoms were a covariate of interest, as diatom silicification is highly plastic and significantly modifies copepod-diatom interactions (Grønning & Kiørboe 2020; Ryderheim *et al.* 2022); however, this information was notably not reported in studies in our dataset.

The manipulation type can be split into three broad categories: direct manipulations, indirect manipulations established via nutrient supply, and indirect manipulations established via other factors including species composition (Table 2). These three levels represent a hierarchy regarding the strength of inferences that can be made to causally link specific nutrients to

zooplankton performance. The distinction is particularly important for P manipulation due to known indirect effects of P limitation on phytoplankton food quality. Specifically, P manipulation studies were classified as “direct” if they added a spike in dissolved P supply that was rapidly absorbed by P-limited algae in order to increase algal P:C directly prior to feeding to zooplankton (and change nothing else between treatment and control; see e.g., Plath and Boersma (2001) for a description of this method). Studies were classified as “indirect (via nutrient supply)” if they manipulated nutrient supply over the scale of days to weeks in order to create high and low P:C treatments, as they represent a combination of direct P effects as well as the many indirect physiological effects which can manifest in P-limited algae. Note that fatty acids and sterols only fall within the classifications as “direct” (e.g., direct addition via EFA emulsions or liposomes containing EFAs) or “indirect (other)” (e.g., gradients in EFA due to different species composition or environmental factors).

When applicable, we also categorized algae as having important food quality characteristics beyond nutrient content when these were indicated by study authors (e.g., digestion resistance, large size, toxicity). There is a distinction in predicted effects between constitutive (expressed in both high- and low-nutrient conditions) and induced (only expressed with limited nutrients) digestion resistance. Therefore, we differentiated these and classified each case as either constitutive or induced digestion resistance based on descriptions by the authors. All data were extracted from either figures (using WebPlotDigitizer V4.5 (Rohatgi 2021)), tables, text, or original raw data provided by authors (when required data was not available in full texts).

## **Effect size calculations**

For the treatment versus control style studies, we calculated Hedges’  $g$  (a bias-corrected measure of the standardized mean difference) using the function ‘escalc’ in the R package ‘metafor’

(Viechtbauer 2010) using package version 3.0-2 and R version 4.0.3 (R Core Team 2020). We calculated Hedges'  $g$  following Hedges (1983):

$$g = \frac{\mu_{\text{replete}} - \mu_{\text{limited}}}{SD_{\text{pooled}}} \times J \quad \text{where} \quad J = \frac{\Gamma(m/2)}{\sqrt{m/2} \times \Gamma(\frac{m-1}{2})}$$

$$\text{And } SD_{\text{pooled}} = \sqrt{\frac{(n_{\text{replete}} - 1) \times SD_{\text{replete}}^2 + (n_{\text{limited}} - 1) \times SD_{\text{limited}}^2}{n_{\text{replete}} + n_{\text{limited}} - 2}}$$

Here,  $\mu_{\text{replete}}$  and  $\mu_{\text{limited}}$  are the mean zooplankton responses under the most nutrient-replete and most nutrient-limited treatments in each case study, respectively, and  $SD_{\text{pooled}}$  is the pooled standard deviation ( $SD$ ) with  $n$  as the number of replicates for each effect size estimate. The term  $J$  corrects for bias in the estimate of  $g$  at small sample sizes;  $\Gamma$  denotes the gamma function and  $m$  denotes the degrees of freedom such that  $m = n_{\text{replete}} + n_{\text{limited}} - 2$ . The sampling variance was estimated using the unbiased estimator given in Equation 9 of Hedges (1983), as this is preferable for minimizing bias when small sample sizes are commonly used in studies (i.e., using the setting `vtype = "UB"` in the metafor 'escalc' function (Viechtbauer 2010)). Effect sizes were weighted by the inverse of the sampling variance, such that studies with more precise estimates received greater weight. Hedges'  $g$  was preferred over log response ratios because negative or zero values of means were common for certain response variables, in which case log response ratios cannot be used.

For studies with a correlational/gradient design, we calculated effect sizes as correlation coefficients (Pearson's  $r$ ) for all combinations of nutrient type and zooplankton response category. We then converted these values of  $r$  into Fisher's  $Z$ -transformed correlation coefficients (hereafter  $Z_r$ ), a standard effect size used in meta-analysis to improve distributional properties

relative to  $r$  (Koricheva *et al.* 2013). The  $Z$ -transformation ( $Z_r$ ) and its associated sampling variance ( $v_{Z_r}$ ) are as follows, where  $N$  is the number data points used per correlation:

$$Z_r = \frac{1}{2} \ln \left( \frac{1+r}{1-r} \right) \quad v_{Z_r} = \frac{1}{N-3}$$

Of the 122 studies included in our database, 45 could be placed into both broad design categories; we thus calculated both Hedges'  $g$  and  $Z_r$  for these. The pool of studies therefore contains all categories of direct and indirect manipulations for the correlative ( $Z_r$ ) approach, while the Hedges'  $g$  approach includes only direct and nutrient supply-mediated indirect manipulations from which stronger inferences may be made (Table 2), allowing these two datasets to provide complementary information. Agreement between these two distinct effect size methods would add weight to the generality and rigor of the results (shown in Fig. S7).

We reversed the sign of effect sizes for a subset of zooplankton responses for which a smaller value indicates greater performance, such that a reduction in e.g., age at maturity or mortality rate, would be coded as a positive effect size. As the estimation of the sampling variance ( $v_{Z_r}$ ) requires 4 data points, only those datasets containing a gradient of  $\geq 4$  observations were included in the meta-analysis of  $Z$ -transformed correlation coefficients. For both  $g$  and  $Z_r$ , effect sizes of 0 indicate no effect of increased nutrient availability, while positive effect sizes indicate an increase in zooplankton performance with increasing nutrients.

## Statistical models

We used both random and mixed effects models using the 'rma.mv' function in the 'metafor' R package in order to calculate weighted effect sizes, to assess broad differences among effects of each nutrient type, and to identify variables that significantly modify the effects of nutrient limitation. To test H1, that limitation by each broad nutrient type is significant, we ran a separate

weighted random effects model for each combination of nutrient type (N:C, P:C, overall fatty acids as well as specific EFAs, sterols) and broad zooplankton response type (somatic growth, reproduction, survival, and population growth). For Hedges'  $g$ , article ID, zooplankton species, and phytoplankton species were used as random effects to account for non-independence across publications and taxonomy; we specified zooplankton and phytoplankton species as random effects nested within publication. For  $Z_r$ , only article ID was used as the random effect, as one effect size could represent gradients across species, and low degrees of freedom precluded the use of zooplankton species as a random effect.

For both effect size metrics used, we evaluated H1a based on whether the 95% confidence intervals overlap zero (indicating no overall effect), or do not overlap zero (indicating a significant effect). Similarly, non-overlapping confidence intervals between different nutrient types would indicate differences in the magnitude of nutrient limitation effects (H1b). We tested H2 (that zooplankton experience co-limitation) in two ways: first, by comparing the mean and 95% CI of nutrient co-addition to individual nutrient addition (e.g., for studies where both nutrients were factorially manipulated). Second, we used mixed effects models (see below) to test if the concentration of other nutrients significantly influenced the effects of the focal nutrient that was manipulated (e.g., testing if algal P:C ratios influence the magnitude of EFA supplementation effects).

To examine how other experimental and environmental covariates may alter effect sizes across different nutrients (H3), we used additive mixed effect models using the following moderators which are commonly reported for all types of nutrient manipulation: experimental duration and volume, food quantity, realm (marine versus freshwater), and morphological dimensions of food quality. Mixed models using factors specific to each nutrient type were used to identify additional

influential variables. For instance, with P:C manipulation studies, we also included the manipulation method (direct or indirect P manipulation) as a fixed effect. For all nutrients, a separate model tested whether the initial (pre-supplementation) nutrient levels, as well as the strength of manipulation, influenced effect sizes. Last, we used separate mixed effects models to test for variation among lab versus field phytoplankton communities, phytoplankton taxonomy, and zooplankton taxonomy. In cases with low sample size for each factor, we could not use nested random effects as described above, but instead performed mixed effects models using default settings in the ‘rma’ function in the ‘metafor’ package.

## Results

### Individual effects of essential nutrient limitation

The results from both the direct and indirect nutrient manipulations in this meta-analysis indicate strong effects of limitation by all the nutrient types we examined. For somatic growth of zooplankton, effect sizes (as Hedges’  $g$ ) of experimentally increasing N:C ( $g = 2.06$ , 95% CI: 0.94-3.18), P:C ( $g = 2.53$ , 95% CI: 1.93-3.14), fatty acid content ( $g = 1.35$ , 95% CI: 0.94-1.77), and sterol content ( $g = 7.88$ , 95% CI: 1.38-14.4) were all positive and did not overlap 0, indicating significant effects of alleviating their limitation (Fig. 1A, Table S1), and providing clear support for H1a. All broad nutrient types had significant effects on reproduction as well (Fig 1A, Table S1). In addition to the summarized effects of increasing fatty acid content, individual essential fatty acids often had significant positive effects for somatic growth and/or reproduction as well. Overall, similar results were obtained for population growth and survival, yet these responses had relatively limited underlying data compared to somatic growth and reproduction (Fig. S2). Furthermore, these strong effects of nutrient limitation were supported by the correlation results (Fig. 2). Here, effect sizes (as Fisher’s Z-transformed correlation) for

zooplankton performance in relation to N:C, P:C, fatty acid and sterol content were positive and did not overlap zero for any of the broad nutrient types (shown as squares in Fig. 2), which provides additional support for H1a. The  $\omega$ 3 fatty acids (e.g., ALA, SDA, EPA, total  $\omega$ 3) tended to have strong positive correlations with growth and reproduction, often equal in magnitude to that of N:C and P:C, while correlations for  $\omega$ 6 fatty acids (e.g., LIN, GLA, total  $\omega$ 3) did not differ from zero, indicating no effect. Hedges'  $g$  and  $Zr$  were highly congruent with one another for studies where both could be estimated, providing further evidence that support for H1 is robust (Fig. S7).

Though all nutrient types had significant effects, the magnitude differed depending on both the nutrient type and the zooplankton response category, supporting H1b (Fig. 1, Fig. 2, Table S1). For somatic growth as response, the grand mean effect size of P:C manipulation was higher than the grand mean for EFA manipulation (Fig 1A). For reproduction as a response, however, effect sizes of P:C and EFA manipulation were equivalent (i.e., overlapping CIs in Fig.1B). In other words, fatty acids appear to be relatively more important for zooplankton reproduction than for somatic growth. Additionally, there is substantial variation for individual fatty acids with different zooplankton responses. Supplementation of ALA, ARA, EPA, and mixed EFAs all had similar effects on somatic growth, but these effects were less than those of P:C manipulations (Fig 1, Table S1). However, with reproduction as the response, increasing the content of ARA and EPA has effect sizes equivalent in magnitude to increasing P:C ratios.

### **Co-limitation by multiple nutrients**

Data on co-addition of nutrients from factorially designed experiments was limited to 13 studies, but the results clearly indicate that co-limitation occurs as the mean effect of alleviating limitation of multiple nutrients is substantially greater than for individual nutrients in all cases

(Fig. 1). Specifically, a simultaneous increase in P and EFA availability yields a significantly larger effect on zooplankton growth than increasing either nutrient alone. Other co-additions also create larger mean effects than for individual additions, but these are marked by substantial variability and thus have overlapping confidence intervals.

Another indicator of co-limitation is the relationship between algal P:C ratio and the strength of other nutrient manipulations like fatty acids. The available data ( $n = 9$  studies) show that P:C content significantly modifies the effects of EFA addition, with EFA effect sizes generally being smaller with lower P:C ratios (Fig. 3). EFA addition, however, still had some positive effects even under extreme P-limitation, although effects are much stronger when P is replete.

### **Effects of other factors on the strength of nutrient limitation**

All combinations of added nutrients and response variables were characterized by significant heterogeneity in the responses. Part of this heterogeneity could be explained by moderators reflecting study experimental design and organisms used, providing broad support for H3. For example, the method of P manipulation had a very strong effect on both growth and reproduction responses (Fig. 4). Studies implementing direct limitation of elemental P (i.e., the P-spiking method) had significantly lower effects on zooplankton growth ( $g = 1.42$ , 95% CI: 0.77-2.08) than those which also allowed for indirect effects, e.g., via physiological changes over the course of algal culturing under different dissolved P supplies ( $g = 3.20$ , 95% CI: 2.60-3.79).

Phytoplankton derived from lab isolates (either mono- or polycultures) had greater nutrient manipulation effect sizes than natural seston (Fig. 5), indicating that there are systematic differences in how complex natural versus simple artificial communities serve as a food base for zooplankton.



Results from an additive model reveal further key differences in effects depending upon experimental design and food quality factors beyond nutritional quality (Fig. 6). Variation in the food quantity at which nutrients were manipulated had a significant influence on P:C, EFA, and sterol effect sizes. However, these effects were relatively small in magnitude and varied from negative to positive for different nutrient types. Differences in marine versus freshwater study systems were only detected for sterols. Additionally, the effects of manipulating P:C and EFAs were significantly reduced when algae exhibited morphological traits like constitutive digestion resistance (Fig. 6).

Beyond these factors, additional variability in food quality effects can be partially explained by the magnitude of nutrient manipulation, the degree to which nutrients were constrained pre-manipulation, and taxonomy of phytoplankton and grazers. For instance, P:C manipulation is significantly influenced by both the initial (pre-manipulation) P:C ratio, the magnitude of increase, and the interaction between the two (Table S2, Fig S6). For other nutrients, only marginal effects of manipulation strength were detected. Grazer taxonomy also created significant variation in effects. For example, *Bosmina* (a low-P content cladoceran) showed a weaker response to P supplementation compared to *Daphnia*, while rotifers in the genus *Brachionus* had slightly greater responses than *Daphnia* (Fig. S5). Phytoplankton taxonomy provided further variation in effect sizes. For P:C, EFAs, and sterols, adding nutrients to cyanobacteria cultures had greater effects when compared to cryptophytes or green algae. This clearly reaffirms that cyanobacteria are of poor quality while cryptophytes are of superior quality.

## Discussion

### Individual effects of nutrient limitation

The results of this meta-analysis provide a quantitative summary of the food quality limitations zooplankton experience in terms of both elemental and biochemical composition of phytoplankton, two distinct aspects of herbivore nutrition that are typically assessed separately. Our findings support the hypotheses in recent decades that N, P, fatty acid, and sterol limitation impede zooplankton growth and reproduction, and may thereby significantly structure zooplankton communities in nutrient-limited environments. We show that food quantity alone cannot explain fluxes of energy in aquatic systems, and that food quantity variation was, in fact, only a marginally influential factor across the controlled experiments in this study (Fig 6). Though all broad nutrient limitation categories had significant effects (supporting H1a), there were differences depending on the nutrient type and the response assessed (supporting H1b). In most cases, P limitation was more evident for somatic growth than other nutrient types. This is in agreement with the growth rate hypothesis of stoichiometric theory, which postulates that rapid growth requires greater amounts of P-rich ribosomal RNA (Elser *et al.* 1996; Sterner & Elser 2002). Moreover, fatty acids were shown to be relatively more important for reproduction than for somatic growth, which also supports previous assertions that fatty acids are essential for hormone signaling involved in egg production in aquatic invertebrate consumers (Brett and Müller-Navarra 1997, Parrish 2009, and references therein). In other words, our results suggest that different nutrients can potentially be relatively more or less limiting throughout the life cycle of an individual herbivore, with P more likely to limit juvenile growth rates and EFAs more likely to constrain fecundity of mature adults than to limit juvenile growth. In nature, however, this will be highly contingent upon the specific nutritional status, and severe limitation by any individual nutrient can hinder multiple life stages. It is also worth noting that despite a much

stronger emphasis in the literature on P limitation effects, we also found N limitation to be significant for zooplankton somatic growth. This suggests that limitation (or co-limitation) by nitrogen, not just phosphorus, is a strong stoichiometric constraint on production of zooplankton that are facing low food N:C ratios, although this may become less frequent with anthropogenic increases of global N:P ratios (Peñuelas *et al.* 2020; Peñuelas & Sardans 2022).

### **Co-limitation and interactive effects of nutrients**

Our synthesis provides further evidence regarding the extent of co-limitation and interdependence of various dimensions of food quality (H2). Direct tests of co-limitation in this meta-analysis consistently point towards additive effects, whereby adding multiple distinct nutrient types elicits a stronger response than individual manipulations. Combined increases in N:C and P:C, EFAs and P:C, and EFAs and sterols all had larger effects on zooplankton growth than individual nutrient manipulations, which contradicts the idea that any single nutrient tends to be the main limiting factor for zooplankton. This is analogous to widely observed co-limitation of nitrogen and phosphorus for primary producers (Elser *et al.* 2007; Harpole *et al.* 2011). The evidence for co-limitation in zooplankton thus far remains limited, however, and future research on the effects of nutrient co-addition are required to fill this knowledge gap.

The interactive effect of both nutrient types is also evident from the observation that responses to EFA manipulations increased with higher P:C ratio, which quantitatively addresses the long-standing hypothesis that fatty acid limitation should be more important when phosphorus is replete (Sundbom & Vrede 1997; Sterner & Schulz 1998; Boersma 2000). These studies, which provide measurements of P:C along with EFA manipulations, illustrate the interdependence of elemental and biochemical nutrients quite well. For example, the threshold for P-limitation in *Daphnia* has been estimated to be a molar C:P ratio of ~150-250 (Anderson & Hessen 2005;

Khattak *et al.* 2018); above this ratio, P-limitation should dominate, and below it biochemical nutrients or other limiting factors should become more important. This corresponds to  $\log(P:C)$  ratios of ca. -5 to -5.5, which is approximately the break point at which predominance of phosphorus limitation versus EFA limitation appear to diverge (see Fig. 3). Specifically, EFA manipulation effects are both higher and more variable than those with P-limited food below this threshold. Also noteworthy is that even under severe P-limitation with C:P ratios near 2,000 ( $\log P:C = 7.6$ ), one study still found positive effects of EFA addition (Ferrão-Filho *et al.* 2003). This collectively points to the fact that measurements of multiple aspects of food quality, even when only one nutrient type is the core focus of a particular study, can yield important and novel insights into aquatic food web functioning.

### **Other factors influencing zooplankton nutritional constraints**

Several key factors contributed to the variation in the strength of nutrient manipulations (supporting H3), and may reconcile contrasting views on the nature of zooplankton nutrition. The most striking cause of heterogeneity for P-manipulation studies, for example, was the method in which P was manipulated. The direct method using short-term P-spiking immediately before feeding and thus only manipulating the algal P:C ratio had significantly smaller effects compared to the indirect method of manipulating dissolved P in the medium over a longer incubation of days to weeks. The indirect approach includes both direct effects of low P content as well as the complex suite of physiological changes resulting from algal P-stress. Notably, these indirect effects are strong for rotifers (Rothhaupt 1995; Zhou *et al.* 2018) as well as cladocerans, indicating that these effects are not constrained to a single taxon. These indirect effects include induced changes in cell size, cell wall thickness, and digestibility (van Donk & Hessen 1993; Lüring & Van Donk 1997; Lüring *et al.* 1997; Van Donk *et al.* 1997), but also changes in essential fatty acid content (Müller-Navarra 1995; Ahlgren *et al.* 1998; Ravet & Brett 2006;

Grzesiuk *et al.* 2018; Bi & Sommer 2020). Moreover, other food quality dimensions that have received less attention so far may be affected. The extent to which fatty acids like EPA are influenced by P limitation varies, however, and can either increase (Bi & Sommer 2020), not change (Park *et al.* 2002), or decline sharply (Müller-Navarra 1995) depending upon phytoplankton taxa. Therefore, we can only speculate as to the most important causes of these indirect effects. In sum, our results confirm that zooplankton can experience direct P limitation, that indirect effects of P-limitation often occur, but that more work is needed to identify the mechanisms by which P-stress induces these indirect effects on zooplankton production.

We also found large differences in the effects of nutrient manipulation for natural phytoplankton communities versus laboratory isolates. Here, we can also only make limited inferences as to the exact mechanisms causing these differences. The main hypotheses regarding food quality of phytoplankton involve a) mineral/biochemical content, b) toxicity, c) size/shaped-based grazing resistance, and d) other morphological defenses/digestion resistance (Gulati & DeMott 1997; Sterner & Schulz 1998). As most studies filtered natural seston to include only edible particles (< 35- $\mu$ m fraction), and there were no differences between filtered and unfiltered seston (Fig. 6), we can rule out size in this analysis. However, we cannot rule out any of the other three factors for explaining the large differences between lab and field algae. The relatively low effect sizes of nutrient addition for natural communities suggest three general possibilities: 1) the diverse natural communities were already of high quality and therefore did not increase much in quality with supplementation, 2) zooplankton were limited by algal defenses (e.g. indigestibility, toxicity, morphology) and not nutrients, and/or 3) zooplankton were limited by a mineral or biochemical nutrient not explicitly tested by the authors. Our findings suggest that natural communities may have been of poorer average quality within this dataset, as zooplankton had slightly lower (though still positive) growth rates on non-supplemented natural seston (mean =

0.22 d<sup>-1</sup>, 95% CI: 0.20-0.24) than on non-supplemented isolates (mean = 0.26 d<sup>-1</sup>, 95% CI: 0.22-0.30). Interestingly, the range in zooplankton somatic growth rates is smaller for natural communities than for isolates (Fig. S8), indicating that these more diverse communities may buffer the most severe effects of nutritional deficiency observed in monocultures. This would support previous findings that phytoplankton biodiversity can decrease the variance in zooplankton production (Striebel *et al.* 2012), possibly increasing the stability of the autotroph-consumer link in food webs. However, such effects clearly depend on the specific array of traits that come with increasing algal diversity (Marzetz *et al.* 2017), and the ability of the grazer to select for high-quality food. We included certain algal traits like digestion resistance as moderators, however these are only included to the extent that authors explicitly acknowledged such factors. The consequences of algal defenses for the observed effects of food nutrition will depend on their response to changes in nutrient availability. For example, if algal defenses increase with nutrient limitation, then the reported effects of nutrient content *per se* will be overestimated, as increased nutrient availability is correlated with reduced defenses. This is in fact seen in the slightly positive influence of induced digestion resistance, compared to the strong negative influence of constitutive digestion resistance, on values of Hedges' *g* (Fig. 6). The full extent to which such traits are linked to nutrient availability, and how they together affect zooplankton performance, remains to be elucidated (Meunier *et al.* 2017).

### **Nutrient limitation in natural systems**

In our analysis, we only included studies which controlled for any confounding effects of the biomass quantity available to zooplankton, which selects for mostly lab experiments, and predominantly those using common study organisms like *Daphnia* and *Scenedesmus* or *Synechococcus* spp. (see sample sizes by taxa in Fig. S4, S5). This largely reflects the biases in this field, where the focus on clear study systems has greatly supported our mechanistic insights

into zooplankton nutrient limitation. Nevertheless, that we have “a dearth of knowledge about other anomopods, cyclopoid and calanoid copepods, and rotifers” (Sterner & Schulz 1998) relative to *Daphnia* is as true today as it was in the 1990s.

Some of the limitations posed by the controlled laboratory experiments included in this synthesis do not apply for field and mesocosm studies. Therefore, a complementary meta-analysis, using field and mesocosm studies that tested effects of seston food quality (and quantity) on zooplankton production, could provide additional insights based on more field-realistic settings, if (somewhat) controlled for factors like species sorting and food quantity. The relative degree of nutrient limitation in nature is, of course, determined by the distribution of nutrient concentrations in aquatic systems, which varies considerably across sites (Brett *et al.* 2000; Elser *et al.* 2000). At large scales, concentrations of different nutrient types may be highly interactive, especially if P enrichment selects for phytoplankton communities of lower biochemical food quality (Müller-Navarra *et al.* 2004; Taipale *et al.* 2019). The mechanistic results from our meta-analysis will therefore pair well with systematic assessments of elemental and biochemical nutrient concentrations (e.g., field monitoring of nutrients over time and across different lakes). Specifically, our results illustrate the potential magnitude of nutrient limitation effects on zooplankton physiology with phytoplankton species composition ranging from monocultures to natural communities. The extent to which low quality taxa dominate natural systems, combined with the observed nutrient concentrations in a given system, should provide clear predictions regarding the extent of nutrient limitation in nature. However, particularly the combination of field observations across sites with controlled manipulations (e.g., Hartwich *et al.* (2012)), will allow making strong inferences regarding limitation of zooplankton production in nature.

## **Knowledge gaps**

We identify several recommendations for future research based on our systematic review and meta-analysis. As described above, the need is clear for more information on zooplankton nutrient co-limitation, simultaneous measurements of multiple food quality factors, and how these signals of algal nutritional quality are inherently connected to other functional traits of both producers and consumers. In addition to this, we find our database to be less data-rich for nutrient manipulation effects involving sterols,  $\omega$ 6-EFAs like ARA (Ilić *et al.* 2019), controlled manipulations of nutrients in natural phytoplankton communities, marine systems in general, and interactions between algal defenses (e.g., digestion resistance, silicification) and nutrient content. Therefore, we see clear gaps in our knowledge based on these lines of taxonomy, study system, level of biological complexity, and nutrients examined, which would benefit from additional future research. Moreover, our synthesis emphasizes that integration across disparate nutritional categories provides insights into food web dynamics. Our mechanistic understanding on zooplankton eco-physiology will therefore benefit from focusing on interactive effects of different nutrients on food web structure, as well as the linkages between functional diversity of zooplankton and phytoplankton, and environmental change.

## **Conclusion**

In conclusion, we find that every type of phytoplankton nutritional quality we considered in our meta-analysis significantly influenced growth and reproduction of herbivorous zooplankton. We show that several additional factors beyond low concentrations of nutrients *per se* can explain substantial variation in reported effects of algal food quality on planktonic consumers and therefore variation in the efficiency of aquatic food web functioning.

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697 Table 1. Components of food quality investigated in this meta-analysis. Bold denotes essential  
 698 fatty acids following Kainz et al. (2004). Units used are molar ratios (N:C, P:C) or mass ratios  
 699 ( $\mu\text{g mg C}^{-1}$  for fatty acids and sterols).

Abbreviation	Description
N:C	molar N:C ratio
P:C	molar P:C ratio
<b>LIN</b>	C18:2 $\omega$ 6; linoleic acid
GLA	C18:3 $\omega$ 6; $\gamma$ -linolenic acid
<b>ALA</b>	C18:3 $\omega$ 3; $\alpha$ -linolenic acid
SDA	C18:4 $\omega$ 3; stearidonic acid
<b>ARA</b>	C20:4 $\omega$ 6; arachidonic acid
<b>EPA</b>	C20:5 $\omega$ 3; eicosapentaenoic acid
<b>DHA</b>	C22:6 $\omega$ 3; docosahexaenoic acid
DHA:EPA	Ratio of DHA:EPA
$\omega$ 3	total omega-3 fatty acids
$\omega$ 6	total omega-6 fatty acids
$\omega$ 3: $\omega$ 6	$\omega$ 3: $\omega$ 6 ratio
PUFA	total polyunsaturated fatty acids
EFA	essential fatty acids (sum of LIN, ALA, ARA, EPA, DHA)
LCEFA	long-chain ( $\geq 20$ C) essential fatty acids (sum of ARA, EPA, DHA)
TFA	total fatty acids
sterol	total sterols

703 Table 2. Categories of nutrient manipulations used in the meta-analysis.

Nutrient manipulation category	Description	Illustrative studies using each method	Model used in	
			Hedges' g	Zr
Direct	Only the nutrient of interest is manipulated; all else held constant	Plath and Boersma (2001); Zhou et al. (2018) "LP+P" treatment compared to "LP"	×	×
Indirect (via nutrient supply)	Nutrient supply is manipulated; however, the manipulation may simultaneously alter other food quality factors	Sterner et al. (1993); Zhou et al. (2018) "HP" treatment compared to "LP"	×	×
Indirect (other)	Factor other than nutrient supply is manipulated (i.e., most commonly, species composition)	Müller-Navarra et al. (2000); Mandal et al. (2018)		×

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## Figure captions

Fig. 1. Effect sizes (mean and 95% CI) of nutrient manipulation on zooplankton growth and reproduction from a weighted random effects meta-analysis. Error bars that do not overlap zero indicate significant effects of increasing availability in zooplankton diets. Squares denote overall effects of N:C, P:C, essential fatty acids (EFA), and sterols, as well as their combined addition (below horizontal line); circles denote effects of specific fatty acids. The sample size for each combination of nutrient type and response is shown as the number of unique publications ( $n$ ) and the number of effect size estimates ( $k$ ).

Fig 2. Effect of nutrient gradients ( $Z$ -transformed correlation coefficient,  $Z_r$ ) for the subset of studies using indirect and/or gradient style manipulations to test effects of nutrient limitation. This dataset includes e.g., studies correlating zooplankton response by algal species, by different environmental conditions, or along nutrient manipulation gradients. Points show mean and 95% CI of effect sizes for each nutrient type, squares denote summary effect sizes, circles denote individual EFA correlations, and the sample size for each combination of nutrient type and response is shown as the number of unique publications ( $n$ ) and the number of effect size estimates ( $k$ ).

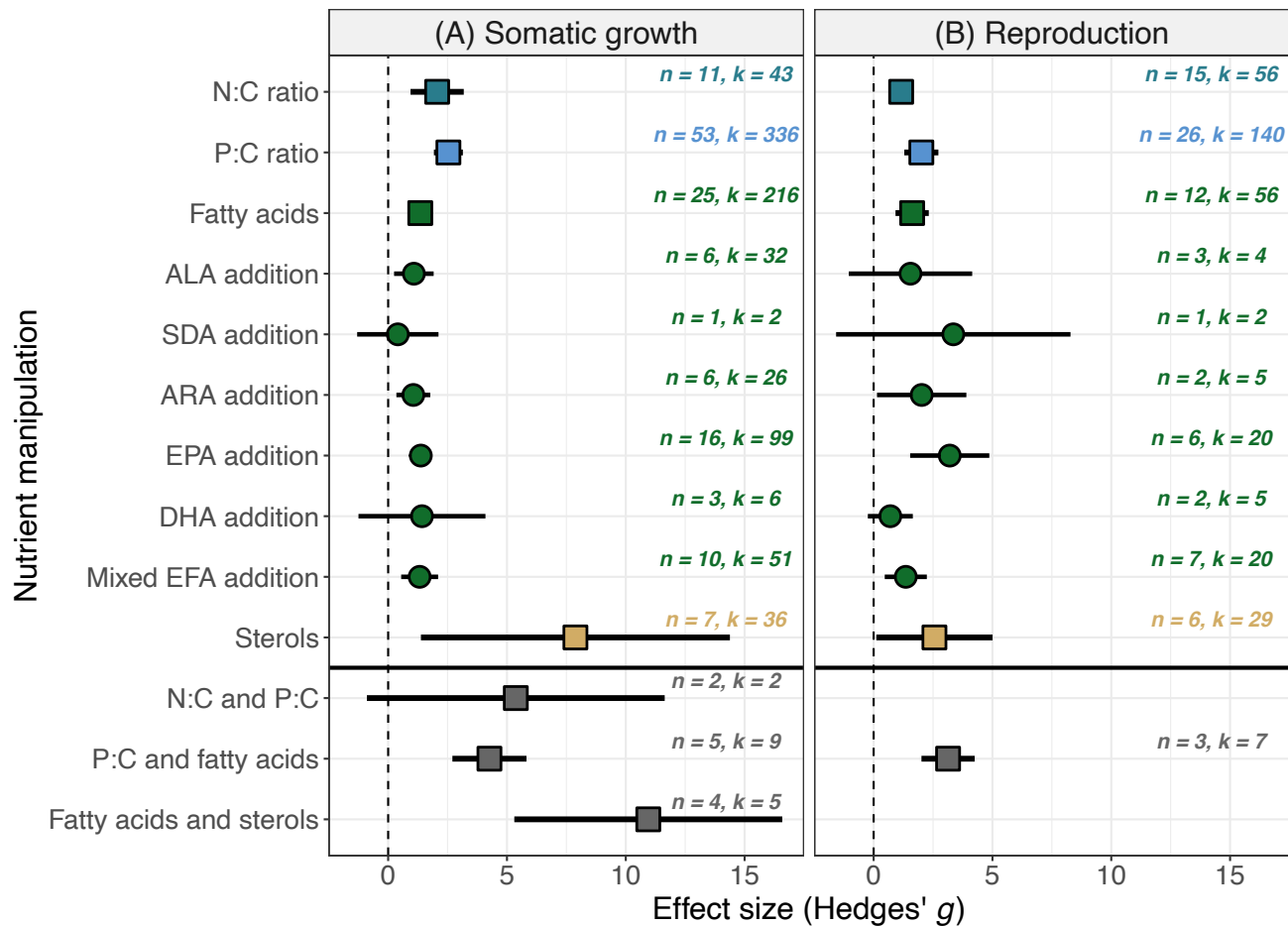
Fig 3. Relationship between variation in log-transformed P:C ratio and the effect size of EFA addition. Effect sizes of EFA manipulation generally increase with greater phosphorus availability, as shown by the 'loess' fit curve. The size of each point indicates the reliability of each effect size (i.e., the weighting in the form of inverse variance), and colors indicate different studies. For reference, log(P:C) values of -7, -6, -5, and -4 are equivalent to C:P ratios of ca. 1100, 400, 150, and 55, respectively.

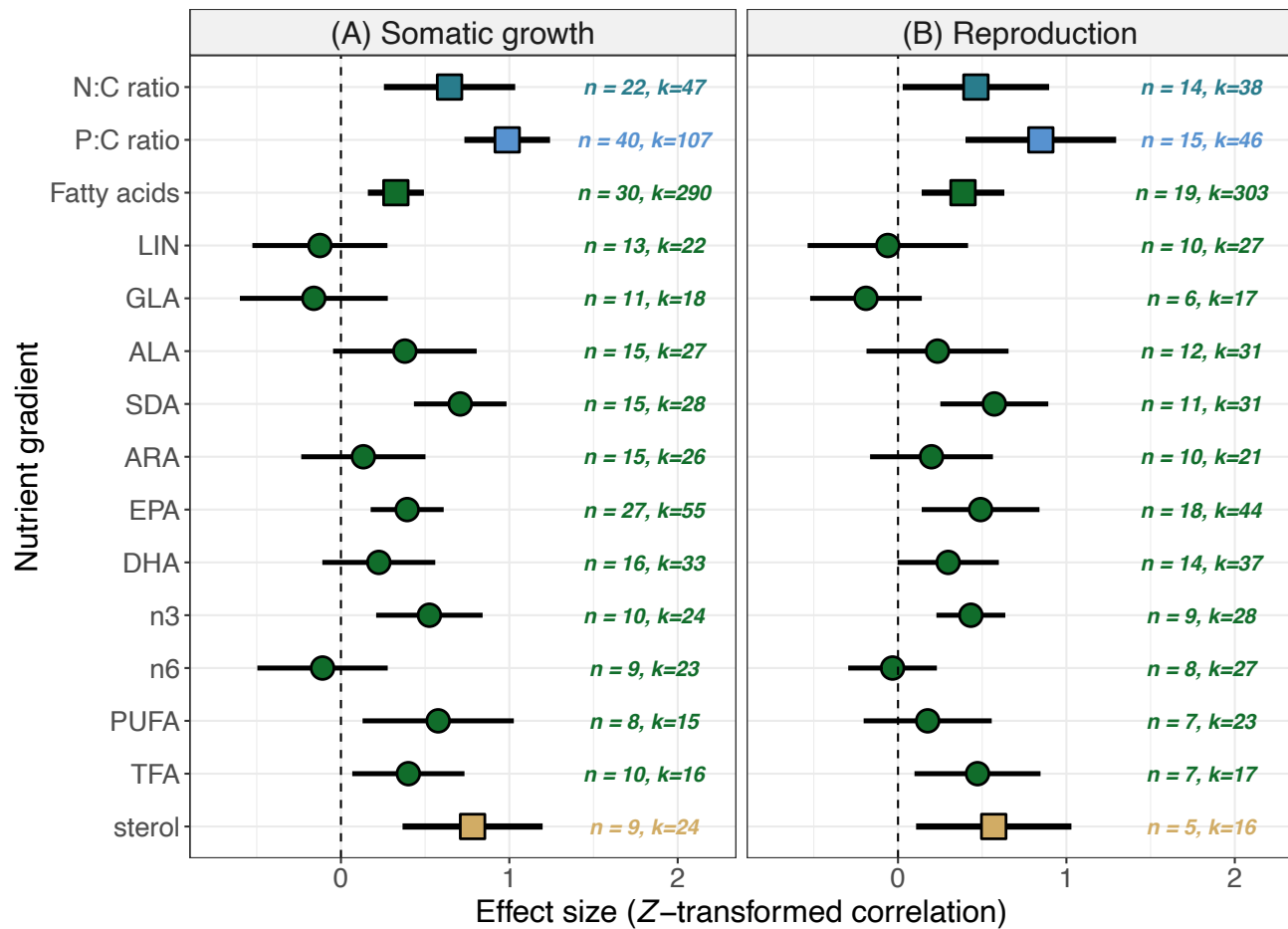
Fig. 4. The method by which P is manipulated has strong effects on P-limitation effect sizes.

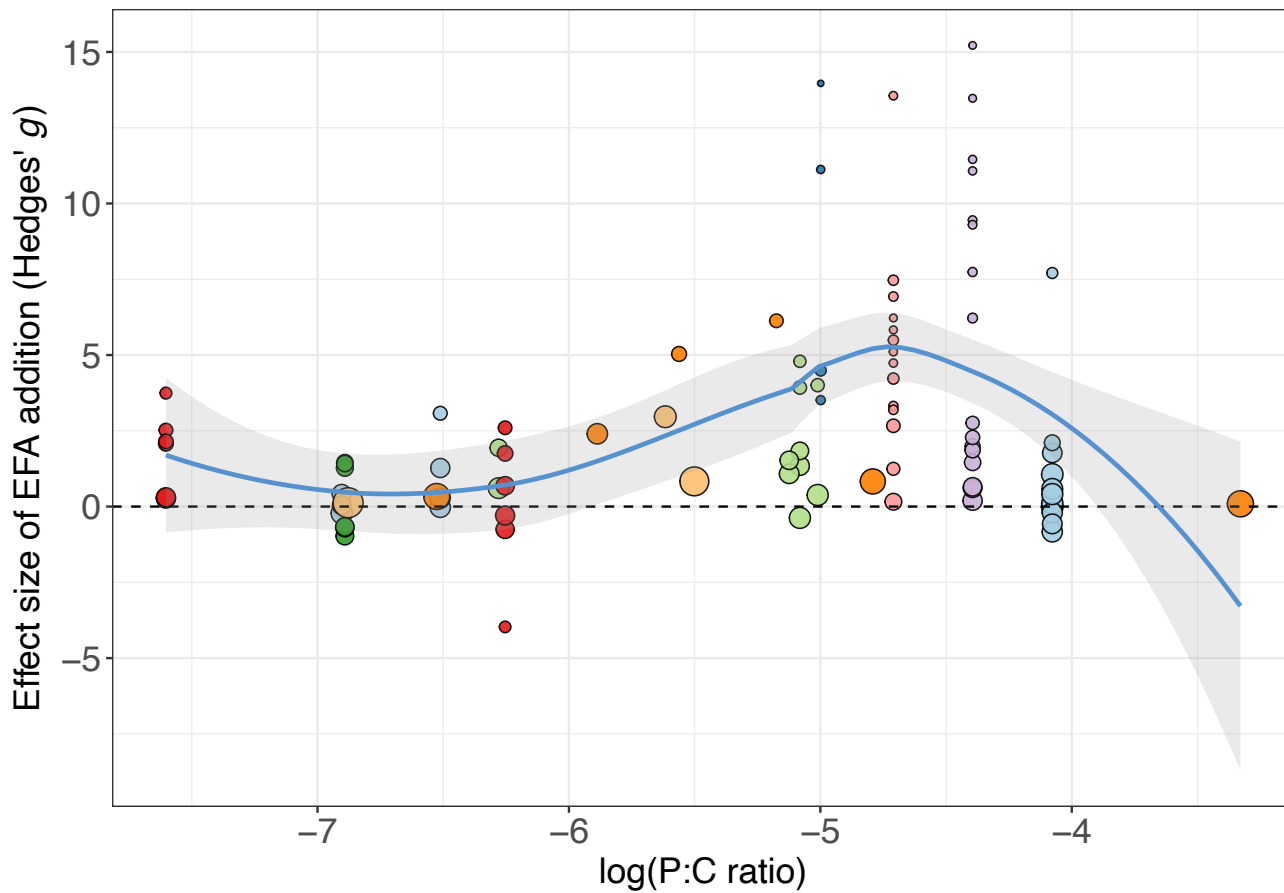
“Indirect” indicates that phytoplankton were exposed to differences in P supply during cultivation, which can induce indirect physiological effects of P-limitation (in addition to direct effects), whereas “Direct” indicates a more strictly direct test of elemental P-limitation, as these indirect effects are controlled for by spiking P such that phytoplankton physiology does not change before they are consumed, and only P:C varies between treatment and control. Values of  $n$  and  $k$  represent number of studies and number of effect size estimates, respectively.

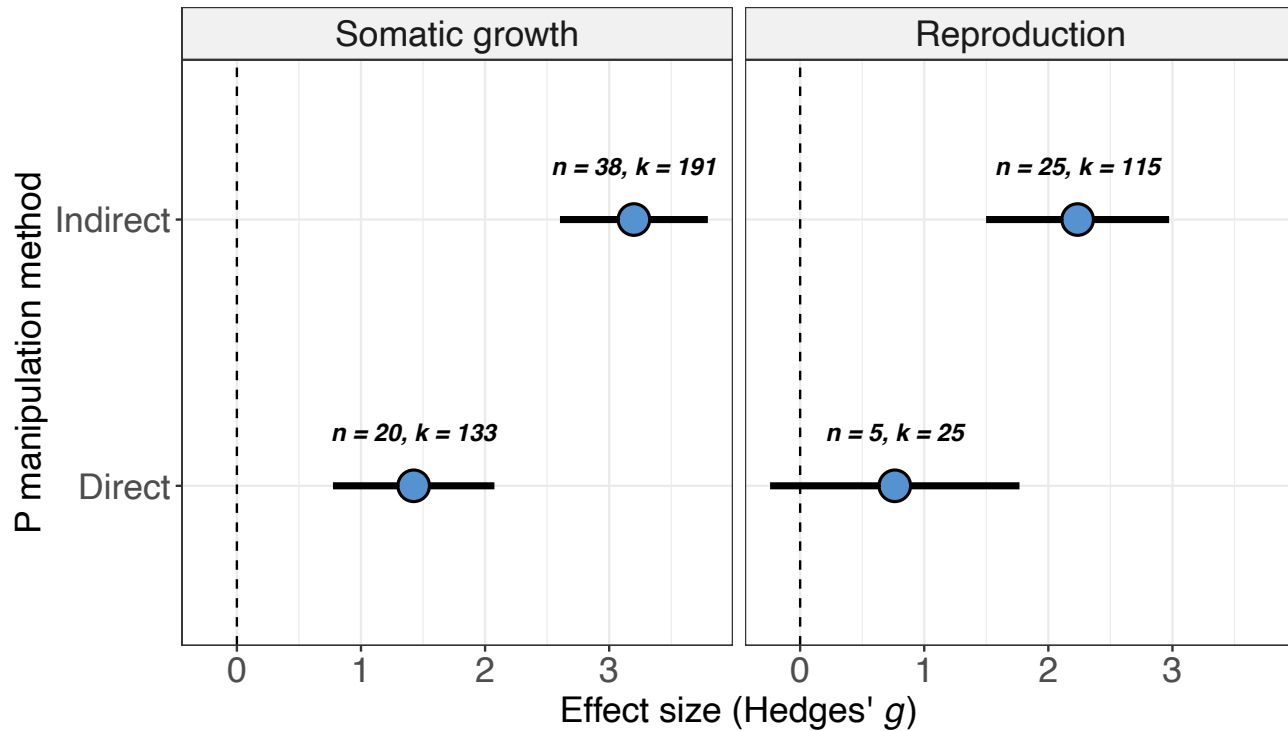
Fig. 5. Effect sizes for P:C and essential fatty acid manipulation, separated by zooplankton response category and whether the phytoplankton used were from a natural community or from isolated laboratory strains. Size of points corresponds to the number of studies per effect size estimate.

Fig. 6. Effects of additional experimental and morphological factors on nutrient limitation effects. Significant effects are indicated by bolded diamonds. The mean and 95% CI represent results of a mixed effects regression model.

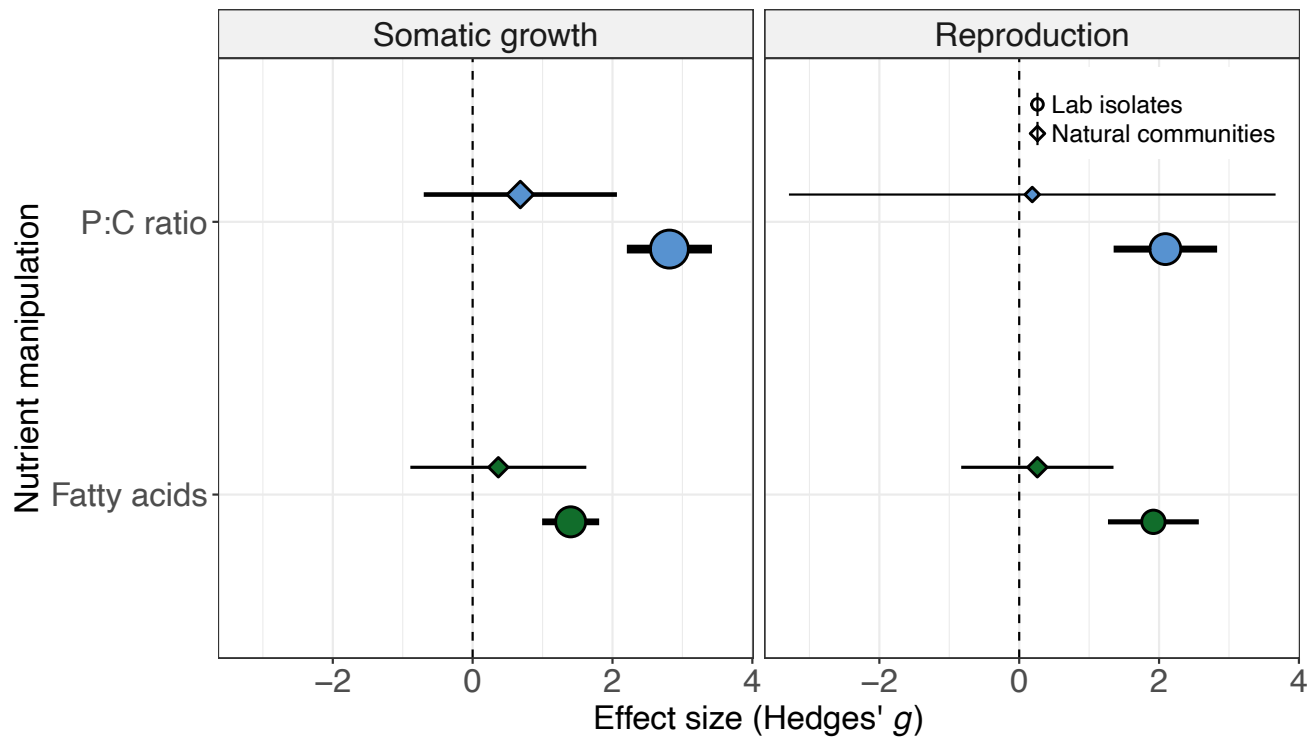












Other factor

