

Microbiome community composition and zoonotic bacterial pathogen prevalence in synanthropic *Peromyscus* mice

Janine Mistrick¹, Evan Kipp¹, Sarah Weinberg², Collin Adams¹, Peter Larsen¹, and Meggan Craft¹

¹University of Minnesota Twin Cities

²University of Pennsylvania

August 1, 2023

Abstract

Rodents are key reservoirs of zoonotic pathogens and play an important role in disease transmission to humans. Importantly, anthropogenic land-use change has been found to increase the abundance of synanthropic rodents, particularly rodent reservoirs of zoonotic disease. Anthropogenic environments also affect the microbiome of synanthropic wildlife, influencing wildlife health and potentially introducing novel pathogens. Our objective was to characterize the microbiome and investigate the prevalence of zoonotic bacterial pathogens in synanthropic rodents in native and anthropogenic environments to better understand their role in pathogen maintenance and transmission. We sampled wild *Peromyscus* mice in agricultural and undeveloped landscapes and forest and synanthropic habitat in Minnesota, USA and conducted 16S amplicon sequencing using long-read Nanopore sequencing technology on fecal samples to characterize the rodent microbiome. We compared community composition and diversity between habitats and screened for the presence of putative pathogenic bacteria species. Microbiome community composition differed significantly between agricultural and undeveloped landscapes and forest and synanthropic habitat while microbiome richness, diversity, and evenness were lower in undeveloped-forest habitat compared to all other habitats. We detected overall low abundance and diversity of putative pathogenic bacteria, though the greatest number of pathogenic bacteria were detected in the agricultural-forest habitat. Our findings show that rodent microbiome community composition differs across landscapes and habitat types but suggest that landscape-level anthropogenic factors may be most important to predict zoonotic pathogen abundance. Ultimately, understanding how anthropogenic land-use change and synanthropy affect rodent microbiomes and pathogen prevalence is important to managing transmission of rodent-borne zoonotic diseases to humans.

1 **Microbiome community composition and zoonotic bacterial pathogen prevalence in**
2 **synanthropic *Peromyscus* mice**

3 *Running title:* Microbiome composition of synanthropic mice

4

5 Janine Mistrick ¹, Evan J. Kipp ², Sarah I. Weinberg ³, Collin C. Adams ⁴, Peter A. Larsen ^{2†},
6 Meggan E. Craft ^{1†}

7 [†] co-senior authors

8

9 ¹ Department of Ecology, Evolution, and Behavior, College of Biological Sciences, University of
10 Minnesota, St. Paul, MN, USA

11 ² Department of Veterinary and Biomedical Sciences, College of Veterinary Medicine, University
12 of Minnesota, St. Paul, MN, USA

13 ³ School of Veterinary Medicine, University of Pennsylvania, Philadelphia, PA, USA

14 ⁴ Itasca Biological Station and Laboratories, University of Minnesota, Lake Itasca, MN, USA

15

16 Corresponding author: Janine Mistrick - mistr033@umn.edu

17 **ABSTRACT**

18 Rodents are key reservoirs of zoonotic pathogens and play an important role in disease
19 transmission to humans. Importantly, anthropogenic land-use change has been found to
20 increase the abundance of synanthropic rodents, particularly rodent reservoirs of zoonotic
21 disease. Anthropogenic environments also affect the microbiome of synanthropic wildlife,
22 influencing wildlife health and potentially introducing novel pathogens. Our objective was to
23 characterize the microbiome and investigate the prevalence of zoonotic bacterial pathogens in
24 synanthropic rodents in native and anthropogenic environments to better understand their role
25 in pathogen maintenance and transmission. We sampled wild *Peromyscus* mice in agricultural
26 and undeveloped landscapes and forest and synanthropic habitat in Minnesota, USA and
27 conducted 16S amplicon sequencing using long-read Nanopore sequencing technology on fecal
28 samples to characterize the rodent microbiome. We compared community composition and
29 diversity between habitats and screened for the presence of putative pathogenic bacteria
30 species. Microbiome community composition differed significantly between agricultural and
31 undeveloped landscapes and forest and synanthropic habitat while microbiome richness,
32 diversity, and evenness were lower in undeveloped-forest habitat compared to all other habitats.
33 We detected overall low abundance and diversity of putative pathogenic bacteria, though the
34 greatest number of pathogenic bacteria were detected in the agricultural-forest habitat. Our
35 findings show that rodent microbiome community composition differs across landscapes and
36 habitat types but suggest that landscape-level anthropogenic factors may be most important to
37 predict zoonotic pathogen abundance. Ultimately, understanding how anthropogenic land-use
38 change and synanthropy affect rodent microbiomes and pathogen prevalence is important to
39 managing transmission of rodent-borne zoonotic diseases to humans.

40

41 **KEYWORDS**

42 microbiome, Nanopore sequencing, *Peromyscus*, synanthropy, zoonoses, 16S amplicon
43 sequencing

44

45 **1 INTRODUCTION**

46 Rodents are an important source of zoonotic disease spillover, accounting for a greater diversity
47 of zoonotic pathogens than any other mammalian order (Han et al., 2016). While many factors
48 have been proposed to contribute to this (e.g. fast-paced life history, Han et al., 2015; cyclic
49 population fluctuations, Kallio et al., 2009) recent studies have suggested that the tendency of
50 particular rodent species to occasionally or exclusively live in human-built environments
51 (synanthropy) is likely a key factor (Ecke et al., 2022).

52

53 Anthropogenic land-use change, leading to habitat fragmentation and the intensification of
54 agricultural development and urbanization, is the major driver of zoonotic pathogen spillover
55 (Gottdenker et al., 2014). Indeed, urbanized habitat has been found to have a significant,
56 positive effect on the abundance of rodent hosts of zoonotic pathogens compared to areas of
57 native vegetation (Mendoza et al., 2019). Shifts in rodent biodiversity in anthropogenic
58 landscapes could further increase zoonotic risk, as rodent hosts and non-host rodents show
59 opposite responses to agricultural and urban habitat, with the abundance of host species
60 increasing and non-host species decreasing compared to areas of minimally disturbed primary
61 vegetation (Gibb et al., 2020).

62

63 However, spillover of zoonotic pathogens at the human-wildlife interface does not solely flow
64 from wildlife into humans. Synanthropic wildlife (including rodents) also show increased
65 prevalence of human pathogens: *Escherichia coli*, *Clostridioides difficile*, *Salmonella enterica* in
66 Norway rats in New York City, New York (Firth et al., 2014); antimicrobial-resistant *E.coli* in
67 racoons in Chicago, Illinois (Worsley-Tonks et al., 2021); *Salmonella* in urbanized white ibis in

68 southern Florida (Hernandez et al., 2016), representing both a concern for wildlife health and a
69 potential source for spillback into human populations. As such, while the relationships between
70 land-use change, rodents, and zoonotic pathogen prevalence are still being explored,
71 synanthropic wildlife represent both important reservoirs for zoonotic pathogens and likely
72 drivers of pathogen maintenance and spillover in anthropogenic landscapes (Hassell et al.,
73 2017).

74
75 Synanthropy has also been shown to impact the gut microbiome of wildlife. The gut microbiome
76 plays a role in host health (Marchesi et al., 2016) and immune function (Schluter et al., 2020)
77 and disruption of the normal microbiome has been linked to various health conditions in wildlife,
78 livestock, and domestic animals (Funosas et al., 2021; Monteiro & Faciola, 2020; Suchodolski,
79 2022). Wildlife living in close proximity to humans often experience changes to the composition
80 of their microbiome compared to counterparts in native habitat (e.g. rodents, Anders et al.,
81 2022; sparrows, Berlow et al., 2021) though whether anthropogenic habitats decrease or
82 increase microbiome diversity may vary by species (Diaz et al., 2023; Dillard et al., 2022). It is
83 likely that changes in microbiome diversity associated with synanthropy could increase the
84 prevalence of pathogenic bacteria in wildlife, but studies linking microbiome shifts with pathogen
85 prevalence are limited (but see Murray et al., 2020).

86
87 Here, we characterize the microbiome and compare the abundance of zoonotic bacterial
88 pathogens in *Peromyscus* mice in agricultural developed and undeveloped landscapes and
89 forest and synanthropic habitat in Minnesota, USA. Our research questions were two-fold: 1)
90 How does the microbiome community of *Peromyscus* mice differ between forest and
91 synanthropic habitat? and 2) Are zoonotic bacterial pathogens more abundant in agricultural
92 developed landscapes?

93

94 We expected the microbiome community of *Peromyscus* to be shaped by the surrounding
95 landscape and specific habitat as they influence the availability of food resources and exposure
96 to humans and their pathogens. We predicted that microbiome richness and diversity would be
97 lower in the agricultural landscape and synanthropic habitat compared to the undeveloped
98 landscape and forest habitat due lower diversity of food resources. We predicted that the
99 agricultural landscape would have a higher abundance and diversity of pathogenic bacteria
100 since the area is dominated by crop fields and human habitation and thus increased exposure to
101 manure as fertilizer, wastewater and runoff, and trash; whereas we predicted that the
102 undeveloped landscape would have lower pathogen abundance because the surrounding area
103 is largely forested with little anthropogenic development. Characterizing rodent microbiomes
104 across development gradients is important for quantifying the risk of rodent-borne zoonotic
105 pathogen spillover and understanding how microbiome shifts associated with synanthropy may
106 influence pathogen abundance.

107

108 **2 MATERIALS & METHODS**

109 **2.1 Study Sites**

110 Three major North American biomes intersect in Minnesota: the eastern deciduous forest,
111 northern coniferous forest, and western prairie, providing diverse habitats and biological
112 communities of hosts and pathogens. With respect to land-use, the state is dominated by
113 agricultural cropland and forest with interspersed developed areas ranging from dense
114 metropolitan areas to small, rural communities. Together, the biological and anthropogenic
115 factors create a heterogeneous landscape of natural areas mixed with agricultural and urban
116 developed landscapes where synanthropic rodents have many opportunities to overlap with
117 humans. We focus our study on mice of the genus *Peromyscus* (i.e. *Peromyscus leucopus* and
118 *Peromyscus maniculatis*) which are highly adaptable generalists that are common throughout
119 Minnesota and can thrive in agricultural and urban settings as well as forests and grasslands.

120 Importantly, *Peromyscus* mice are known reservoirs of zoonotic and foodborne pathogens
121 (e.g. *Borrelia*, *Campylobacter* spp., *E. coli*, *Giardia* spp., hantavirus; Jahan et al., 2021).

122
123 For our study, we focused on two landscape types: native, contiguous forest with little
124 permanent human habitation or agriculture (hereafter “undeveloped landscape”) and a mosaic
125 of fragmented forest interspersed with crop fields and low-density housing (hereafter
126 “agricultural landscape”). Within each landscape, four study sites were chosen to represent two
127 habitat types (two sites per habitat): forest habitat and synanthropic habitat around human-
128 frequented structures (e.g. cabins, tent platforms, field station buildings, maintenance sheds and
129 garages). Rodent sampling was conducted at two locations: the Itasca Biological Station and
130 Laboratories at Itasca State Park (“Itasca”, undeveloped landscape) and Cedar Creek
131 Ecosystem Science Reserve (“Cedar Creek”, agricultural landscape). Itasca is located in
132 northern Minnesota in the northern coniferous boreal forest biome. Though the state park is
133 frequented by hikers and visitors, the surrounding landscape is contiguous forest with no
134 agricultural development and very sparse permanent human habitation (Figure 1-A). Cedar
135 Creek is located in central Minnesota in the eastern deciduous forest and oak savanna biome.
136 The landscape surrounding the reserve is dominated by agricultural development (e.g.
137 pasture/hay, cultivated crops), woody and herbaceous wetlands, and low-medium intensity
138 housing communities (Figure 1-B).

139

140 **2.2 Rodent Trapping**

141 Two consecutive nights of rodent trapping were conducted at each study site (a “trapping
142 session”) using 100 Sherman live-capture traps baited with oats. Traps were opened at dusk
143 and checked at dawn the following morning. Traps were closed during the day between trap
144 nights at a single site and were reopened at dusk for the second night. Captured *Peromyscus*
145 mice were sampled and then released at the point of capture. Due to the difficulty in

146 distinguishing *P. leucopus* and *P. maniculatus* – two species found across our study landscapes
147 in Minnesota – based solely on morphologic features, we did not attempt to identify captured
148 *Peromyscus* mice to the species level. Captured non-target (i.e. non-*Peromyscus*) species were
149 released immediately and were not sampled. Longitudinal trapping was conducted at the
150 agricultural landscape sites. Each site was sampled three times throughout the summer (June,
151 July, and August 2019) with 3-4 weeks between trapping sessions. Captured *Peromyscus* mice
152 were marked with a metal ear tag to identify individuals at subsequent recaptures. Only one
153 trapping session (July 2019) was conducted at the undeveloped landscape sites. For each
154 captured *Peromyscus*, a fecal sample was collected and body mass, sex, and reproductive
155 status were recorded (reproductive individuals identified by the presence of scrotal testes for
156 males or any of the following traits for females: perforate vagina, enlarged nipples, palpable
157 embryos). Individuals captured a second time within a trapping session were not resampled and
158 were promptly released at the point of capture.

159

160 All rodent trapping and handling methods were reviewed and approved by the University of
161 Minnesota Institutional Animal Care and Use Committee (protocol no. 1903-36892A). The
162 objective of this study was live-capture and release but trap fatalities (3.4% [16/477] of capture
163 events of target and non-target species) were collected with approval by the Minnesota
164 Department of Natural Resources (MN-DNR) under Special Permit No. 28440 and were
165 accessioned with the Bell Museum of Natural History collections.

166

167 **2.3 DNA Extraction**

168 We collected 176 fecal samples representing 153 unique individuals. Fecal samples of up to
169 250 mg were stored without buffer or ethanol and frozen at -80°C immediately after sampling
170 until DNA extraction. DNA was extracted using a QIAamp PowerFecal Pro kit (Qiagen, Hilden,
171 Germany) following manufacturer instructions both manually and using a QIAcube robotic

172 workstation (Qiagen, Hilden, Germany). DNA extracts were quantified using a Qubit 4
173 fluorometer (ThermoFisher Scientific, Waltham, MA, USA) using the Qubit dsDNA BR Assay Kit
174 (ThermoFisher Scientific, Waltham, MA, USA) following the manufacturer's instructions. Samples
175 with low DNA yield (<24 ng/ μ L, $n=16$) were excluded from downstream analysis.

176

177 **2.4 Library Prep & Nanopore Sequencing**

178 For the remaining 160 samples, the Rapid 16S Barcoding Kit (SQK-16S024 [utilizing 'Kit 9'
179 chemistry]; Oxford Nanopore Technologies [ONT], Oxford, UK) was used to prepare barcoded
180 amplicon libraries for sequencing, largely following the manufacturer's protocol (methods
181 described in detail in Jahan et al., 2021). First, all fecal DNA extracts were diluted in nuclease-
182 free water to a concentration of 10-30 ng/ μ L. The full-length bacterial 16S rRNA gene region
183 (1.6kb) was amplified via PCR using specific primers and between 20-40 ng of DNA template, a
184 long-range master mix (LongAmp Hot Start Taq, 2X; New England Biolabs, Ipswich, MA, USA),
185 and sample-specific barcode identifier. PCR products were purified and prepared for
186 sequencing through a series of magnetic bead wash steps (AMPure XP beads; Beckman
187 Coulter Life Sciences, Indianapolis, IN, USA). Barcoded samples were pooled with ONT rapid
188 sequencing adapter mixture into a final library for sequencing. Seven sequencing runs were
189 performed with a total of 160 samples, including 24 (run 1, 2, 4), 23 (run 6), 22 (run 3, 5), and 21
190 (run 7) barcoded samples from individual mice. Libraries were sequenced on a FLO-MIN106
191 MinION flow cell utilizing R9 sequencing chemistry (Oxford Nanopore Technologies, Oxford,
192 UK), run for 24 hours using the ONT MinKNOW GUI (v4.3.20; Oxford Nanopore Technologies,
193 Oxford, UK).

194

195 **2.5 Bioinformatic Pipeline**

196 Raw Fast5 data from the sequencing runs were base-called using the ONT Guppy basecaller
197 using the 'super accuracy' basecalling model (ONT configuration file:

198 dna_r9.4.1_450bps_sup.cfg). The barcoded samples were further de-multiplexed using the
199 Guppy barcoder to identify reads as belonging to one of the 24 unique barcodes. Reads were
200 quality filtered (Nanopore Q score ≥ 8 , corresponds to 84.15% base call accuracy) and filtered
201 for target length (full-length bacterial 16S region approx. 1600 bp in length) using NanoFilt
202 (v2.8.0; De Coster et al., 2018). Only reads 1200-1800 bp in length were retained for onward
203 analysis. Summary reports were generated using Nanoq (v0.9.0; Steinig & Coin, 2022).

204

205 Taxonomic abundance profiles were generated using Emu, an expectation-maximization
206 algorithm designed specifically to account for the increased read length and error rate often
207 associated with long-read data such as ONT-generated sequences (v3.4.4; Curry et al., 2022).
208 Compared to conventional taxonomic identification algorithms, Emu is able to reduce the false
209 positive rate of identification and accurately identify long reads to species level (Curry et al.,
210 2022). Reads were mapped using the Emu default database settings: a combination of rrnDB
211 v5.6 (Stoddard et al., 2015) and NCBI (National Center for Biotechnology Information) 16S
212 RefSeq downloaded on 17 September 2020 (O'Leary et al., 2016). The output of Emu is an
213 estimated abundance (read count) of each identified species in a given sample. Because read
214 counts are estimated based on likelihood probabilities, outputted values are not necessarily
215 integer counts. Values were rounded to the nearest integer for analysis.

216

217 **2.6 Data Analysis**

218 The full fecal microbiome was characterized at the sample level using measures of alpha and
219 beta diversity (to quantify within-sample and between sample bacterial diversity, respectively).
220 Alpha diversity indices included species richness, Shannon-Weiner diversity, Simpson diversity,
221 and species evenness. Shannon diversity and Simpson diversity make different assumptions
222 about species evenness and how it contributes to diversity: Shannon diversity assumes all
223 species are present and are randomly sampled while Simpson diversity gives more weight to

224 common species. Calculating both indices can suggest how common or rare species may affect
225 diversity estimates for different populations. Beta diversity was quantified using the Bray-Curtis
226 dissimilarity index to compare bacterial microbiome community composition at the species level
227 between all pairs of samples. As a subset of the full fecal microbiome, the presence of
228 pathogenic bacteria (foodborne and zoonotic pathogens of concern for human, livestock, and
229 domestic animal health) was quantified at the sample level, then grouped by landscape-habitat
230 pairing.

231

232 Rodent sampling was conducted across three months (June, July, and August) in the
233 agricultural landscape and 18 individuals were captured and sampled in multiple months. To
234 control for non-independence between repeat samples of the same individuals, only one sample
235 per mouse (n=140) was included in the alpha and beta diversity analyses. We chose to include
236 only the July sample for all recaptured mice to avoid introducing variation based on sampling
237 month (all recaptured animals were sampled in July, but not in June or August) and to better
238 align with the undeveloped landscape sampling (which was only conducted in July). For the
239 analysis of pathogenic bacteria species, all samples (n=160) were used.

240

241 For the analyses of alpha and beta diversity, all samples were rarefied to the number of reads of
242 the least abundant sample using the 'phyloseq' R package (v1.38.0; McMurdie & Holmes,
243 2013). Alpha diversity indices (richness, Shannon, Simpson, and evenness) were estimated
244 from the rarefied data using the 'vegan' R package (v2.6.4; Oksanen et al., 2022). We
245 investigated whether alpha diversity was affected by landscape or habitat type by developing a
246 linear regression model for species richness and Shannon diversity and a beta regression
247 model for Simpson diversity and species evenness. In all models, the response variable was the
248 alpha diversity index and the explanatory variables were landscape type (i.e. anthropogenic or
249 undeveloped), habitat type (forest or synanthropic), the interaction of landscape and habitat

250 type, mouse sex, reproductive status (reproductive or non-reproductive), body mass, and
251 sampling month (June, July, or August). Beta diversity was visualized using non-metric
252 multidimensional scaling (NMDS) ordination performed on the rarefied data using the Bray-
253 Curtis dissimilarity index in the 'vegan' package. NMDS was first performed with 2-dimensions
254 (k) and the k value was iteratively increased until the stress value was below 0.1. Non-
255 parametric statistical analyses were performed on the rarefied distance matrices using the
256 'adonis2', 'anosim', 'betadisper', and 'permutest' functions also in the 'vegan' package.

257

258 For the analysis of pathogenic bacteria, species-level abundances were not rarefied and the raw
259 estimated read counts output by the Emu pipeline were used. A list of 209 putative pathogenic
260 bacteria species was generated using the PHI-base pathogen database (Urban et al., 2020;
261 accessed on 13 Feb. 2023, plant-specific pathogens removed); resources from the U.S. Centers
262 for Disease Control and Prevention on 'foodborne germs and illnesses' (CDC, 2022); and
263 foodborne and mastitis-causing pathogens screened for by Jahan *et al.* 2021 (Jahan et al.,
264 2021; For full list of pathogens, see Table S1). The species-level read count abundance data
265 from the sequenced samples was filtered for reads assigned to the pathogen species on this
266 list. We thresholded read count per pathogen species to at least 50 reads and visualized the
267 patterns of pathogen read count per mouse, grouped by landscape-habitat pairing.

268

269 All analyses were performed in R Statistical Software (v4.1.2; R Core Team, 2021).

270

271 **3 RESULTS**

272 **3.1 Rodent Samples**

273 160 fecal samples were sequenced, representing 140 unique *Peromyscus* mice. In the
274 agricultural landscape, 50 and 39 total fecal samples were collected from forest and
275 synanthropic habitats, respectively, across three months of rodent trapping (Figure 1-C).

276 Excluding recaptures, 40 and 29 unique mouse fecal samples were collected in forest and
277 synanthropic habitats, respectively. In the undeveloped landscape, 31 and 40 unique mouse
278 fecal samples were collected from forest and synanthropic habitats, respectively (Figure 1-D).

279

280 **3.2 Nanopore Sequencing Summary**

281 After quality filtering, over 32.7 million high quality reads were retained (mean Q score $12.8 \pm$
282 0.31 s.d; Q score of 12.8 corresponds to base call accuracy of 94.75%). The mean number of
283 reads per sample was 204,772.4, though the number of reads per sample was highly variable
284 (standard deviation: 82,970.5; range: 74,517-517,058 reads; Table 1).

285

286 The Emu algorithm identified 1212 unique bacterial species across the 160 fecal samples. The
287 mean number of species per sample was 211 (standard deviation: ± 55.8 ; range: 82-367).

288

289 Rarefaction curves were plotted for all sequenced samples ($n=160$). The asymptotic nature of
290 these curves suggest reasonable sequencing depth was achieved for all samples (Figure S1).
291 To enable direct comparisons between samples for the alpha and beta diversity analyses, the
292 samples used in the diversity analysis ($n=140$) were rarefied to the minimum number of reads
293 per sample (74,517 reads) and species were selected without replacement to reach the desired
294 number of reads. After rarefaction, 36 species were removed because they were no longer
295 present in any sample after random subsampling.

296

297 **3.3 Alpha Diversity**

298 The interaction of landscape and habitat type had a moderate effect on observed species
299 richness, Shannon diversity and Simpson diversity indices (all $p < 0.05$; Table S2). The effect of
300 the interaction of landscape and habitat type on species evenness was weaker and only
301 marginally significant ($p=0.087$; Table S2), though there was a significant effect of landscape

302 alone ($p=0.016$; Table S2). Mean observed species richness, Shannon diversity, Simpson
303 diversity, and species evenness were lower in the undeveloped-forest habitat compared to all
304 other landscape-habitat pairings (Figure 2; Table S3). However, contrary to our hypotheses,
305 there was no difference in species richness, diversity, or evenness between forest and
306 synanthropic habitats in the agricultural landscape or between agricultural-synanthropic and
307 undeveloped-synanthropic habitat. Reproductive status (reproductive or non-reproductive
308 individual, as noted by external morphology) had a moderate effect on Shannon and Simpson
309 diversity and species evenness (all $p<0.01$; Table S2). None of the other parameters tested
310 (sex, body mass, sampling month) had an effect on any alpha diversity index.

311

312 **3.4 Beta Diversity**

313 Across the four landscape-habitat pairings, the microbiome communities of sampled mice were
314 dominated by three phyla: Firmicutes, Proteobacteria, and Bacteroidetes (relative abundance
315 $\geq 5\%$) though Melainabacteria (a candidate phylum related to Cyanobacteria, Di Rienzi et al.,
316 2013) and Deferribacteres were observed at relative abundances $\geq 1\%$ in some samples (Figure
317 3). Firmicutes was the dominant phyla in most samples (relative abundance 90.1% mean ± 11.1
318 s.d.) followed by Proteobacteria ($16.8\% \pm 20.0$) and Bacteroidetes ($8.92\% \pm 3.07$).

319

320 Bacterial microbiome community composition at the species level was compared between all
321 pairs of samples using the Bray-Curtis dissimilarity index based on rarefied count data. A
322 nonparametric analysis of similarities test ('anosim' function, 'vegan' R package) comparing
323 dissimilarity indices between samples from the four landscape-habitat pairings suggested that
324 the between-group dissimilarity in microbiome community composition was significantly greater
325 than the within-group dissimilarity ($p=0.001$).

326

327 An NMDS ordination plot calculated based on Bray-Curtis dissimilarity indices showed a high
328 degree of overlap between samples from the four landscape-habitat pairings (Figure 4).
329 Samples from agricultural-synanthropic and undeveloped-forest habitat showed the greatest
330 dissimilarity while samples from agricultural-forest and undeveloped-synanthropic habitat were
331 more similar. The variability among samples was high, but an analysis of multivariate
332 homogeneity of group dispersion ('betadisper' and 'permutest' functions, 'vegan' R package) by
333 landscape-habitat pairing showed no significant difference in variance between the groups
334 (permutation test, $p=0.96$), indicating that the differences in community composition were not
335 only due to differences in sample variance.

336

337 A nonparametric PERMANOVA analysis was used to test the effects of landscape, habitat type,
338 mouse sex, reproductive status, body mass, and sampling month on differences in microbiome
339 community composition using the 'adonis2' function in the 'vegan' R package with the
340 by="margin" option to determine the marginal effect of each parameter. There was a small but
341 significant effect of landscape and habitat, suggesting that the microbiome of sampled mice was
342 different between agricultural and undeveloped landscapes and between forest and
343 synanthropic habitats (PERMANOVA, $R^2_{\text{Landscape}}=0.06$, $R^2_{\text{Habitat}}=0.04$, both $p=0.001$; Table S4).
344 Mouse reproductive status and body mass also had small, but significant effects (both $p<0.05$).
345 However, much of the variance in microbiome community composition was not explained by the
346 modeled parameters (residual $R^2=0.85$).

347

348 **3.5 Putative Pathogen Detection**

349 The presence of putative pathogenic bacteria was investigated using raw read counts of all
350 sequenced samples ($n=160$). Read counts from mice captured in more than one month in the
351 agricultural landscape were pooled by bacterial species across fecal samples from a single
352 mouse. Of the 209 putative pathogenic bacteria species screened for, 18 were identified in

353 sampled mice (read count ≥ 50). At the population level, putative pathogen species richness
354 was higher in agricultural-forest and undeveloped-synanthropic habitat (13 species identified;
355 Figure 5) compared to agricultural-synanthropic and undeveloped-forest habitat (7 species
356 identified). However, at the individual level, putative pathogen species richness was higher in
357 mice in the agricultural landscape (agricultural-forest: mean putative pathogen species/mouse
358 1.42 ± 1.17 s.d.; agricultural-synanthropic: 1.24 ± 1.06) compared to mice in the undeveloped
359 landscape (undeveloped-forest: 0.42 ± 0.77 ; undeveloped-synanthropic: 0.83 ± 0.93).

360

361 Read counts of detected putative pathogens were similar across landscape-habitat pairings with
362 many mice having low read counts (<200 reads), though the number of mice with high read
363 counts (>500 reads) was greatest in the agricultural-forest habitat (Figure 5). Across all sampled
364 mice, *Clostridioides difficile*, *Streptococcus sanguinis*, *Enterococcus gallinarum*, *Citrobacter*
365 *freundii*, and *Morganella morganii* were the most frequently detected putative pathogens (Figure
366 5).

367

368 **4 DISCUSSION**

369 Our objective was to characterize and compare the microbiome of synanthropic rodents and the
370 abundance of zoonotic bacterial pathogens in agricultural landscapes and synanthropic habitat
371 in Minnesota. We found that landscape-habitat pairing affected microbiome richness and
372 diversity but species evenness was only affected by landscape. Overall, undeveloped-forest
373 habitat had lower mean alpha diversity (richness, Shannon and Simpson diversity, evenness)
374 than the other three landscape-habitat pairings. Microbiome community composition at the
375 species level was also significantly different between landscapes (agricultural versus
376 undeveloped) and habitat types (forest vs. synanthropic). We detected reads for a number of
377 putative pathogenic bacteria across the four habitats, mostly at low read counts. The mean

378 number of putative pathogenic bacteria detected per mouse was higher in the agricultural
379 landscape than the undeveloped.

380

381 Across landscape-habitat pairings, the microbiome of sampled mice was dominated by three
382 phyla (Firmicutes, Bacteroidetes, Proteobacteria). These phyla are typical of the gut microbiome
383 of wild *Peromyscus*, though we observed higher levels of Firmicutes and lower levels of
384 Bacteroidetes compared to previous studies (e.g. Diaz et al., 2023; Schmidt et al., 2019). This
385 suggests that the core fecal microbiome of the mice in our study is similar to *Peromyscus*
386 *maniculatus* in other regions of North America. Only one other study has compared microbiome
387 communities of free living *Peromyscus* in developed and undeveloped habitats (Diaz et al.,
388 2023). We found lower richness and alpha diversity in the undeveloped-forest habitat compared
389 to all other habitats, conversely, Diaz *et al.* found lower mean richness and Shannon diversity in
390 urban habitats compared to undeveloped habitats. However, the directionality of alpha diversity
391 shifts between undeveloped and developed populations is likely affected by multiple species-
392 and system-specific factors; research in other wildlife systems has documented an increase in
393 alpha diversity between undeveloped and developed populations (Dillard et al., 2022). Despite
394 the differences in the direction of alpha diversity shifts, our finding that the microbiome
395 community composition (beta diversity) between mice from undeveloped and agricultural
396 developed landscapes was significantly different aligned with the findings of Diaz *et al.* These
397 shifts in microbiome composition could be attributed to dietary shifts based on habitat type and
398 food availability, particularly in synanthropic environments (Anders et al., 2022). In future
399 studies, stable isotope analysis similar to those conducted by Anders *et al.* could provide
400 additional insights into the diet of synanthropic and forest mice. Such information would likely
401 inform the microbiome composition observed in our data, as the PERMANOVA modeling
402 approach utilized herein indicated a high degree of unexplained microbiome composition
403 variability that was not accounted for by landscape or habitat type.

404

405 We detected 16S sequences of a number of putative pathogenic bacteria in samples from all
406 four landscape-habitat pairings. The greatest number of mice carrying putative pathogenic
407 bacteria and the highest mean diversity of putative pathogen species per mouse was found in
408 the anthropogenic-forest habitat while the lowest was found in the undeveloped-forest habitat.
409 These differences are likely explained by the landscape surrounding our sampling locations
410 which could represent a source of infection for many of these pathogens. The forest sampling
411 sites in the agricultural landscape were located on the periphery of a research reserve which is
412 surrounded by crop fields, pastures, and low-density housing. By contrast, the forest sites in the
413 undeveloped landscape were contained in a state park and the forest continues uninterrupted
414 beyond the park boundary with little agricultural development, limiting sources of pathogen
415 exposure. *Peromyscus* are known to forage in crop fields as well as forest habitat, so it is likely
416 that the abundance of putative pathogens in mice in the anthropogenic-forest habitat are
417 representative of exposure to the surrounding agricultural landscape. Indeed, *Clostridioides*
418 *difficile* was the most frequently detected putative pathogenic bacteria in the agricultural
419 landscape, aligning with literature documenting this bacteria in many species of livestock and
420 wildlife, including antimicrobial resistant strains in urban rodents and those living on or near
421 farms (reviewed in Weese, 2020). In agricultural settings, manure used as fertilizer may serve
422 as a source of environmental contamination for *C. difficile* (Frentrup et al., 2021) which could
423 provide a transmission route to rodents and other wildlife. Contrary to our predictions, the mean
424 number of putative pathogenic bacteria per mouse was similar between forest and synanthropic
425 habitat within a landscape, suggesting similar levels of pathogen exposure for mice between
426 these two habitats. The synanthropic habitats sampled were all at the interface of forest and
427 human-habitated areas. It is possible the synanthropic mice only occasionally visit the human
428 structures where they were trapped (maintenance garages and storage areas, cabins and tent
429 platforms, etc.) and predominantly reside in the nearby forest. Frequent movement of mice

430 between native vegetation and synanthropic habitat could account for similar putative pathogen
431 exposure within a landscape type.

432

433 Accurate detection and taxonomic assignment of reads is a key assumption for community
434 diversity and metagenomic analyses. Species richness and diversity estimates can be sensitive
435 to the presence of rare species. The Emu algorithm has a built-in abundance threshold of 10
436 reads for large samples (over 1,000 reads) to control against long tails of low-abundance
437 species which are an artifact of the probabilistic expectation-maximization model (Curry et al.,
438 2022). As a result, Emu has a limited ability to detect rare species and thus our estimates of
439 species richness and diversity are likely underestimations of the true community composition.
440 However, Emu's strength is that it was specifically designed for taxonomic identification of long-
441 read sequence data. The Emu pipeline helps to correct errors and improve the accuracy of
442 Nanopore 16S amplicon sequencing through the expectation-maximization algorithm and has
443 been shown to outperform algorithms designed for short-read (i.e. Illumina) data when
444 classifying 16S Nanopore sequences (Curry et al., 2022). Because we were most interested in
445 the species-level identification of reads for the detection of putative pathogenic bacteria, we
446 chose to prioritize accurate taxonomic assignment over the ability to detect rare species and
447 more accurately estimate species richness and diversity. Furthermore, Nanopore sequencing
448 provides a key advancement over short-read microbiome sequencing in that species-level
449 identification is possible and accurate. In future research, we see great utility for taxonomic
450 assignment algorithms like Emu designed specifically for long-read Nanopore sequences and
451 expect these novel methods to continue to improve the ability to accurately characterize and
452 study species-level microbiome composition. Indeed, already the Nanopore 'Kit 12' chemistry
453 and R10 flow cells (released in late 2021) are able to outperform Illumina sequencing with less
454 noise and higher accuracy, specifically for species-level classification of 16S amplicon
455 sequencing of gut microbiota (Szoboszlay et al., 2023).

456

457 It is important to clarify that, while we can be confident in accurate taxonomic assignment of the
458 bacterial species detected in the sampled mice, their presence does not guarantee zoonotic
459 potential. Many of these bacteria are commensal in the human and mammalian gut and may
460 only be opportunistic pathogens or only certain serotypes possess virulence factors capable of
461 infecting humans. Determining pathogenicity requires more in-depth genotyping or lab cultures
462 that were outside the scope of this research. Nonetheless, our detection of these bacteria
463 species serves to inform the potential of *Peromyscus* mice to be reservoirs for zoonotic
464 pathogens and can inform future studies that characterize the pathogenicity of these bacteria.

465

466 Our research supports and expands upon previous work done in Minnesota using Nanopore
467 sequencing to identify pathogenic bacteria in synanthropic rodents. Jahan *et al.* pointed to the
468 role that farms play in the increased abundance of putative pathogenic bacteria in synanthropic
469 rodents (Jahan et al., 2021). However, farms are a unique anthropogenic environment with
470 many routes of pathogen introduction, and rodents at this interface may not be representative of
471 synanthropic rodents more broadly. Our work expands upon the foundation set by Jahan *et al.*
472 by investigating less disturbed environments to understand the abundance and diversity of
473 zoonotic bacterial pathogens in undeveloped and agricultural (cropland) landscapes. The
474 diversity of putative pathogenic genera found in *Peromyscus* mice generally align between our
475 studies: Jahan *et al.* similarly identified putative pathogenic genera including *Bacillus*,
476 *Clostridium*, *Enterococcus*, and *Streptococcus* circulating in synanthropic rodents on Minnesota
477 farms. However, we identified a higher abundance of *Clostridioides* and no pathogenic species
478 of *Helicobacter* in our study. It is possible that these differences can be attributed to differences
479 in how pathogen abundance was quantified: Jahan *et al.* reported abundance of reads identified
480 at the genus level (summed across all sampled *Peromyscus*) whereas we focused on read
481 abundance of specific pathogenic species per individual mouse. Interestingly, Jahan *et al.* found

482 lower abundance of putative pathogenic genera in *Peromyscus* mice compared to other rodent
483 species trapped on farms including *Mus musculus*, *Microtus pennsylvanicus*, and *Rattus*
484 *norvegicus*. While our study did not include other rodent species, the limited abundance of
485 putative pathogenic bacteria found in *Peromyscus* herein corroborates the findings of Jahan *et*
486 *al.* and could indicate lower exposure for these mice compared to other synanthropic rodents.

487

488 Overall, we found that *Peromyscus* in undeveloped and agricultural landscapes in Minnesota
489 carried low abundance and diversity of putative pathogenic bacteria (we detected, on average,
490 1-2 putative pathogens per mouse and zero putative pathogens in many mice). Further, many of
491 these were opportunistic pathogens which may pose a limited risk to zoonotic transmission in
492 the human population. Our findings suggest that agricultural landscapes play a role in
493 increasing the abundance of zoonotic pathogens in wild rodents; however, synanthropic habitat
494 may be less informative of the abundance of zoonotic bacterial pathogens, particularly in
495 environments where mice are expected to be highly mobile across interfaces between native
496 vegetation and synanthropic areas. Taken together, our research suggests that *Peromyscus* are
497 occasional hosts of zoonotic bacterial pathogens when sources of exposure are high (i.e.
498 agricultural settings like crop fields and farms) but their flexibility to thrive in natural vegetation
499 as well synanthropic habitat may act as a buffer to higher levels of zoonotic pathogen
500 abundance.

501

502 **5 CONCLUSIONS**

503 The data presented herein provide a glimpse into the gut microbiome of *Peromyscus* mice in
504 diverse landscapes of Minnesota. By sampling from populations in agricultural and undeveloped
505 landscapes and in forest and synanthropic habitat, we find that landscape and habitat are
506 important factors influencing microbiome community composition in wild rodents. We also
507 identify low abundance of putative pathogenic bacteria species in these populations and

508 suggest the role of agricultural landscapes in increasing rodent exposure to putative pathogens.
509 Even where transmission risk seems low, infection in wildlife populations could represent
510 sources of novel pathogenic strains, bridge hosts linking environmental contamination back to
511 human or livestock infection, or vectors to translocate pathogens across the landscape. As
512 such, this research underscores the importance of investigating zoonotic pathogen prevalence
513 in synanthropic rodents and other wildlife to better characterize their potential as reservoirs and
514 vectors for pathogen spillover at the human-wildlife interface.

515

516 **ACKNOWLEDGEMENTS**

517 We thank Dr. Forest Isbell, Dr. Caitlin Barale Potter, and the staff of the Cedar Creek
518 Ecosystem Science Reserve; Dr. Jonathan Schilling and the staff at Itasca Biological Station
519 and Laboratories and Itasca State Park for their coordination and support of this research.
520 Rodent trapping at Cedar Creek was conducted under Research Project Proposal RP922,
521 Experiment No. E331. Rodent trapping at Itasca State Park was conducted under approval of
522 the MN-DNR Division of Parks and Trails Special Permit No. 201938. We thank Dr. Sharon
523 Jansa for providing rodent sampling permits and field supplies and Dr. Laramie Lindsey, Lauren
524 Agnew, and Abby Guthmann for assisting with field sampling. Suzanne Stone provided logistical
525 assistance with the molecular lab work. Dr. Pat Schloss, Dr. Noelle Noyes, and Jasmine Veitch
526 provided guidance on data analysis. The Minnesota Supercomputing Institute provided
527 computational and data storage resources. Research funding was provided by: the Itasca
528 Biological Station American Indian Fund student research internship; Alexander and Lydia
529 Anderson Fellowship through the Graduate School at the University of Minnesota (UMN); the
530 Zoological Society Fund & the Dayton Bell Museum Fund of the Bell Museum of Natural History;
531 the Cedar Creek Graduate Fellows program; the Florence Rothman Research Fellowship
532 through the Department of Ecology, Evolution, and Behavior at UMN; and the Graduate Student
533 Grant-in-Aid of Research from the American Society of Mammalogists. JM was supported by

534 the National Science Foundation Graduate Research Fellowship Program under Award No.
535 2237827. Any opinions, findings, and conclusions or recommendations expressed in this
536 material are those of the authors and do not necessarily reflect the views of the National
537 Science Foundation. EJK was supported by UMN startup funds awarded to PAL and SIW was
538 supported by the Van Sloun Foundation.

539

540 REFERENCES

- 541 Anders, J. L., Mychajliw, A. M., Moustafa, M. A. M., Mohamed, W. M. A., Hayakawa, T., Nakao,
542 R., & Koizumi, I. (2022). Dietary niche breadth influences the effects of urbanization on
543 the gut microbiota of sympatric rodents. *Ecology and Evolution*, *12*(9), e9216.
544 <https://doi.org/10.1002/ece3.9216>
- 545 Berlow, M., Phillips, J. N., & Derryberry, E. P. (2021). Effects of urbanization and landscape on
546 gut microbiomes in white-crowned sparrows. *Microbial Ecology*, *81*(1), 253–266.
547 <https://doi.org/10.1007/s00248-020-01569-8>
- 548 CDC. (2022, December 19). *Foodborne Illnesses and Germs*. Centers for Disease Control and
549 Prevention. <https://www.cdc.gov/foodsafety/foodborne-germs.html>
- 550 Curry, K. D., Wang, Q., Nute, M. G., Tyshaieva, A., Reeves, E., Soriano, S., Wu, Q., Graeber,
551 E., Finzer, P., Mendling, W., Savidge, T., Villapol, S., Dilthey, A., & Treangen, T. J.
552 (2022). Emu: Species-level microbial community profiling of full-length 16S rRNA Oxford
553 Nanopore sequencing data. *Nature Methods*, *19*(7), Article 7.
554 <https://doi.org/10.1038/s41592-022-01520-4>
- 555 De Coster, W., D'Hert, S., Schultz, D. T., Cruets, M., & Van Broeckhoven, C. (2018). NanoPack:
556 Visualizing and processing long-read sequencing data. *Bioinformatics*, *34*(15), 2666–
557 2669. <https://doi.org/10.1093/bioinformatics/bty149>
- 558 Dewitz, J. (2021). *National Land Cover Database (NLCD) 2019 Products* [Data set]. U.S.
559 Geological Survey. <https://doi.org/10.5066/P9KZCM54>
- 560 Di Rienzi, S. C., Sharon, I., Wrighton, K. C., Koren, O., Hug, L. A., Thomas, B. C., Goodrich, J.
561 K., Bell, J. T., Spector, T. D., Banfield, J. F., & Ley, R. E. (2013). The human gut and
562 groundwater harbor non-photosynthetic bacteria belonging to a new candidate phylum
563 sibling to Cyanobacteria. *ELife*, *2*, e01102. <https://doi.org/10.7554/eLife.01102>
- 564 Diaz, J., Redford, K. H., & Reese, A. T. (2023). Captive and urban environments are associated
565 with distinct gut microbiota in deer mice (*Peromyscus maniculatus*). *Biology Letters*,
566 *19*(3), 20220547. <https://doi.org/10.1098/rsbl.2022.0547>
- 567 Dillard, B. A., Chung, A. K., Gunderson, A. R., Campbell-Staton, S. C., & Moeller, A. H. (2022).
568 Humanization of wildlife gut microbiota in urban environments. *ELife*, *11*, e76381.
569 <https://doi.org/10.7554/eLife.76381>
- 570 Ecke, F., Han, B. A., Hörnfeldt, B., Khalil, H., Magnusson, M., Singh, N. J., & Ostfeld, R. S.
571 (2022). Population fluctuations and synanthropy explain transmission risk in rodent-
572 borne zoonoses. *Nature Communications*, *13*(1), Article 1.
573 <https://doi.org/10.1038/s41467-022-35273-7>
- 574 Firth, C., Bhat, M., Firth, M. A., Williams, S. H., Frye, M. J., Simmonds, P., Conte, J. M., Ng, J.,
575 Garcia, J., Bhuvu, N. P., Lee, B., Che, X., Quan, P.-L., & Lipkin, W. I. (2014). Detection
576 of zoonotic pathogens and characterization of novel viruses carried by commensal

577 Rattus norvegicus in New York City. *MBio*, 5(5), e01933-14.
578 <https://doi.org/10.1128/mBio.01933-14>

579 Frentrup, M., Thiel, N., Junker, V., Behrens, W., Münch, S., Siller, P., Kabelitz, T., Faust, M.,
580 Indra, A., Baumgartner, S., Schepanski, K., Amon, T., Roesler, U., Funk, R., & Nübel, U.
581 (2021). Agricultural fertilization with poultry manure results in persistent environmental
582 contamination with the pathogen *Clostridioides difficile*. *Environmental Microbiology*,
583 23(12), 7591–7602. <https://doi.org/10.1111/1462-2920.15601>

584 Funosas, G., Triadó-Margarit, X., Castro, F., Villafuerte, R., Delibes-Mateos, M., Rouco, C., &
585 Casamayor, E. O. (2021). Individual fate and gut microbiome composition in the
586 European wild rabbit (*Oryctolagus cuniculus*). *Scientific Reports*, 11(1), Article 1.
587 <https://doi.org/10.1038/s41598-020-80782-4>

588 Gibb, R., Redding, D. W., Chin, K. Q., Donnelly, C. A., Blackburn, T. M., Newbold, T., & Jones,
589 K. E. (2020). Zoonotic host diversity increases in human-dominated ecosystems. *Nature*,
590 584(7821), Article 7821. <https://doi.org/10.1038/s41586-020-2562-8>

591 Gottdenker, N. L., Streicker, D. G., Faust, C. L., & Carroll, C. R. (2014). Anthropogenic land use
592 change and infectious diseases: A review of the evidence. *EcoHealth*, 11(4), 619–632.
593 <https://doi.org/10.1007/s10393-014-0941-z>

594 Han, B. A., Kramer, A. M., & Drake, J. M. (2016). Global patterns of zoonotic disease in
595 mammals. *Trends in Parasitology*, 32(7), 565–577.
596 <https://doi.org/10.1016/j.pt.2016.04.007>

597 Han, B. A., Schmidt, J. P., Bowden, S. E., & Drake, J. M. (2015). Rodent reservoirs of future
598 zoonotic diseases. *Proceedings of the National Academy of Sciences*, 112(22), 7039–
599 7044. <https://doi.org/10.1073/pnas.1501598112>

600 Hassell, J. M., Begon, M., Ward, M. J., & Fèvre, E. M. (2017). Urbanization and disease
601 emergence: Dynamics at the wildlife–livestock–human interface. *Trends in Ecology &
602 Evolution*, 32(1), 55–67. <https://doi.org/10.1016/j.tree.2016.09.012>

603 Hernandez, S. M., Welch, C. N., Peters, V. E., Lipp, E. K., Curry, S., Yabsley, M. J., Sanchez,
604 S., Presotto, A., Gerner-Smidt, P., Hise, K. B., Hammond, E., Kistler, W. M., Madden,
605 M., Conway, A. L., Kwan, T., & Maurer, J. J. (2016). Urbanized white ibises (*Eudocimus
606 albus*) as carriers of *Salmonella enterica* of significance to public health and wildlife.
607 *PLOS ONE*, 11(10), e0164402. <https://doi.org/10.1371/journal.pone.0164402>

608 Jahan, N. A., Lindsey, L. L., Kipp, E. J., Reinschmidt, A., Heins, B. J., Runck, A. M., & Larsen,
609 P. A. (2021). Nanopore-based surveillance of zoonotic bacterial pathogens in farm-
610 dwelling peridomestic rodents. *Pathogens*, 10(9), Article 9.
611 <https://doi.org/10.3390/pathogens10091183>

612 Kallio, E. R., Begon, M., Henttonen, H., Koskela, E., Mappes, T., Vaheri, A., & Vapalahti, O.
613 (2009). Cyclic hantavirus epidemics in humans—Predicted by rodent host dynamics.
614 *Epidemics*, 1(2), 101–107. <https://doi.org/10.1016/j.epidem.2009.03.002>

615 Marchesi, J. R., Adams, D. H., Fava, F., Hermes, G. D. A., Hirschfield, G. M., Hold, G.,
616 Quraishi, M. N., Kinross, J., Smidt, H., Tuohy, K. M., Thomas, L. V., Zoetendal, E. G., &
617 Hart, A. (2016). The gut microbiota and host health: A new clinical frontier. *Gut*, 65(2),
618 330–339. <https://doi.org/10.1136/gutjnl-2015-309990>

619 McMurdie, P. J., & Holmes, S. (2013). phyloseq: An R package for reproducible interactive
620 analysis and graphics of microbiome census data. *PLOS ONE*, 8(4), e61217.
621 <https://doi.org/10.1371/journal.pone.0061217>

622 Mendoza, H., Rubio, A. V., García-Peña, G. E., Suzán, G., & Simonetti, J. A. (2019). Does land-
623 use change increase the abundance of zoonotic reservoirs? Rodents say yes. *European
624 Journal of Wildlife Research*, 66(1), 6. <https://doi.org/10.1007/s10344-019-1344-9>

625 Monteiro, H. F., & Faciola, A. P. (2020). Ruminal acidosis, bacterial changes, and
626 lipopolysaccharides. *Journal of Animal Science*, 98(8), skaa248.
627 <https://doi.org/10.1093/jas/skaa248>

628 Murray, M. H., Lankau, E. W., Kidd, A. D., Welch, C. N., Ellison, T., Adams, H. C., Lipp, E. K., &
629 Hernandez, S. M. (2020). Gut microbiome shifts with urbanization and potentially
630 facilitates a zoonotic pathogen in a wading bird. *PLOS ONE*, *15*(3), e0220926.
631 <https://doi.org/10.1371/journal.pone.0220926>

632 Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P. R.,
633 O'Hara, R. B., Simpson, G. L., Solymos, P., & Stevens, M. H. H. (2022). *Vegan:*
634 *Community Ecology Package. R package version 2.5-2.7. 2020.*

635 O'Leary, N. A., Wright, M. W., Brister, J. R., Ciufo, S., Haddad, D., McVeigh, R., Rajput, B.,
636 Robbertse, B., Smith-White, B., Ako-Adjei, D., Astashyn, A., Badretdin, A., Bao, Y.,
637 Blinkova, O., Brover, V., Chetvernin, V., Choi, J., Cox, E., Ermolaeva, O., ... Pruitt, K. D.
638 (2016). Reference sequence (RefSeq) database at NCBI: Current status, taxonomic
639 expansion, and functional annotation. *Nucleic Acids Research*, *44*(D1), D733–D745.
640 <https://doi.org/10.1093/nar/gkv1189>

641 R Core Team. (2021). *R: A language and environment for statistical computing.* R Foundation
642 for Statistical Computing. Vienna, Austria. URL <https://www.R-project.org/>

643 Schluter, J., Peled, J. U., Taylor, B. P., Markey, K. A., Smith, M., Taur, Y., Niehus, R., Staffas,
644 A., Dai, A., Fontana, E., Amoretti, L. A., Wright, R. J., Morjaria, S., Fenelus, M., Pessin,
645 M. S., Chao, N. J., Lew, M., Bohannon, L., Bush, A., ... Xavier, J. B. (2020). The gut
646 microbiota is associated with immune cell dynamics in humans. *Nature*, *588*(7837), 303–
647 307. <https://doi.org/10.1038/s41586-020-2971-8>

648 Schmidt, E., Mykytczuk, N., & Schulte-Hostedde, A. I. (2019). Effects of the captive and wild
649 environment on diversity of the gut microbiome of deer mice (*Peromyscus maniculatus*).
650 *The ISME Journal*, *13*(5), Article 5. <https://doi.org/10.1038/s41396-019-0345-8>

651 Steinig, E., & Coin, L. (2022). Nanoq: Ultra-fast quality control for nanopore reads. *Journal of*
652 *Open Source Software*, *7*(69), 2991. <https://doi.org/10.21105/joss.02991>

653 Stoddard, S. F., Smith, B. J., Hein, R., Roller, B. R. K., & Schmidt, T. M. (2015). rrnDB:
654 Improved tools for interpreting rRNA gene abundance in bacteria and archaea and a
655 new foundation for future development. *Nucleic Acids Research*, *43*(D1), D593–D598.
656 <https://doi.org/10.1093/nar/gku1201>

657 Suchodolski, J. S. (2022). Analysis of the gut microbiome in dogs and cats. *Veterinary Clinical*
658 *Pathology*, *50*(S1), 6–17. <https://doi.org/10.1111/vcp.13031>

659 Szoboszlay, M., Schramm, L., Pinzauti, D., Scerri, J., Sandionigi, A., & Biazzo, M. (2023).
660 Nanopore is preferable over Illumina for 16S amplicon sequencing of the gut microbiota
661 when species-level taxonomic classification, accurate estimation of richness, or focus on
662 rare taxa is required. *Microorganisms*, *11*(3), Article 3.
663 <https://doi.org/10.3390/microorganisms11030804>

664 Urban, M., Cuzick, A., Seager, J., Wood, V., Rutherford, K., Venkatesh, S. Y., De Silva, N.,
665 Martinez, M. C., Pedro, H., Yates, A. D., Hassani-Pak, K., & Hammond-Kosack, K. E.
666 (2020). PHI-base: The pathogen–host interactions database. *Nucleic Acids Research*,
667 *48*(D1), D613–D620. <https://doi.org/10.1093/nar/gkz904>

668 Weese, J. S. (2020). Clostridium (*Clostridioides*) difficile in animals. *Journal of Veterinary*
669 *Diagnostic Investigation : Official Publication of the American Association of Veterinary*
670 *Laboratory Diagnosticians, Inc*, *32*(2), 213–221.
671 <https://doi.org/10.1177/1040638719899081>

672 Worsley-Tonks, K. E. L., Miller, E. A., Anchor, C. L., Bender, J. B., Gehrt, S. D., McKenzie, S.
673 C., Singer, R. S., Johnson, T. J., & Craft, M. E. (2021). Importance of anthropogenic
674 sources at shaping the antimicrobial resistance profile of a peri-urban mesocarnivore.
675 *Science of The Total Environment*, *764*, 144166.
676 <https://doi.org/10.1016/j.scitotenv.2020.144166>

677

678 **DATA ACCESSIBILITY STATEMENT**

679 Upon manuscript submission, all Nanopore sequence data will be uploaded to the National Center
680 for Biotechnology Information Sequence Read Archive. Metadata associated with all samples will
681 be made available on Dryad. The R code used for analysis will be archived on Zenodo.

682

683 **BENEFIT-SHARING STATEMENT**

684 This research was supported by the Itasca Biological Station American Indian Fund whose goal
685 is to foster scientific growth and collaboration between local students from area high schools
686 and the researchers at the Itasca Biological Station and Laboratories. A recent high school
687 graduate (CCA) collaborated on the field research and is included as a co-author. As described
688 above, all data resulting from this research have been publicly shared via appropriate research
689 databases.

690

691 **AUTHOR CONTRIBUTIONS**

692 JM, MEC, and PAL conceived of and designed the research and acquired funding. JM, SIW,
693 and CCA performed the field sampling. JM and EJK conducted sequencing. JM conducted the
694 bioinformatic and statistical analyses with support from EJK and PAL. JM led the writing of the
695 manuscript under the mentorship of MEC and PAL. All authors contributed to revisions and
696 editing. All authors have read and agreed to the published version of the manuscript.

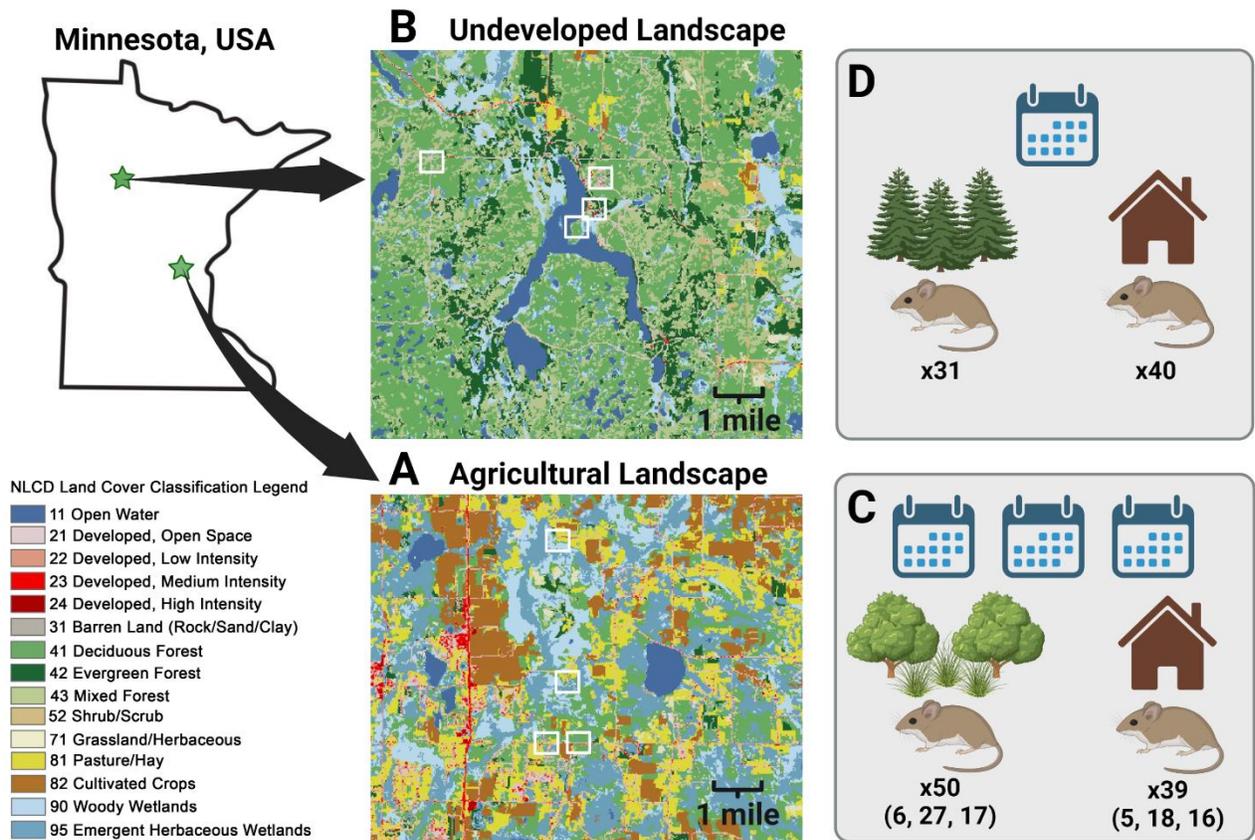
697

698 **CONFLICTS OF INTEREST**

699 The authors declare no conflict of interest.

700

701

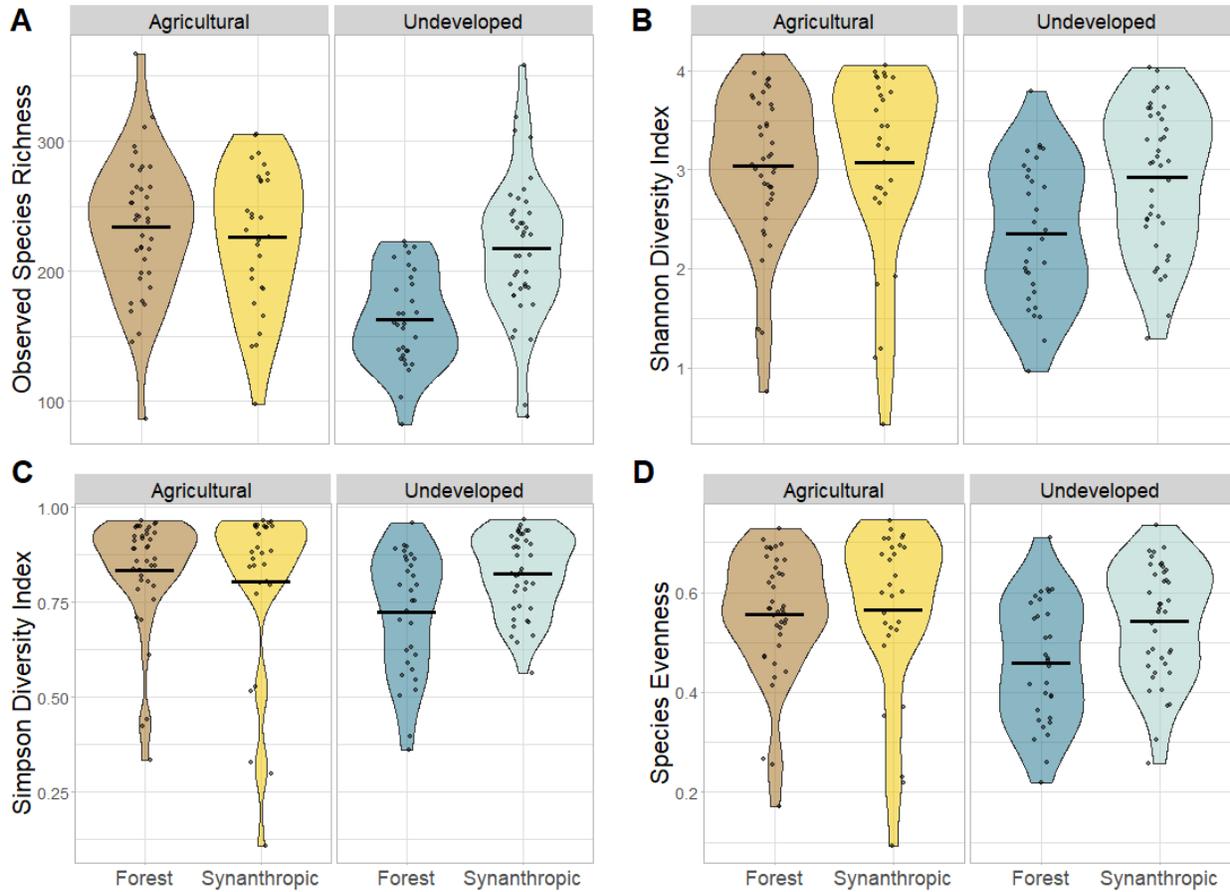


703
 704 Figure 1. Rodent sampling locations and sample size summary. Sampling was conducted at two
 705 locations in Minnesota, USA representing undeveloped and agricultural landscapes. Study sites
 706 are outlined with white boxes (A, B). Sample size (total number of fecal samples) in forest and
 707 synanthropic habitat is shown for each landscape (C, D). Sampling was conducted once in the
 708 undeveloped landscape and three times in the agricultural landscape. Total number of samples
 709 per landscape-habitat pairing is noted first with samples per month in parentheses below
 710 (includes multiple samples from individual mice). Maps and land cover classification legend from
 711 National Land Class Database (NLCD) 2019 (Dewitz, 2021). Figure created with
 712 BioRender.com.
 713

714 Table 1. Summary statistics of 16S Nanopore sequencing data of mouse fecal sample DNA
 715 (after quality filtering) by landscape, habitat type, and sampling month. Mean and standard
 716 deviation are reported for number (N) of reads per sample (reported in units of thousands of
 717 reads), number of basepairs per sample (reported in units of millions of basepairs [Mb]), and
 718 read quality (Q) score. Individual sampling months in the agricultural landscape shown in italics,
 719 rows shaded in gray. Mean values across all three months shown in bold. Number of samples
 720 represents total number of fecal samples sequenced (includes repeat sampling of unique mice).
 721

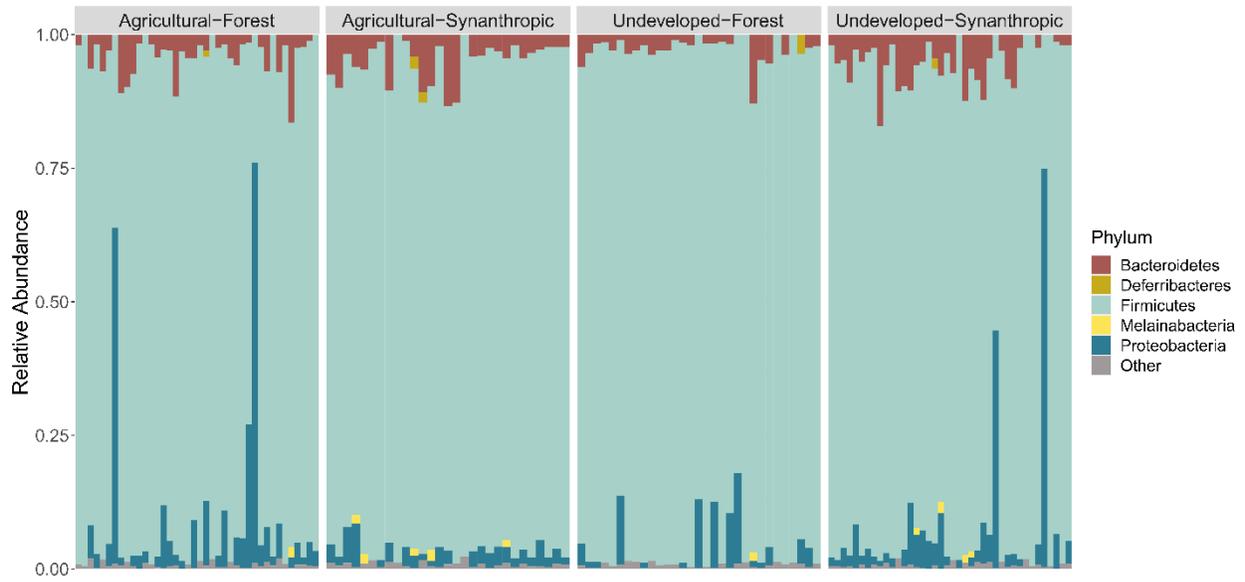
Landscape	Habitat Type	Month	N samples	N reads/sample (thousands of reads)	N basepairs/sample (Mb)	Q Score
<i>Agricultural</i>	<i>Forest</i>	<i>June</i>	6	272.18 ± 39.12	433.27 ± 61.68	13.08 ± 0.04
<i>Agricultural</i>	<i>Forest</i>	<i>July</i>	27	262.35 ± 62.91	418.69 ± 100.39	13.14 ± 0.06
<i>Agricultural</i>	<i>Forest</i>	<i>August</i>	17	248.61 ± 116.91	395.76 ± 186.39	12.84 ± 0.31
Agricultural	Forest	Summer	50	258.86 ± 82.36	412.64 ± 131.34	13.03 ± 0.23
<i>Agricultural</i>	<i>Synanthropic</i>	<i>June</i>	5	326.35 ± 73.84	517.32 ± 115.67	13.08 ± 0.08
<i>Agricultural</i>	<i>Synanthropic</i>	<i>July</i>	18	215.45 ± 21.66	342.53 ± 33.57	12.82 ± 0.04
<i>Agricultural</i>	<i>Synanthropic</i>	<i>August</i>	16	88.02 ± 10.85	139.68 ± 17.33	12.4 ± 0
Agricultural	Synanthropic	Summer	39	177.39 ± 88.31	281.72 ± 139.93	12.68 ± 0.25
Undeveloped	Forest	July	31	139.49 ± 22.44	223.04 ± 35.6	12.43 ± 0.22
Undeveloped	Synanthropic	July	40	214.45 ± 59.77	342.01 ± 95.9	12.88 ± 0.11

722



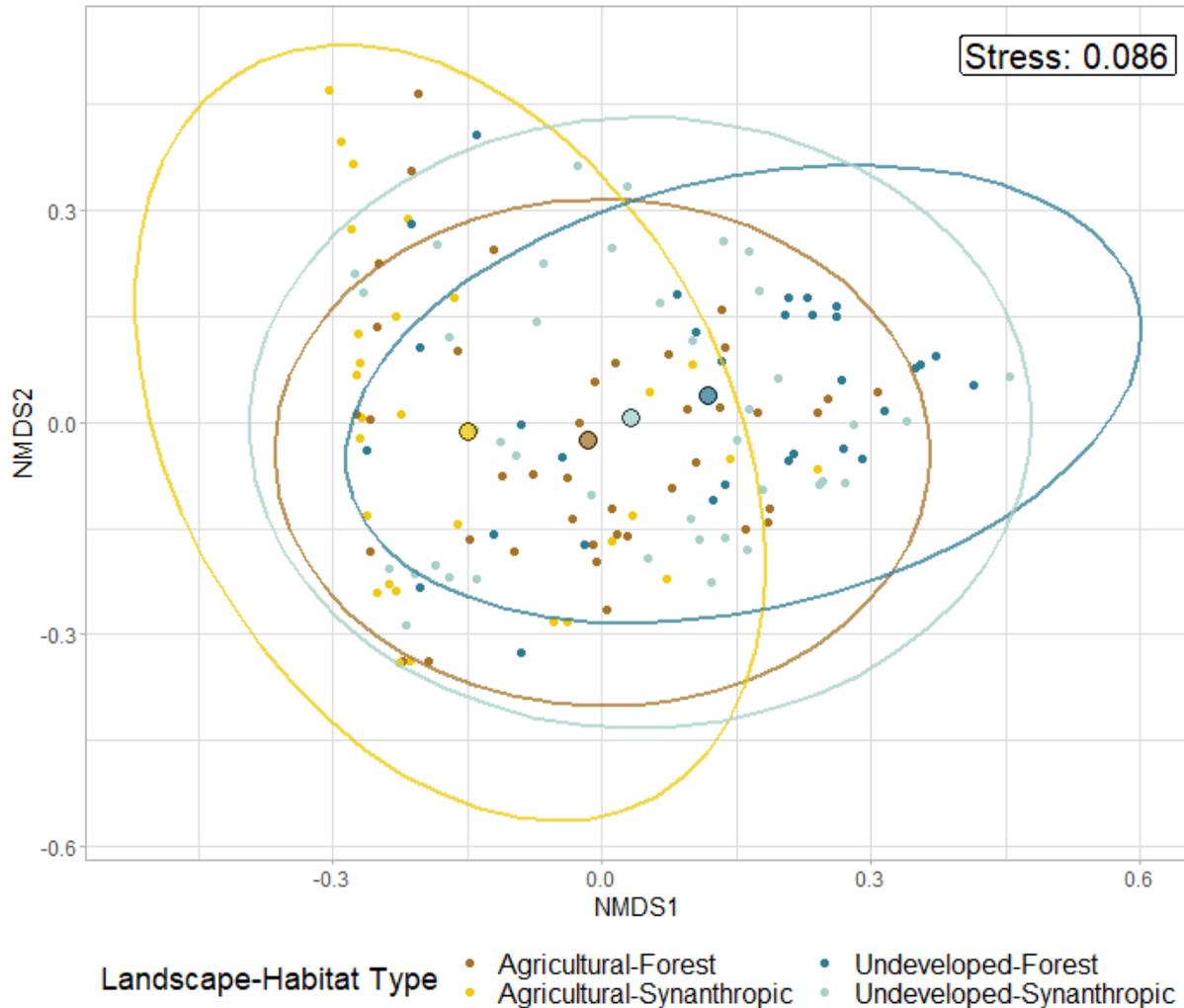
723

724 Figure 2. Alpha diversity for all unique mouse fecal samples (n=140) in anthropogenic and
 725 undeveloped landscapes and in forest and synanthropic habitat according to A) observed
 726 species richness B) Shannon diversity index C) Simpson diversity index and D) species
 727 evenness.
 728



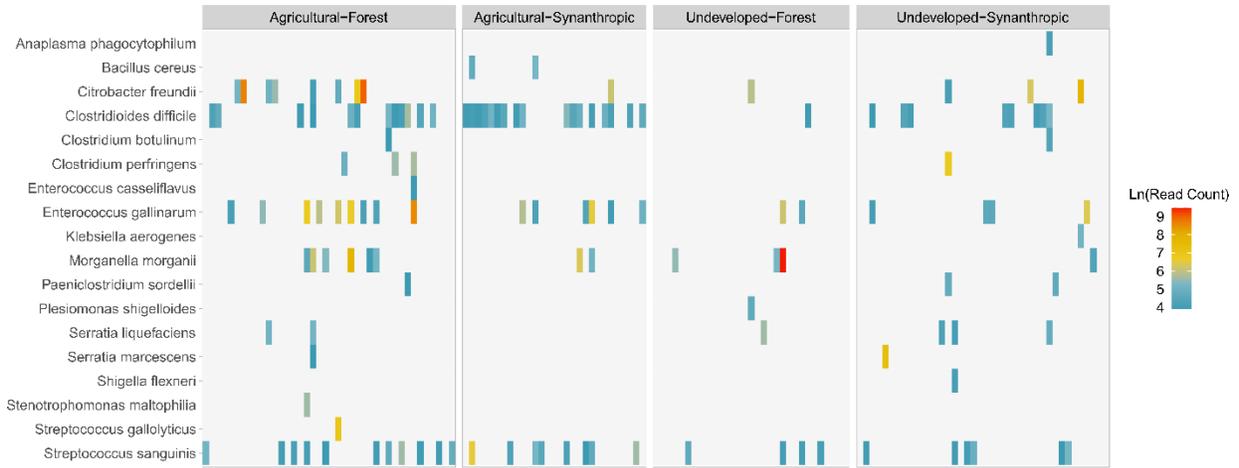
729

730 Figure 3. Relative abundance of bacteria phyla per sample (n=140) by landscape-habitat pairing
 731 showing phyla present at $\geq 1\%$ relative abundance. Phyla observed at $< 1\%$ relative abundance
 732 were grouped in a single category "Other". The microbiome of sampled mice was dominated by
 733 three phyla: Bacteroidetes, Firmicutes, and Proteobacteria.
 734



735

736 Figure 4. Non-metric multidimensional scaling ordination on microbiome community composition
 737 by Bray-Curtis dissimilarity index. Points represent individual samples, colored by landscape-
 738 habitat pairing. Ellipses denote the 95% confidence level for a multivariate t-distribution of the
 739 data points per group (centroids marked with larger points). Stress value: 0.086 (k=4).
 740



741

742 Figure 5. Heatmap of read counts of putative pathogenic bacteria species per mouse in each
 743 landscape-habitat pairing (count threshold >50 reads). The vertical axis represents samples
 744 from an individual mouse. Warmer colors indicate higher read abundance (natural log scale).